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TRANSBOUNDARY RIVER SOCKEYE
SALMON ENHANCEMENT ACITIVITES
FINAL REPORT FOR SUMMER 1992
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## EXECUTIVE SUMMARY

Joint Canada/U.S. enhancement of transboundary river sockeye stocks began in 1989, when eggs were taken at Tahltan Lake on the Stikine River, incubated at Snettisham Central Incubation Facility (CIF) located near Juneau, Alaska, and the resulting fiy back-planted to Tahitan Lake. In 1990, eggs were again taken at Tahltan Lake and enhancement of Taku River sockeye salmon stocks began, with egg-takes at Little Trapper and Little Tatsamenie lakes to produce fry for outplanting to Trapper and Tatsamenie lakes, respectively. Annual egg-takes were conducted at all sites from 1991 through 1995, with the exception of Little Trapper, where they were suspended in 1995. Under terms of the Pacific Salmon Treaty, ${ }^{1}$ brood years (BY's) 1991, 1992, 1993 and 1994 Tahitan Lake origin fry were divided between Tahitan and Tuya lakes. Activities up to the spring of 1992 have been previously reported (PSC 1994); the present report begins with the egg-takes in the fall of 1992 and continues through to smolt migrations and fry outplants in the spring of 1995. Results of the 1995 egg takes are included as well, for infonnative purposes. This report does not deal with adult returns, which began with the return of 4-yr-old Tahltan fish in 1993.

Methods are described, including egg-take and hatchery operations, otolith marking, fry outplanting, monitoring of outplants including hydroacoustic/limnological surveys and smolt sampling, and ancillary activities. Hatchery and otolith mark related activities are also described. Results are presented for each lake for the period from egg-take through smolt migration, followed by limnological observations and ancillary activities. No attempt was made to estimate benefit/cost ratios in this report. It is recommended these analyses should be done in the near future, using recently acquired adult return data and actual costs of the projects. The report concludes with a summary of major results and recommendations. Inportant results are summarized below.

## Hatchery Operations

Major modifications to convert the existing hatchery building at Port Snettisham into central incubation facility (CIF) for sockeyewere completed in August, 1993. In addition to more space and better stock isolation, the new facility has improved capability for water treatment. All eggs from brood year (BY) 93 and subsequent years will be incubated in this new facility. The new CIF is fully modularized, with four of ten modules committed to transboundary sockeye salmon incubation, allowing for much improved isolation of separate stocks. The physical plant is also much improved, with thermal marking, water quality, and egg and fry handling methods modernized.

## Otolith Marking and Reading

A laboratory has been established in Juneau to examine thermal marking techniques and develop methods for mass processing of otoliths from returning adults. Marks have been recovered from Alaskan domestic sockeye stocks and transboundary sockeye juveniles and smolts arising from the outplants. In addition, marks from the first returns of sockeye salmon adults resulting from enhancement activities at Tahltan Lake were recovered in 1993 (BY 89). Initial problems regarding the clarity of some of the marks during the initial years of the program and the ability to identify these fish in mixed stock fisheries have been identified and are

[^0]being worked on. At this time it is clear that in-season recovery and analysis of thermally marked fish is possible and that results of these analyses can be made available to fishery managers in a timely manner.

## Tahltan Lake Outplant Project

There were no problems meeting egg-take goals and outplanted fry have grown and survived well. The maximum carrying capacity of Tahltan Lake has not been defined. The lake is capable of supporting current levels of outplanting, but abnomally high wild fry production could result in fry densities which may not be sustainable on a continued basis. Caution is therefore advised, and outplant numbers should be reviewed annually through analysis of data sets from ongoing limnological and fry and smolt monitoring programs.

## Tuya Lake Outplant Project

Tahltan Lake provides a ready source of broodstock for fry outplants to Tuya Lake. As expected, growth of outplanted fry was exceptional. Survival, although not precisely determined, appears to be good. Final confirmation of this depends on adult returns. The lake appears capable of supporting outplants in excess of those to date. Current outplant levels are considerably below those allowed by the euphotic volume model. However, it may be prudent to proceed cautiously until changes in the observed zooplankton community structure induced by current fry outplant levels, stabilize .

## Tatsamenie Lake Outplant Project

Escapement levels in several years have restricted availability of broodstock. Because of genetic concerns, the egg-take site was moved from Little Tatsamenie to Tatsamenie Lake in 1994. It is recommended that fishery management strategies be refined to allow greater escapements in the future. Outplanted fry grew well; survival has been difficult to determine but there are indications it is less than expected; confirmation by adult returns is required. Tatsamenie Lake appears capable of supporting fry outplants considerably in excess of those to date at current escapement levels; however, increased wild production in combination with outplants could conceivably tax nursery lake carrying capacity, The number of natural spawners may have to be considered in future when determining appropriate numbers for outplant.

## Trapper Lake Outplant Project

Although there were no problems obtaining sufficient eggs, this stock appears to be more susceptible to Infectious Haematopoietic Necrosis Virus (IHNV) than others. Outplanted fiy have grown well but it has not been possible to deternine survival with any degree of accuracy. Egg-takes were suspended in 1995 because of survival uncertainty as well as concerns about the early outmigration of outplanted fry. The early outmigration of enhanced fry could have a detrimental affect on wild stocks of sockeye fry rearing in Little Trapper Lake. It is recommended this suspension remain in effect until adult returns from initial outplants are assessed. It is also recommended the lost production be replaced by increasing outplants to Tatsamenie, and/or by beginning outplants to other Taku drainage lakes, such as Nakina Lake.

### 1.0 INTRODUCTION

Joint Canada/United States enhancement of transboundary river salmon stocks began in 1989, when 3.3 million sockeye salmon eggs were taken at Tahitan Lake in the Stikine River watershed (Figure 1). These eggs were incubated at the Snettisham temporary Central Incubation Facility (CIF) near Juneau, Alaska, and the resulting fry back planted to Tahltan Lake in the spring of 1990 . Details of this project, including observations on the planted fry during their first summer and ancillary enhancement activities conducted during 1989 and early 1990, have been previously reported (PSC 1991). In 1990, eggs were again taken at Tahitan Lake and enhancement of Taku River sockeye salmon stocks began, with egg takes at Little Trapper and Little Tatsamenie lakes to produce fry for outplanting to Trapper and Tatsamenie lakes, respectively (Figure 2). Further egg-takes were conducted at all three sites in 1991, 1992, 1993, and 1994. Details and results of these enhancement activities from the fall of 1990 through to the spring of 1992 have been previously reported (PSC 1994). This report presents details and results to the smolt emigration stage, for all transboundary enhancement projects, from the summer of 1992 through the spring of 1995. Adult returns, including a small enhanced component, began in 1993 with returns of 4 -yr-old Tahitan fish (4-yr-olds are a minor component of sockeye returns in this area), and will be the subject of later reports.

### 2.0 METHODS

### 2.1 Egg-takes

Egg-take methods for 1992 and 1993 were the same as those previously reported (PSC 1994). In 1994 minor procedural changes were implemented at all egg take sites. One of the changes involved secondary rinsing of eggs with a 100 ppm iodophor solution, as per the revised Alaska Department of Fish and Game (ADF\&G) protocol to reduce IHNV transmission. A second modification involved transporting the fertilized eggs in plastic bags instead of muslin lined aluminum baskets.

### 2.2 Hatchery Operations

All eggs from BY 92 were incubated at the Snettisham temporary CIF. All eggs from BY 93 onwards were incubated at the Snettisham permanent CIF. There are no plans to change the incubation site. Hatchery methods documented previously (PSC 1994), have remained constant with the exception of the method of fry ponding. Instead of physically removing the fry when they reach a 3-5\% yolk sac ratio, the fry from BY 92 through BY 94 have been allowed to volitionally emerge into temporary holding containers. When the majority of fry have exited an incubator, the remainder are removed and transferred to the holding container . Although this method is less stressful to the fish, emergence is extended over a longer period of time. This delay makes timing of flights to complete outplanting more problematic and may also result in the excessive absorption of yolk sac reserves. As a result, hatchery staff are reviewing the volitional emergence technique and plan to use it for a shorter period before physically removing the remaining fry in future years.

### 2.3 Otolith Marking and Reading

### 2.3.1 Alaska

In 1989, the Transboundary Technical Committee agreed to mass mark all sockeye from transboundary river enhancement projects by manipulation of hatchery water temperatures during incubation to induce patterns of ring deposition on otoliths. A central laboratory for processing of otoliths has been developed by ADF\&G in Juneau, Alaska. Responsibilities of this laboratory include refinement of techniques for marking and development of methods for mass processing of otoliths taken from returning adults. The laboratory supports the sampling of otoliths in Alaskan commercial fisheries and is capable of supplying estimates of the contribution of enhanced fish to catch and escapement either during or after the fishing season.

### 2.3.2 Canada

Canada began developing the expertise to examine otoliths for thernal marks in 1994, with the hiring of a technician working under the direction of the Aging Lab at Nanaimo, B.C. Prior to this, marks had been read at either Eric Volk's Washington Department of Fisheries Lab or, later, at the ADF\&G Lab in Juneau. In 1995, all Canadian transboundary otoliths from juveniles, adults, and smolts were processed at the ADF\&G Lab in Juneau, Alaska.

### 2.4 Outplanting

Outplanting procedures are consistent with those previously reported (PSC 1994). However, the volitional emergence technique initiated by Snettisham Hatchery with fry from BY 92, has shifted the timing of fry outplants to later in the season (Appendix 3).

### 2.5 Monitoring of Outplants

### 2.5.1 Hydroacoustic/Limnological Surveys

1992: Hydroacoustic and limnological surveys were conducted to evaluate the freshwater growth and survival of the fry outplanted to Tahltan, Tatsamenie, and Trapper lakes. Tuya Lake was also surveyed, this being the first year of smolt migration from this lake. This work was contracted to Triton Environmental Consultants, Ltd., with subcontracting to B. Mercer and Associates. Four surveys were conducted: June 1825 (limnological only), July 24- August 5 (hydroacoustic/limnological), August 20-25 (limnological only), and September 17 to October 4 (hydroacoustic/limnological). All surveys included beach seining in the littoral (near shore) zone to determine the relative proportions of wild and outplanted fry. Trawl sampling in the pelagic (offshore) zone was conducted in conjunction with the hydroacoustic surveys to obtain population estimates of wild and enhanced fish. The limnological surveys included measurements of water clarity, water temperature, dissolved oxygen, total dissolved solids, total phosphorous, total nitrate and chlorophyll $a$, as well as collection of zooplankton and phytoplankton samples.

1993: The first limnological survey, June 16-20, was conducted by B. Mercer and Associates, and included collection of plankton samples, Sechii depths, and temperature profiles. Lakes surveyed included Tahltan, Tuya, Tatsamenie, and Trapper as well as Little Trapper. It was felt closer monitoring of Little Trapper

Lake was required because of the observed early outmigration of Trapper Lake fry outplants. Nakina Lake, on the Taku system, was examined the first time for its suitability for future fry outplants. The two remaining surveys were contracted to Triton Environmental Consultants, Ltd. with subcontracting to B. Mercer and Associates. The second survey, July 28 to August 4, included all lakes surveyed in June. The Nakina Lake survey included only zooplankton sampling. For all other lakes, limnological measurements of water clarity, water temperature, dissolved oxygen, total dissolved solids, total phosphorus, total nitrate, and chlorophyll a were taken. As well, zooplankton and phytoplankton samples were collected. Nitrates and total dissolved solids were not collected at Tuya, Tatsamenie, or Trapper lakes. Beach seining was conducted at all lakes except Nakina. Hydroacoustics and trawling were omitted from this survey since surveys in previous years at this time of the season had revealed large numbers of fry remaining onshore where they are inaccessible to hydroacoustic enumeration. From September 5 through September 21 a final survey of all lakes was conducted following the same sampling regime as the previous survey, although with the addition of hydroacoustics and trawl sampling.

1994: Surveys in 1994 were very similar to those of 1993. Lakes surveyed included Tahltan, Trapper, Tatsamenie, Tuya, Little Trapper, and Nakina. There was also a single survey of Little Tatsamenie Lake; this lake was included since there was the possibility for early outmigration of Tatsamenie fry outplants from Tatsamenie into Little Tatsamenie, as has been observed at Trapper/Little Trapper lakes. The first survey, June 16 to 18, was conducted by B. Mercer and Associates and included collection of plankton samples, Sechii depths, and temperature profiles. Lakes surveyed included Tahltan, Tuya, Tatsamenie, Little Tatsamenie, Little Trapper, and Nakina. Attempts to survey Trapper Lake were unsuccessful due to adverse weather. The second survey, conducted by B. Mercer and Associates (subcontract from Triton Environmental Consultants Ltd.), from July 23 to 29, included Tahltan, Tatsamenie, Tuya, Trapper, and Little Trapper lakes. Nakina and Little Tatsamenie were excluded due to cost considerations. Sampling included beach seining, zooplankton hauls, water clarity and temperature, oxygen profiles, chlorophyll-a, and total phosphorus. Total dissolved solids, Nitrogen, and phytoplankton were not sampled, as has been done in previous years. A final survey was conducted September 2 to 22 on all lakes except Little Tatsamenie, which was excluded because of cost considerations. Measurements and sampling regimes were consistent with the July surveys with the addition of hydroacoustics combined with mid-water trawling to obtain population estimates

Problems associated with obtaining accurate population estimates were discussed extensively in a previous report (PSC 1994). These included fry remaining onshore where they are missed in hydroacoustic estimates and difficulties in obtaining adequate numbers of fish in trawl samples during midsummer surveys, as well as the questionable accuracy of the hydroacoustic surveys. The number of transects was increased in Tahltan, Tuya, and Trapper lakes in 1993 in an attempt to improve accuracy, however, many of the other problems still remain. The large, undetermined, numbers of fry remaining onshore in early summer was a major factor in the decision to discontinue the July hydroacoustic surveys and to do them only in the fall when fry have moved offshore in most lakes. Specific problems encountered in deriving estimates at each lake are described in the appropriate sections to follow.

### 2.5.2 Smolt sampling

Smolt sampling was conducted at the outlets of Tahltan, Tatsamenie, Trapper, and Little Trapper lakes in 1993, 1994, and 1995. In 1993 smolt sampling commenced at Tuya Lake, following the first fry outplanting in 1992. At Tahltan Lake smolt migrations were enumerated and sampled in a weir program conducted by Fisheries and Oceans Canada (DFO), Whitehorse, as described in a previous report (PSC 1994).

At Little Trapper, Trapper, and Tatsamenie lakes, samples were collected on several occasions over the period of smolt migration by employing a fyke net and procedures documented by Hyatt et al. (1984). This sampling was conducted by B. Mercer and Associates. Sampling at Tuya Lake was conducted by Triton Environmental/Tahitan Tribal Council. Access to Tuya Lake during the early portion of smolt out-migration is often restricted to helicopter and due to the high cost of this mode of travel, it was decided to camp at the lake outlet and trap nightly for several days rather than make intermittent trips. Samples from these lakes provide smolt size, age and thermal mark information. When possible, smolt production is estimated from the in-lake hydroacoustic surveys conducted the fall prior to smolt outmigration.

### 2.6 Ancillary Enhancement Activities

A number of sockeye enhancement related studies were conducted which included genetic analysis of Tatsamenie river and lake sub-stocks; Tatsamenie Lake broodstock capture, holding and spawning; assessment of Nakina Lake (Taku system) for enhancement potential for fry outplanting; collection of more detailed limnological data from the outplant lakes and Little Trapper Lake; and continued study of changes in diel migration patterns of the zooplankton in Tuya Lake to determine possible response to sockeye introductions. Short term holding studies on outplanted fry to assess mortality after transport and outplanting were conducted at Tatsamenie and Trapper lakes in 1993. A short term feeding experiment was conducted at Trapper Lake with a portion of the BY 1994 outplanted fry with the objective of determining if short term rearing would result in increased fry survival.

### 3.0 RESULTS

### 3.1 Hatchery Operations

The construction of a larger, permanent CIF at Snettisham (renovation of the existing main hatchery building) was completed in August, 1993. All transboundary sockeye eggs from brood years 1993 to the present have been incubated in this new facility.

The new CIF has ten modules, four of which are dedicated to transboundary river sockeye salmon incubation. The new CIF has improved capability for heating and chilling incubation water, isolating separate sockeye salmon stocks, treating incubation water to increase hardness (with $\mathrm{CaCl}_{2}$ ), and controlling gas supersaturation by stripping dissolved nitrogen and adding oxygen. The new CIF is also more spacious, facilitating egg receipt and fiy transfer procedures.

Preliminary results of in-hatchery survival in the new CIF were encouraging. For BY 1993 and 1994, only one incubator of fiy was lost to IHNV. No Little Trapper Lake fish were lost to IHNV; and the in-hatchery
survival for this stock was improved although still short of the of $80 \%$ in-hatchery survival biostandard (BY 1993 in-hatchery survival of 78.1\%). In-hatchery survivals for Tahltan Lake eggs, (those destined for both Tahltan and Tuya stocking), exceeded the in-hatchery biostandard (BY 1993 and 1994 in-hatchery survival's were $91 \%$ and $89 \%$, respectively). BY 1993 Tatsamenie Lake sockeye salmon eggshad a poor in-hatchery survival of $46 \%$, due to IHNV losses and poor green to eyed-egg survival, but BY 1994 had an improved inhatchery survival of $73 \%$. The reasons for the poor green to eyed egg survivals are not known, but since other stocks had good in-hatchery survivals it seems unlikely that the poor Tatsamenie Lake in-hatchery survivals were due to water quality or other hatchery factors.

### 3.2 Otolith Marking and Processing

### 3.2.1 Alaska

## Development of Marking Procedures

Successful otolith marking is directly related to the ability of a hatchery to maintain at least two levels of rearing temperatures, and to provide control for rapidly switching between temperatures to induce a unique banding pattern. Successful marking is also influenced by the proportion of fish in a group that can be simultaneously marked, and the extent that marking can be completed within a particular time frame. At the Snettisham hatchery, learning to induce recoverable marks has been a gradual process, with each year of marking fish providing a new example of what factors can confound the marking process.

A number of early attempts to identify Transboundary River stocks at Snettisham were limited to three and four ring marks (Appendix 1.) We have since learned that three ring patterns can appear in wild stocks. In addition, many of these early markings took place during hatching. Hatching can induce variable patterns in otoliths; it may mask the appearance of the marks or may induce additional rings which look similar to thermal marks. Marking protocols now include increasing the number of thernal rings, and completing the marking before hatching or starting the marking soon after hatching.

Difficulties also arise when trying to balance schedules to uniquely mark a number of different groups given limited availability of heated and chilled water. Separating different brood years by unique thermal marks has added to this problem. The ADF\&G otolith laboratory has since modified marking protocols to include an accessory band of marks that identifies brood year. In addition the lab has started to investigate counting annuli patterns in the otolith while processing for thermal marks and use of that information to identify brood year.

To increase understanding of the processes that control thermal marking, the laboratory entered into a cooperative project with the University of Alaska. The project involved experimenting with different marking protocols that would have been too risky to apply to the Snettisham releases. The results confirmed the laboratory's concerns about marking during hatching and provided indications that applying less than two Temperature Units between thermal rings can produce a ring spacing that is less than one micron. A one micron ring spacing requires more effort to resolve ring count than can be accomplished using rapid mass processing methods.

## Development of laboratory procedures

The laboratory has explored various options for mass processing otoliths. During this investigation it became apparent that manual processing of individually mounted otoliths by trained personnel was the fastest means of accurately detecting thermal marks from mixed stock fisheries. The primary reason why manual processing is the most expedient approach is that the morphology of otoliths are quite variable. In an examination of the shape of left and right otoliths from various individuals it was found that individual variation accounted for almost half the amount of variation found between the individual fish. Because recovery of the thermal marks requires grinding enough material away to get to an optimum viewing plane within an otolith, this high degree of shape variation precludes using a machine based processing approach in which a standard amount of material is removed for each otolith.

In addition to the high degree of shape variation, otoliths also contain a good deal of variation in the background of natural patterns in their microstructure. When preparing to process otoliths, laboratory personnel first examine otoliths from the voucher collection. These voucher otoliths are collected from a sub-sample of marked fry prior to release into lakes. The purpose of examining the voucher otoliths is to develop a visual search image of the thermal mark pattern. By processing otoliths individually, the laboratory personnel are able to make a judgment about the presence or absence of the particular mark as the otolith core is gradually exposed. Development is underway to utilize image processing as an aid to the recovery of marks, but for this application the human eye is still the most sophisticated tool for pattern recognition.

In addition to examining voucher otoliths, the laboratory also examines samples from smolts or fry that are captured in the lakes targeted for enhancement. Information from these samples is used to determine survivorship of the marked fish. These samples also provide an opportunity for personnel to develop skills in recognizing the thermal mark patterns of the juveniles in preparation for adult returns.

The summer of 1993 also provided the first opportunity to recover thermal marked adult sockeye returning to Alaska's district 108 fisheries. These marked fish were the returning 2-ocean age class and as a result did not constitute a significant portion of the catch.. Notwithstanding, 1000 sockeye were collected in an effort to gain experience in sampling the commercial fishing catches and dissecting fish and removing the otoliths in the ports. From these samples the laboratory recovered seven marked Tahltan otoliths from the 1989 brood year.

The 1994 fishing season marked the first test of the ADF\&G's otolith processing facility to meet the objectives identified as part of the US/Canada agreements in enhancing sockeye production. The lab was able to provide fisheries managers with an in-season estimate of the proportion of enhanced sockeye caught in 52 commercial openings over a 10 week period. These initial estimates were made by processing 4,653 otoliths taken from seven different districts and sub-districts, with the information given to managers in time for their next weekly opening. The program continued in a similar manner in 1995.

Following the in-season estimates, post-season processing continues during the summer and fall. The postseason processing is intended to confirm the initial estimates, allow for replicate readings as patt of a quality control program, and provide overall estimates of contribution of the enhanced fish to the commercial fisheries. To help prioritize the samples to process during the post-season period, the U.S. implemented and
evaluated an approach developed by ADF\&G's statewide Salmon Biometrician. This approach uses an adaptive sampling scheme which involves periodically updating the estimates on stock contribution as new fisheries come in and more otoliths are processed. The method determines which mix of the remaining otoliths will minimize the overall uncertainty on the numbers of enhanced fish. The approach was evaluated and utilized on the District 108 and 106 fisheries. Because it involves Baysian sampling theory, the uncertainty estimates from those fisheries are in the form of credible intervals. For all practical purposes a credible interval can be treated essentially the same as a confidence interval. Further work on this approach will be conducted, and the extent to which it will be more fully utilized in the lab depends, in part, on the development of a database system.

Development of an otolith processing quality control program continued. The program involved blind replicate readings as well as occasional planting of marked otoliths during training periods. Overall, there was consistency in detecting the Tahltan and Tuya marked fish. More work is needed to develop the analytical approach to evaluate the data quantitatively.

The Tatsamenie 1990 mark was very poor and an expansion factor by which estimates of the number of enhanced fish based on the readability rates has yet to be developed. It is uncertain if this is possible. Based on observations from voucher samples, it was estimated that about a quarter of the enhanced Tatsamenie stock may be completely unreadable and a significant number of the others may be difficult to interpret. Similarly, the Trapper 1990 stocks are poorly marked with over $40 \%$ considered difficult to distinguish with certainty. Therefore, contribution estimates for the BY 1990 Taku Transboundary enhanced stocks must be considered preliminary at this time.

### 3.2.2 Canada

Fisheries and Oceans Canada began developing capabilities to examine for otolith marks in 1994, with the hiring of a Technician to work under the direction of the Aging Laboratory staff at the Pacific Biological Station Nanaimo, B.C. This lab has extensive experience aging both salmonid and non-salmonid fish by otoliths but had not previously worked with thermally marked otoliths. The Technician visited the otolith labs in both Juneau and Olympia (Washington Department of Fisheries) to observe processing techniques and to develop standardized terminology.

In 1994 a portion of the transboundary juvenile and smolt otoliths were read in the Nanaimo lab; aurd all adult otoliths and the remainder of the juvenile and smolt otoliths were read in Juneau. In 1995 all Canadian transboundary juvenile, smolt, and adult sockeye otoliths as well as chinook otoliths from three southem British Columbia hatcheries were read at the Nanaimo lab. A manual explaining thermal marking and processing techniques has now been written (Hoyseth 1995). Although the lab is not presently funded as a support service, this is the eventual goal.

### 3.3 Disease Testing and Outplant Dates, all Lakes

Levels of IHNV in the broodstock varied dramatically between brood-years for all lakes, for reasons not clearly understood. Levels of BKD varied much less, being relatively low in all cases. Broodstock disease testing results for brood years 1988 through 1994 for all lakes are presented in Table 26, and discussed in detail in Appendix 2. Stocking dates for all brood years for each lake are summarized in Appendix 3.

Outbreaks of IHNV occurred in alevins from Little Trapper Lake BY 1990, 1991 and 1992, but the prevalence of the virus in the sampled brood stock varied from $96.1 \%$ for BY $1990,13.3 \%$ for BY 1991, and $97.3 \%$ for BY 1992. The prevalence of the virus in BY 1993 adult spawners, which had no losses to IHNV, was $60.0 \%$. No outbreaks occurred in Little Tatsamenie Lake alevins of BY 1990 and 1991; the prevalence of the virus in these years varied from $64.0 \%$ (BY 1990) to $3.3 \%$ (BY 1991). There were losses to IHNV for Tatsamenie sockeye from BY 1992 and 1993, the prevalence of the virus in these groups of fish was $63.3 \%$ and $63.1 \%$ respectively, very close to BY 1990. BY 1994 had no losses to IHNV after virus prevalence dropped to $1.0 \%$. For BY 1992 through 1994 Tahitan eggs, no IHNV losses were experienced for BY 1993 and 1994 (virus prevalence was $71.8 \%$ and $50.0 \%$, respectively) but approximately 500,000 alevins were lost from BY 92 (virus prevalence was $91.6 \%$ ).

### 3.4 Egg-takes, Incubation, Outplanting, Growth and Survival, by Lake

Results of egg takes, incubation, outplanting, and egg to outplant fry survival for all enhancement sites are presented in Tables 1 through 3. Details of juvenile samples obtained by beach seining (onshore) and trawling (offshore) in evaluation surveys during the summer and fall of 1992 through 1995 are presented for all lakes in Tables 4 through 13 and Figures 3 through 11. It should be noted that supplemental beach seining was performed on those surveys where a sufficient number of juveniles were not captured at the index sites. Therefore in some instances the total number of juveniles sampled may exceed the actual total index catch as given in Tables 4 through 13.

Population estimates based on hydroacoustic and trawl surveys for the summer and/or fall of 1992 through 1994 for all lakes are presented in Table 14. Total estimates are apportioned to enhanced or wild and to age class based on the trawl catches detailed in Tables 4 through 13. Survival values for enhanced fish, based on these numeric estimates, are presented in Table 15. Smolt observations are summarized in Tables 16 through 20 and Figure 12. Results for each individual lake are discussed below.

### 3.4.1 Tahltan Lake

## Egg-take Through Outplant Activities (Tahltan and Tuya) ${ }^{2}$

Results for egg take to outplant activities for BY 1992 through 1995 are summarized in Table 1a (Tahltan) and 1b (Tuya). Project details for BY 1989 to 1991 have been previously reported (PSC 1991; PSC 1994), and details for BY's 1992-1995 are given below. Annual reports detailing the activities and results of the Tahltan egg takes (Triton Environmental Consultants Ltd. 1989-1995), have been prepared by the contractor and submitted to the DFO contracting authority. As previously noted, under terms of the existing Pacific Salmon Treaty, a portion of the fiy resulting from Tahltan Lake egg-takes are to be planted to Tuya Lake provided the BY escapement to Tahltan Lake exceeds 15,000 ; this occurred for the first time in 1991 and has continued each year during the 1992-1995 period.

[^1]The total number of fish utilized for brood stock was 3,694 , ( 1,847 females and 1,847 males), captured from an escapement of 59,907 fish. Eggs were shipped in ten lots during the period September 2 to September 20. All eggs were shipped the day they were taken. The average fertilization rate was $93 \%$. The goal of 5.4 million was not met based on the hatchery adjusted green egg estimate of 4.9 million eggs.

Since the Tahltan escapement exceeded 15,000 in 1992, the 1992 BY Tahltan eggs were divided into two groups and marked distinctively for planting to Tahltan and Tuya lakes. Survival at the hatchery was $90 \%$ from green egg to outplanted fry for the Tahltan Lake group, which was well above the established biostandard.

The otoliths of all 1992 BY Tahltan fry planted in Tahltan Lake were marked with a 7 ring pattern ( 7 cycles of 48 h warm/ 48 h chilled water) at the pre-hatch stage. A total of $1,947,000$ fry were planted on June 23 and 26 and July 2, 1993.

The Tuya lake green egg to survival was $72 \%$; this low survival rate was due to the loss of approximately 520,000 fish to IHNV prior to stocking. The otoliths of all BY 1992 fry planted in Tuya Lake were marked with a 5 ring pattern ( 5 cycles of 48 h warm / 48 h chilled) at the pre-hatch stage. A total of $1,990,000$ fry were planted on June 16 and 25 and July 7, 1993.

## 1993 Brood year

The total number of fish utilized for brood stock was 4,506, ( 2,253 females and 2,253 males), captured from an escapement of 53,362 fish. Eggs were shipped in ten lots during the period September 2 to September 21. The average fertilization rate was $95 \%$. Brood stock was readily captured due to the high escapement and the egg-take target of 6.0 million was marginally exceeded with the adjusted green egg estimate being 6.1 million.

Since the Tahltan escapement exceeded 15,000 in 1993, the BY 1993 Tahltan eggs were divided into two groups and each marked distinctively for planting to Tahltan and Tuya lakes. Survival at the hatchery from green egg to outplanted fry was $93 \%$ which was the highest in-hatchery survival to date.

The otoliths of all 1993 BY Tahltan stock planted in Tahltan Lake were marked with a $6+5$ ring pattern, 6 rings prior to hatching with 5 rings post hatching. A total of 904,000 fry were planted on June 24 and 28, 1994.

The Tuya Lake destined fry had a green egg to fry survival of $91 \%$. The otoliths of all BY 1993 fry planted in Tuya Lake were marked with a $4+5$ ring pattern, a 4 ring pattern during pre-hatch stage and a 5 ring pattern during post-hatch. A total of $4,691,000$ fry were planted on June 24, 28 and 30 and July 1, 12, and 13, 1994.

## 1994 Brood year

The total number of fish utilized for brood stock was $3,378,1,689$ females and 1,689 males, captured from
an escapement of 46,363 fish. Eggs were shipped in 9 lots during the period September 1 to September 22. Although brood stock was readily captured because of the high escapement, the egg-take target of 6.0 million was not met, resulting in an adjusted green egg estimate of 4.2 million (The 1994 egg take at Tahltan was stopped after the crash of the egg transport plane on September 22).

Since the Tahltan escapement exceeded 15,000 in 1994, the 1994 BY Tahltan eggs were again divided into two groups and marked distinctively for planting to Tahltan and Tuya lakes. Survival at the hatchery was $94 \%$ from fertilized egg to outplanted fry for those fry planted in Tahltan.

The otoliths of all 1994 BY Tahltan stock planted in Tahltan Lake were marked with a 6 ring pattern. A total of $1,143,000$ fry were planted on June 26 and July 3, 1995.

The Tuya Lake destined fry had an in-hatchery survival of $82 \%$. The otoliths of all BY 1994 fry planted in Tuya Lake were marked with a 4 ring pattern. A total of $2,267,000$ fry were planted on June 21, and 25 and July 3, 1995.

## 1995 Brood year

The total number of fish utilized for brood stock was 4,850 , ( 2,425 females and 2,425 males), captured from an escapement of 42,317 fish. Eggs were shipped in 13 lots during the period August 31 to September 25 . The fertilization rate was $95 \%$. As brood stock was readily captured because of the high escapement, the egg-take target of 6.0 million was exceeded and the adjusted green egg estimate was 6.9 million. A higher than average fecundity resulted in an egg take larger than the original estimate of 6.1 million eggs.

Since the Tahltan escapement exceeded 15,000 in 1995, the 1995 BY Tahltan eggs were divided into two groups and marked distinctively for planting to Tahltan and Tuya lakes. The otoliths of all BY 1995 fry destined for Tahltan Lake were marked with a 6 ring pattern, while fry destined for Tuya Lake were marked with a 4 ring band. Fry will be planted in June and July of 1996.

## Growth and Survival

## Juvenile observations

From 1992 through 1994, beach seine index catches from Tahltan Lake indicated relatively low numbers of fry remained onshore by the end of July and almost none by late August and early September (Tables 4, 5). Trawl catches in both July and October of 1992 were poor; however, catches at Tahitan are generally small and appear to be attributable more to high trawl net avoidance in the extremely clear water, rather than low abundance. No hydroacoustic population estimates were obtained in 1992 due to equipment problems. Only age $0+$ juveniles (BY 91) were captured, and the percentage of enhanced fish (Table 4, Figure 3) ranged from 0 to $18 \%$ (Table 4, Figure 3). Part of this variability is likely attributable to small sample sizes. Average lengths of both enhanced and wild juveniles were very similar.

In 1993 (Table 5, Fig. 4), the percentage of enhanced fish was very low in the early August beachseines, but increased to $36 \%$ for age $1+$ and $13 \%$ for age $2+$ in the September trawls. This is likely a good estimate for the total population, since very few fish remained onshore and the sample size was large. A hydroacoustic
estimate of 817,429 juvenile sockeye was made on September 18 (Table 14). Based on the trawl sample composition, this was made up of approximately 294,274 age $0+$ enhanced, 417,706 age $0+$ wild, and 105,448 age $1+$ wild (Table 14), indicating a survival of $15 \%$ for enhanced fish from time of outplanting (Table 15). As discussed at the end of this section, hydroacoustic estimates at Tahltan usually vastly underestimate the true population size and survivals based on smolt estimates, presented below, are generally considered to be more accurate. As in 1992, the size of enhanced and wild fish of the same age was very similar.

In 1994 (Table 5, Fig 4) there were no fish captured in July. A small trawl sample in the September survey indicated $5.9 \%$ enhanced among age $0+$ juveniles. Very few fish, all wild, were captured in beachseining. A hydroacoustic population estimate of 436,634 was estimated to contain 25,761 age $0+$ enhanced juveniles (Table 14), indicating a survival of only $2.8 \%$ from time of outplanting (Table 15). Enhanced and wild fish were again of similar size.

## Smolt Observations

Detailed observations of Tahltan smolt migrations for 1991 through 1995, as well as migration estimates are presented in Table 16. The total numbers emigrating each year, separated into enhanced or wild (all ages combined), are plotted in Figure 15. The percentage of enhanced smolts has ranged from $6.7 \%$ (1995) to $51.7 \%$ (1992).

Survivals of planted fry to fall juvenile and to age 1+ and age $2+$ smolts for each brood year are presented in Table 21. Survivals of outplanted fry have been quite high, averaging $18.7 \%$ to age $1+$ smolt. The notable exception is the 1993 brood year (1994 outplant), where survival was only $4.9 \%$. The average smolt survival excluding the 1993 BY is $22.1 \%$. The low percentage of enhanced smolts in the 1993 migration ( $12.7 \%$ ) is the result of exceptionally high survival from wild fish of the 1991 brood year, rather than poor survival of outplanted fry ( $26.2 \%$ to age $1+$ smolt). Survival of juveniles from the time of outplanting to the fall are based on hydroacoustic estimates. Comparison of the values for age $0+$ juveniles with those for the associated age $1+$ smolts the following spring show they generally underestimate the smolt values, even without allowing for any mortality between the time of the two observations (presumed to be quite low). The exception is the brood year 1992, when the fall and spring estimates were identical. The abnormally low survival to age $1+$ smolt for the 1993 brood year is supported by the exceptionally low associated fall fry survival estimate based on the hydroacoustics survey.

Tahltan Lake smolt size in relation to juvenile abundance in the spring of the first year of lake rearing is given in Table 22 and plotted in Figure 13. Smolt observations include two years when the lake was artificially fertilized (1986 and 1987). Estimates of spring juvenile abundance were calculated using known numbers of fry outplanted and/or smolt estimates back-calculated using the Tahltan Lake mortality curve presented in a previous report (PSC 1994). There is a noticeable relationship (Figure 13) between juvenile abundance and smolt size for the 1990 brood year only, when the 1991 spring juvenile population was exceptionally high (estimated at 12.4 million). However, even then the age $1+$ smolt sizes of 3.9 g (wild) and 3.8 g (enhanced), are not dramatically different from other years and actually exceed a 1983 value of 3.8 g . (Table 22).

The average age $1+$ smolt size for each lake based on all years of observations (from Tables 16 through 20) are plotted in Figure 12. As in all lakes where wild sockeye are present, the size of wild smolts slightly
exceeds that of enhanced smolts. It is likely this is a function of earlier emergence and commencement of feeding for wild fry, rather than differences in growth rate. As can be seen, Tahltan smolts are large in comparison to most other lakes, being second only to Tuya Lake (Figure 12).

## Hydroacoustic Estimates

Accuracy of the hydroacoustic estimates at Tahltan Lake was discussed extensively in a previous report (PSC 1994). Based on associated spring smolt counts and estimated fall to spring mortality, it was calculated that fall juvenile populations were underestimated by a factor of 2 to 3 times. However, it was found that there was a significant correlation between the fall hydroacoustic estimates and total associated smolts; (i.e., for a fall hydroacoustic estimate in year $t$, the total of all age $1+$ and age $2+$ smolts in year $t+1$, all age $2+$ smolts in year $t+2$, and all age 3 smolts in year $t+3$ ). This relationship was re-examined, with inclusion of two additional years of observations from 1993 and 1994 (Fig. 14). Excluding the 'outlier' 1990 hydroacoustic estimate, there is a highly significant relationship ( $\mathrm{p}=.02$ ); slightly better than that previously calculated ( $\mathrm{p}=.03$ ). Thus, the hydroacoustic estimates at Tahltan lake appear to have some predictive value for estimating the magnitude of annual smolt migrations.

### 3.4.2 Tuya Lake

## Egg-take Through Outplant Activities

No egg takes occur at Tuya Lake, broodstock for these outplants is obtained at Tahltan Lake. Fry outplants to Tuya Lake are detailed in the previous section on Tahltan Lake and presented in Table 1 b .

## Growth and Survival

## Juvenile Observations

The first outplant was made in 1992. As previously mentioned, there are no wild sockeye in Tuya Lake. The June survey of 1992 was the only occasion when substantial numbers of fry were captured onshore, in beachseine catches (Table 6). In later 1992 surveys and in 1993 and 1994, when surveys did not begin until late July (Table 7), virtually all fry had moved offshore. Trawl catches in 1992 and 1993 were very low. This was likely due to low fish density, since in 1994 when the number outplanted ( 4.7 million) was more than doubled, the catch was substantial. Fry grew extremely rapidly (Table 6 and 7; Figures 5 and 6), and by the time of fall surveys Tuya fry lengths generally exceeded those of Tahltan Lake fry, even though Tahltan surveys were conducted about two weeks later.

Hydroacoustic estimates (Table15), indicate survivals from time of outplant to the time of fall surveys ranged from $22 \%$ to $41 \%$, (average $33 \%$ ). It is not possible to assess the accuracy of this estimate, since there was no smolt enumeration program conducted at Tuya Lake. It should be noted that the estimated survival for this period at Tahltan Lake using the smolt estimate derived mortality model (PSC 1994) is 63\%, while the survival, derived from hydroacoustic estimates, averages only $10.4 \%$ (range $2.8 \%-15.1 \%$, Table 15). There are reasons why hydroacoustic estimates might be expected to be more accurate at Tuya Lake, notably higher water turbidity and lower fish densities (PSC 1994). However, given the magnitude of the fry underestimation at Tahltan Lake, it seems probable the Tuya Lake estimates are low as well, and actual
survival values may be higher than the acoustics based survival index, and perhaps closer to that predicted by the Tahltan smolt estimate mortality model.

## Smolt Observations

Details of the smolt emigrations observed at Tuya Lake in 1993, 94, and 95 are given in Table 17. Tuya smolts (originating from Tahltan stock) are exceptionally large due to the rich zooplankton forage base, the structure of which indicated low vertebrate predation levels prior to the first outplant in 1992. Age 1 smolts average about 9 g and 98 mm fork length, approximately twice or more the size of enhanced smolts from any other lake (Figure 12). While smolt emigrations at Tuya lake have not been enumerated, survival to the fall fry stage, as estimated from hydroaccoustic surveys, is relatively high (average $33 \%$, range $22.0 \%$ $33.3 \%$, Table 15). Given that fiy survival derived from hydroaccoustic surveys are probably underestimated, it is possible that survival to the smolt stage at Tuya Lake could be similar to the $63 \%$ value determined by the Tahltan Lake smolt count mortality model discussed above.

### 3.4.3 Tatsamenie Lake

## Egg-take Through Outplant Activities

The first egg-take was conducted in 1990. Results of the egg-take and outplant activities from BY 1990 and 1991 are summarized in earlier documents (PSC 1994), activities conducted from BY 1992 through 1995 are summarized in Table 2 and discussed in detail below. Detailed annual summary reports on the Tatsamenie egg takes have been prepared by the contractor and submitted to the DFO contracting authority (B. Mercer \& Associates Ltd. 1990-1995).

## 1992 Brood Year

Brood stock for outplants to Tatsamenie Lake was captured at the Little Tatsamenie Lake weir located 1 km downstream of the outlet of Little Tatsamenie Lake. Little Tatsamenie is situated approximately 5 km . downstream from Tatsanenie Lake. The connecting stream is passable to adult salmon and spawning occurs in both Tatsamenie Lake and the connecting stream. The proportion spawning in each location is uncertain but instream foot surveys by the Little Tatsamenie weir personnel and egg take crews suggest a connecting stream spawning population of 500-1500 through the period 1990-1994.

The total number of fish used for brood stock in 1992 was 791 ; with 435 females and 356 males spawned. From an escapement of 6,576 fish, 610 females and 475 females were captured and held. Female pre-spawn mortality was $9.0 \%$ ( 55 of 610) and male pre-spawn mortality was $6.5 \%$ ( 31 of 475 ). Eggs were shipped in six lots during the period September 12 to October 10. The overall fertilization rate was $86 \%$.

The egg-take target of 1.75 million was not achieved primarily due to a bear which tore a hole in one of the holding pens, resulting in the release of 56 females. The adjusted green egg estimate was 1.5 million eggs.

Survival at the hatchery was $71.2 \%$ from fertilized egg to outplanted fiy. The relatively poor survival was due in part to a loss of approximately 246,000 alevins to IHNV. This was the first time that Tatsamenie Lake stocks have had an in-hatchery mortality associated with IHNV.

The otoliths of all 1992 brood Little Tatsamenie stock were marked with a $4+3$ ring pattern, 4 rings at the pre-hatch stage and 3 at the post-hatch stage. A total of 909,000 fry were planted into Tatsamenie Lake on July 9 and 14, 1993.

## 1993 Brood Year

The total number of sockeye captured for brood stock was 1,041 from an escapement of 5,028 through the Little Tatsamenie weir. A total of 331 females and 312 males were spawned from the held broodstock. Prespawn holding mortality was 90 females and 65 males; $18 \%$ and $12 \%$ of held fish respectively. The eggs were shipped in six lots during the period September 8 to October 6. One lot of eggs was held overnight at the lake because of poor weather, the remainder were shipped on the day of spawning.

Included in the total number of fish captured listed above, 96 females and 75 males were captured and held near the outlet of Tatsamenie Lake. From these fish 45 females and 44 males were spawned as part of a pilot program undertaken to determine the feasibility of capturing, holding and spawning brood stock at the outlet of Tatsamenie Lake. Holding mortality for this group of fish was similar to that experienced at the Little Tatsamenie site ( $13 \%$ for females and $9 \%$ for males). These results indicate that it is possible to collect eggs from fish which would have spawned in Tatsamenie Lake, thus eliminating potential genetic concerns of stocking Tatsamenie Lake with fiy from the inter-connecting stream sub-stock or lake/stream hybrids.

The 1993 egg take target was increased from 1.75 million (1992 target) to 2.5 million. The egg-take target was not achieved due to lower than expected escapement; resulting in an adjusted green-egg estimate of 1.1 million. The average fertilization rate was $62 \%$, the lowest for this broodstock group to date. Based on the results of the pilot brood stock collection program carried out near the outlet of Tatsamenie Lake in 1993, it was decided to collect brood stock for this project at a weir at this location in future years. This eliminates any potential for genetic concern with regard to mixing of the two sub-stocks.

Survival at the hatchery for BY 1993 eggs was poor, $45 \%$ from green egg to planted fry. This was due to poor green to eyed egg survival and an IHNV loss of 169,000 alevins. Reasons for the poor green to eyed egg survival are unknown.

The otoliths of all BY 93 Little Tatsamenie stock were marked with a $5+5$ ring pattern. A total of 521,000 fry were planted on July 14, 1994.

## 1994 Brood Year

Brood stock for outplants to Tatsamenie Lake was captured in a weir at the outlet of Tatsamenie Lake. The total number of fish captured for brood stock in 1994, from an estimated escapement of 4,371 , was 1,035 , with 381 females and 332 males spawned. There were 51 female pre-spawn mortalities and 29 male prespawn mortalities for a combined total of $7.7 \%$ of held fish. Eggs were shipped in five lots during the period September 16 to October 20.

The egg-take target of 2.0 million was not achieved primarily due to low brood stock availability; the adjusted green egg estimate was 1.2 million eggs.

Survival at the hatchery was $73 \%$ from green egg to outplanted fry. The relatively poor survival was due in part to low survival to the eyed stage.

The otoliths of all 1994 brood Tatsamenie stock were marked with a 5 ring pattern. A total of 898,000 fry were planted on July 18 and 21, 1995.

## 1995 Brood Year

The total number of fish utilized for brood stock was 1,329 , made up of 726 females and 603 males captured from an estimated escapement of 8,000 fish. There were 26 female pre-spawn mortalities and 38 male pre-spawn mortalities. Eggs were shipped in 8 lots during the period September 15 to October 16. The fertilization rate was $84 \%$. Brood stock was readily captured because of the high escapement, the egg-take target of 2.5 million was almost met, with an adjusted green egg estimate of 2.4 million. Slightly lower than average fecundities were responsible for missing the egg take goal, original estimate was 2.6 million eggs. The otoliths of all 1995 brood Tatsamenie stock were marked with a 5 ring pattern and were planted in June of 1996 .

## Growth and Survival

## Juvenile Observations

As at Tahltan and Tuya Lakes, the numbers of fry remaining onshore in Tatsamenie Lake declined over the season in all years, as evidenced by declining beachseine catches (Tables 8 and 9). However, compared to Tuya and Tahltan Lakes, at Tatsamenie (and at Trapper Lake), relatively more fry remained onshore at the time of the fall surveys. In 1992, the June survey preceded the outplants, accounting for the absence of enhanced fry in the beachseine catches. Proportions of enhanced age $0+$ fry in the remaining catches in 1992 ranged from $2.2 \%$ to $7.3 \%$. The early August trawl catch was only 4 fry, all wild. The fall trawl catch was substantial and contained $7.6 \%$ enhanced fry. This was probably a better estimate of the enhanced proportion of the population.

In 1993, beachseine catches of age $0+$ fry contained 4.1 and $9.1 \%$ enhanced fry, the latter being based on a very small sample. The fall trawl catch was substantial, with $12.4 \%$ of the age $0+$ fry of enhanced origin.

In 1994, observed proportions of enhanced fish in age $0+$ fry were lower, $1.7 \%$ and $5.9 \%$ in beachseine catches and only $1.8 \%$ in the fall trawl sample (two of these were relatively small samples, Table 9). Reasons for this include the relatively low number outplanted and are discussed below.

In all years, outplanted fry grew well (Figures 7, 8) but wild fry were generally slightly larger at any time, probably indicating the presence of earlier emerging fry in the wild population.

Hydroacoustic estimates (Table 14, 15) indicate survivals from time of outplant to the time of fall surveys ranged from $3.6 \%$ to $29.9 \%$, with an average of $16.4 \%$. As at Tuya Lake, the accuracy of the Tatsamenie fall survival estimates based on hydroacoustics cannot be assessed, since there is no smolt enumeration program for comparison. Evaluating accuracy of the enhanced fry survival estimates at Tatsamenie Lake is
particularly difficult since the number of natural spawners in the lake is not accurately known, consisting of an uncertain portion of the enumerated Tatsamenie total system return ${ }^{3}$. Therefore, it is difficult to say whether the wild population survival estimates are reasonable, which would give some indication of accuracy of the enhanced fry estimates. Another complicating factor at Tatsamenie Lake is the presence of a relatively (in comparison to Tahltan and Tuya lakes) high, indeterminable number of fry remaining onshore at the time of the fall hydroacoustic estimates. Assuming hydroacoustic estimates are likely to be low, as they are at Tahltan Lake, for the years in which fall estimates are available it appears that fall fry survivals at Tatsamenie could be comparable to the average survival seen at Tahltan. The exception to this would be the 1993 brood year, when spring to fall fry survival was estimated at only $3.6 \%$. Examination of the data indicates this exceptionally low survival is real. Even allowing for $90 \%$ of the total enumerated 1993 Tatsamenie escapement spawning in the lake, and exceptionally high wild egg to fall fry survival ( $46 \%$, the extreme estimate for Tahltan Lake), enhanced fry survival could not have exceeded about $10.5 \%$, given the observed proportion of $1.8 \%$ enhanced fry in the trawl sample. Although based on quite a small sample ( 56 fish), this proportion appears to be realistic, since it agrees quite closely with the proportion of $2.3 \%$ enhanced fish in the age $1+$ smolts the following spring as discussed below (Table 18). As previously discussed, the 1993 brood-year survival was exceptionally low at Tahltan as well. There may have been a common factor involved, perhaps environmental or some aspect of fish culture procedures. It should also be noted however, that proportions of enhanced smolts outmigrating have been relatively low for all brood-years (1990-1994, Table 18); and particularly for BY 1992, do not always correlate with the enhanced fry ratios of the preceding fall.

## Smolt Observations

Details of the smolt emigrations observed at Tatsamenie Lake, 1992 through 1995, are given in Table 18. Enhanced smolts are in most cases smaller than wild smolts, as were the juveniles. Again, this probably indicates the presence of some early emerging fry in the wild population. Although mean smolt length and weight have varied considerably, it is difficult to relate these to juvenile in-lake abundance given the uncertainties about the wild spawner abundance and the hydroacoustic estimates. Overall, smolts are of a good size, comparable to those from Tahitan Lake (Figure 12).

Although no attempts have been made to enumerate the smolt run at Tatsamenie Lake, in general the enhanced/wild smolt ratios of age $1+$ fish (the numerically predominate age class), have been relatively low for all brood years, ranging from $2.0 \%-6.3 \%$. Although the number of wild spawners and their relative production are not known, it appears that enhanced production rates to date have been lower than the wild production rates. For the period 1990-1993 the number of fish used for broodstock purposes has ranged from $10 \%-16 \%$ of total annual escapements of Tatsamenie Lake and stream spawners combined. (PSC 1994). However, annual production of enhanced smolts from Tatsamenie Lake over the same period has ranged from $2.5 \%-6.3 \%$ (all age classes combined, Table 18). Several factors could account for the low enhanced smolt ratios and are discussed in section 4.4 .3 below. Regardless of the causes, indications are that the enhancement objectives of increased smolt production at Tatsamenie Lake are not presently being met.

[^2]
### 3.4.4 Trapper Lake

## Egg-take Through Outplant Activities

The first egg take was conducted at Little Trapper Lake in 1990. Results in 1990 and 1991 have been summarized in a previous report (PSC 1994). Egg take and outplant activities from 1992 through 1995 are summarized in Table 3 and detailed below. Detailed annual reports summarizing all Little Trapper egg take activities from 1990 through 1995 have been prepared by the contractor, and submitted to the DFO contracting authority (B. Mercer \& Associates Ltd. 1990-1995).

Fry for outplanting to Trapper Lake are incubated from eggs collected from Little Trapper Lake broodstock. Little Trapper Lake, located approximately 3 km downstream from Trapper Lake, has experienced average annual wild sockeye escapements of approximately 12,000 (1983-1995). Trapper Lake is not accessible to these anadromous spawners since the stream connecting the two lakes appears to be impassable to upstream migration of adult sockeye.

Little Trapper Lake broodstock were captured using a 35 m beach seine. The seine net was deployed around the mouth of the inlet stream where sockeye congregate prior to entering the inlet stream to spawn. Captured fish were held in net pens and sexually mature fish were sorted, removed, and spawned at intervals that precluded the formation of over-ripe gametes in held fish, and/or provided the requisite number of fertilized eggs for delivery to the hatchery.

## 1992 Brood Year

The total number of fish utilized for brood stock was 1,566 , with 784 females and 782 males spawned. A total of 1,080 females and 1,013 males were captured and held from an escapement of 14,372 fish. Prespawn holding mortality was 39 and $34(3.5 \%$ total) of held females and males respectively. Eggs were shipped in eight lots during the period August 29 to September 12. All eggs were shipped the day they were taken. The average fertilization rate was $90 \%$. The egg-take goal of 2.75 million was not met due to a misunderstanding between the field crew and hatchery crew regarding incubator loading density resulting in the collection of $2,521,000$ eggs ( $92 \%$ of the goal).

Survival at the hatchery for this group of eggs was $44 \%$ from green egg to outplanted fry. The poor survival was due to IHNV; approximately 917,000 alevins were destroyed when the prerelease screening was positive for the virus.

The otoliths of all 1992 BY Trapper fry were marked with a $7+3$ ring pattern, 7 rings in the-pre-hatch stage and 3 in the post-hatch stage (.Appendix 1). A total of 1,113,000 fry were planted June 25 and July 2, 1993 (Appendix 3 and Table 3).

## 1993 Brood Year

The total number of fish utilized for brood stock was 700 , with 350 females and 350 males spawned. A total of 646 females and 457 males were captured and held from an escapement of 17,432 fish. Male and female pre-spawn holding mortality was 18 and 23 fish respectively ( $3.7 \%$ of held fish total). Eggs were shipped in
three lots during the period August 29 to September 5 . The fertilization rate was not determined by hatchery personnel, however the green egg to planted fry survival was calculated to be $78 \%$. Brood stock was readily captured because of the high escapement, the egg-take target of 1.0 million was marginally exceeded, and the adjusted green egg estimate was 1.2 million.

Survival in the hatchery for this group of eggs was $78 \%$ from green egg to outplanted fry which is the highest in-hatchery survival to date for this stock and the first group of Little Trapper eggs not to have mortality associated with IHNV.

The otoliths of all 1993 BY Trapper fry were marked with a $4+5$ ring pattern, 5 rings prior to hatching with 5 rings post hatching (Appendix 1). A total of 916,000 fry were planted on June 16 and 24, 1994 (Appendix 3 and Table 3).

## 1994 Brood Year

The total number of fish utilized for brood stock was 704, with 353 females and 351 males. A total of 667 females and 477 males were captured and held from an escapement of 13,438 sockeye. Pre-spawn holding mortality was 12 females and 5 males ( $1.5 \%$ of total broodstock held). Eggs were shipped in three lots during the period August 30 to September 7. The fertilization rate was not calculated by hatchery personnel. Brood stock was readily captured because of the high escapement, the egg-take target of 1.0 million was marginally exceeded, and the adjusted green egg estimate was 1.06 million.

Survival in the hatchery for this group of eggs was $72 \%$ from green egg to outplanted fry, below the biostandard of $80 \%$. For the second year in a row, the Little Trapper origin eggs did not experience mortality associated with IHNV.

The otoliths of all 1994 BY Trapper fry were marked with a 7 ring pattern (Appendix 1). A total of 773,000 fry were planted on June 21 and 28, and July 3, 1995 (Appendix 3 and Table 3).

## Growth and Survival

Note: Since fry outplanted to Trapper Lake have been observed to migrate prematurely into Little Trapper Lake, this lake has been monitored quite closely as well. These observations are included in this section.

## Juvenile Observations

a) Trapper Lake

Unlike fry in other outplant lakes, fry outplanted into Trapper Lake remained abundant onshore throughout the season in some years, as evidenced by beachseine catches in 1992 and 1993 (Tables 10, 11; Figures 9, 10). This phenomenon did not occur in 1991 (PSC 1994) or 1994, when onshore abundance remained high throughout most of the season, but declined sharply in the final fall surveys. In 1992, all beachseine catches, including those for late September, were $100 \%$ age $0+$ enhanced fry. However, the September trawl catch contained $43 \%$ age $1+$ juveniles and approximately equal numbers of wild and enhanced fish. In 1993, fewer age $1+$ fry were observed in the trawl catch and the proportion of age $0+$ fry was much higher. Results were
similar in 1993, however, the proportions of older juveniles and wild fry in the September trawl catches was not as high. Results for 1991 were similar to 1992 but with fewer older fish in the September trawls. (PSC 1994). The only substantial sample collected in 1994 was in the July beachseine; in which the proportion of wild fry in this sample was slightly higher than in other years. Little can be said regarding the 1994 September trawl survey since the total number of fish captured was so small.

These results indicate outplanted fry generally remain onshore in substantial numbers throughout most or all of their first year in the lake but are probably entirely pelagic in their second and subsequent years. The much higher proportions of age $0+$ wild fish observed in trawl catches compared to beachseine catches suggests this behaviour differs from that of wild fish, with wild fish making much less use of onshore (littoral) regions. Sizes of wild and enhanced juveniles of the same ages were very similar (Figures 9, 10).

The validity of the August 1992 hydroacoustic estimate (Table 14) is doubtful due to the indeterminable numbers of fry onshore, where they would not have been accessible to the sonic gear. This figure almost certainly substantially underestimates the true population. This would be true for the September, 1993 estimate, also. The September, 1991 estimate (PSC 1994), and the September 1994 estimates, may be reasonably accurate since very few fry were observed onshore at the time of the surveys. Estimated enhanced age $0+$ fry survivals from time of outplanting to time of the fall surveys are given in Table 15. These are $24.8 \%$ for the 1990 brood year (based on 1991 surveys) and $8.5 \%$ for the 1992 brood year (1993 survey), although as mentioned above, the latter is probably substantially underestimated. The 1993 brood year survival estimate ( $0 \%$ ) cannot be considered valid because of the extremely small trawl catch ( 3 fish) used in apportioning the total population.

Regarding the 1993 brood year survival, it should be noted that while a valid fall survival estimate could not be made, survival appears to have been poor relative to other years. This is indicated in the 1994 fall surveys by the apparent absence or very low abundance of fry onshore, difficulties capturing fish in trawls, and the low total population estimate compared to other years, even allowing for differences in numbers outplanted. As discussed in previous sections, survival of the 1993 brood year appears to have been low at Tahltan and Tatsamenie lakes as well.

## b) Little Trapper Lake

As reported previously (PSC 1994) a sample of fry collected in Little Trapper Lake in July 1991, approximately 1 month after the first fry plants to Trapper Lake, contained $12 \%$ outplanted fry. This indicated a substantial amount of early outmigration had taken place. These fry were competing with wild fry for the already heavily utilized food supply in Little Trapper. Because of this, Little Trapper was surveyed once again in 1992 in order to more closely monitor what was occurring. In both 1993 and 1994 Little Trapper was surveyed two times each year, including hydroacoustic surveys in the fall. Results of these surveys are given in Tables 12 and 13 and Figure 11.

The beachseine sample collected at Little Trapper in 1992, approximately 1 month after completion of the outplants to Trapper Lake, contained only age $0+$ juveniles, $18 \%$ of which were enhanced. A similar July beachseine sample in 1993 contained no enhanced juveniles in either the age $0+$ or age $1+$ fish captured. This also occurred for a September beachseine sample as well; a trawl sample collected at the same time contained 1 enhanced age $0+$ juvenile, however this was less than $1 \%$ of the total sample. In 1994, a July
beachseine sample again contained no enhanced fry. In September there were no enhanced juveniles in either age $0+$ or age $1+$ trawl sampled fish, however one enhanced age $0+$ fry (from a sample size of 32), was observed in the beachseined juveniles.

These results indicate the amount of early outmigration was either much reduced in 1993, or outplant survival in Trapper Lake was correspondingly low. To allow more accurate comparisons of outmigration, the percentages of enhanced age $0+$ fry were standardized to allow for annual differences in 1 ) the number of fry outplanted to Trapper Lake, and 2) the relative abundance of wild fry in Little Trapper, using the number of wild female spawners the previous year as an index of abundance. Data was standardized relative to 1991 by dividing fry and spawner numbers by the 1991 values for these parameters ${ }^{4}$. The standardized percents were then calculated as:

## standardized percent $=$ observed percent $X$ standardized number of female spawners standardized number outplanted

The results are given in Table 23. The amount of early outmigration in 1991 and 1992 was approximately 2 to 4 times the maximum observed in following years. It should be noted that this maximum, $6.8 \%$ for the September 1994 beachseine sample, was based on a very small sample size and thus the true value could have been lower. It has been suggested that the decreased early outmigration in 1994 may have resulted from changes in hatchery procedures. Beginning in 1993, volitional emergence techniques were employed at Snettisham. With the outplanted fry at a more advanced stage of development they may have been less susceptible to being passively swept from Trapper Lake or may have reduced the tendency to premature outmigration. However, this would not account for the low numbers of enhanced fry observed in 1993 and it should be noted that a host of other environmental factors (water currents, forage base, discharge, temperature, etc.) could influence outmigration behavior of the fry. In addition, the low numbers of enhanced fry in Little Trapper Lake may simply be attributable to low outplant survival. The 1993 and 1994 Trapper Lake hydroacoustic estimates indicated enhanced fry numbers were relatively low (Table 14).

Hydroacoustic estimates for Little Trapper Lake fry are presented in Table 14. These probably substantially underestimate the true population because of technical problems associated with high juvenile densities.

## Smolt Observations

a) Trapper Lake

Results of smolt sampling at the outlet of Trapper Lake from 1992 to 1995 are summarized in Table 19. In 1992, the first year of expected smolt emigration (age 1+ fish from the 1991 outplant), no smolts were captured ${ }^{5}$. Effort was increased in subsequent years, far above that required at other lakes. Despite this, only one smolt was captured in 1993. Procedures were modified somewhat in 1994, which may have contributed to the capture of the first significant number of smolts (38). In 1995, 165 smolts were captured. The most noticeable feature of Trapper Lake smolts was the predominance of older fish, age $2+$ being predominant with substantial numbers of age $3+$, and even one age $4+$ smolt within the sample.

[^3]In spite of the highly glacial nature of Trapper Lake and the less than optimal zooplankton forage base, outplanted fish grew to a relatively large size. Age $1+$ smolts are comparable in weight to those from Tahltan and Tatsamenie Lakes. A possible explanation for this is that the tendency for fry to remain onshore is associated with feeding on benthic or terrestrial organisms in the littoral zone, which may be a superior food source to the pelagic region zooplankton forage. This phenomenon has been observed in other highly glacial lakes (K. D. Hyatt, Pacific Biological Station, personal communication). Analysis of stomach contents of captured fry would have to be performed to substantiate this hypothesis.
b) Little Trapper Lake

Results of smolt sampling at the outlet of Little Trapper Lake from 1992 to 1995 are summarized in Table 20. While the large majority of smolts were of wild origin, in all years there were low proportions of enhanced smolts originating from the outplants to Trapper Lake. A large component of wild outmigrants leaving Little Trapper is to be expected given that smolts leaving Trapper Lake must exit through Little Trapper Lake along with the wild smolts originating there. The exception is the relatively high ratio of enhanced age $1+$ smolts outmigrating in 1992. Given the observed early outmigration of fry from Trapper Lake into Little Trapper and the difficulties in capturing smolts leaving Trapper Lake, it seems likely at least some of these enhanced smolts may have reared in Little Trapper. Of relevance to this issue is the difference in the size of smolts known to have reared in each of the two lakes. The average weight of age $1+$ enhanced smolts captured leaving Trapper Lake is 4.3 g , range 3.4-6.0 (Table 19). The average weight of age $1+$ wild smolts leaving Little Trapper is 2.4 g , range 1.7-2.9 (Table 20); the average weight for age $1+$ enhanced smolts leaving Little Trapper is 2.2 g , range $1.3-3.2 \mathrm{~g}$. This would suggest that most of the enhanced smolts captured leaving Little Trapper are more likely to have reared in Little Trapper rather than in Trapper Lake. Results for age $2+$ smolts are similar.

The numbers of enhanced smolts captured leaving Little Trapper appear to have been low, as are the fall juvenile estimates. Assuming the fall hydroacoustic juvenile estimates to be reasonably accurate and ignoring any winter mortality, total smolt emigrations in 1993 and 1994 would have been approximately 300,000 and 555,000 respectively. The total number of enhanced smolts (ages 1 and 2 combined) in the 1993 and 1994 emigrations, based on the proportions observed, would have been approximately 8,400 and 16,000 , respectively. Having originated from a fry outplant of $1,811,000$ (1991 outplants) and 113,000 (1992 outplants), total fry to smolt survival would have been $1.7 \%$ and $4.5 \%$ for 1993 and 1994 respectively. Even assuming hydroacoustic estimates underestimated smolt migrations by $100 \%$ (the approximate maximum discrepancy observed at Tahltan Lake), fry to smolt survival is still only approximately $4 \%-9 \%$. Therefore, although smolt enumeration has not occurred it is apparent that enhanced smolt production from Trapper and Little Trapper lakes is very low.

The size of age $1+$ smolts emigrating from Little Trapper Lake is plotted against the estimated number of wild female spawners in Little Trapper for the brood year in Figure 16. Assuming fry numbers are directly proportional to spawner abundance, the number of spawners would provide an index of fry abundance during the year spent in the lake. Since there would also be some early emigrating fry from the outplants to Little Trapper, the number of fry from the same brood year planted into Trapper Lake is also shown. Results are irregular with 1993 being an anomaly, but for both wild and enhanced smolts there is an indication (statistically non-significant) of smaller smolts at higher densities. Therefore, although a negative correlation
exists between Little Trapper smolt size and Trapper Lake outplant numbers, evidence indicating Trapper Lake fry outplants have negatively impacted Little Trapper wild sockeye production (based on smolt size), is inconclusive.

### 3.5 Limnological Observations

### 3.5.1 Zooplankton ${ }^{6}$

In this section, zooplankton observations for each of the outplant lakes and those closely associated with them are examined for any impacts resulting from enhanced sockeye fry outplants. Nakina Lake, only recently surveyed for enhancement potential is also included (further details of the Nakina surveys are given in a later section, Ancillary Activities). Zooplankton observations are depicted for all the study lakes in figures 17 through 27. Tables listing the numerical values of zooplankton abundance, length, and biomass from which the figures are derived are located in Appendix 4.

Because sockeye fry typically feed on large bosminid and daphnid cladocerans (Goodlad et al. 1974), changes in zooplankton size and species composition in response to increased sockeye fry recruitment can be used to identify limits to potential sockeye smolt production in nursery lakes and can permit qualitative assessment of the impact of sockeye fry on the zooplankton forage base (Brooks and Dodson 1965; Galbraith 1967; O'Neill and Hyatt 1987).

## Tahltan Lake

Zooplankton data for 1985 through 1994 are presented in Figures 17 and 18. Zooplankton mean length and biomass have increased slightly in recent years, with relatively little change in total abundance (Figure 17). Increased natural escapements in recent years (1991-1994) may have raised lake nutrient levels and helped support this increased zooplankton production even at high sockeye juvenile densities. Species composition data indicate relatively little change among years (Figure 18). Increases in the proportion of daphnids relative to cyclopoids later in the growing season, particularly in 1993 and 1994, may suggest reduced levels of fry predation on more preferable (edible) prey, like cladocerans. Reduced predation pressure through reduced levels of juvenile sockeye may also be responsible for the increased abundance of diaptomids in 1993 and 1994. In summary, these results indicate the zooplankton community has been little impacted by increased densities of juvenile sockeye and appears to be capable of easily suppoting present levels of natural and supplemental fry.

## Tuya Lake

Zooplankton data for 1987 through 1995 are presented in Figures 19 and 20. Sockeye fry outplants to Tuya Lake began in 1992; prior to this there were no sockeye present and it is postulated from the observed plankton community structure that there was very little planktivorous predation. Tuya zooplankton are large and numerically abundant, with resulting high biomass (Fig. 19). Prior to 1994, the community was dominated by large predaceous heterocope and cyclopoid copepods and non-predaceous Diaptomid

[^4]copepods. Cladocerans (Daphnids, Bosmina) were in very low abundance (Fig. 20). At the moderate levels of fry outplanting in 1992 and 1993 this community structure was little impacted by sockeye predation. In 1994, however, with the almost twofold increase in the numbers of fry outplanted, the largest zooplankton were heavily cropped (ie. Heterocope and Diaptomid copepods), resulting in the community being strongly dominated by the smaller, but still relatively large, Cyclopoid copepeds. Their high abundance and large size resulted in a very high biomass, despite the increased predation by sockeye. In summary, results indicate that the Tuya Lake zooplankton community has not been negatively impacted by the fry outplants and could support levels even greater than that of 1994.

## Tatsamenie Lake

Zooplankton data for 1987 through 1995 are presented in Figures 21 and 22. Zooplankton in Tatsamenie Lake are moderate in size and numerically abundant (Fig. 21). Biomass is moderate to high and shows moderate variation between years, and appears to be unrelated to fry density as determined from hydroacoustic estimates. The proportions of cyclopoid copepods and cladocerans (Fig. 22) indicate a well balanced zooplankton community little impacted by predaceous copepods or fish predators. These results indicate Tatsamenie Lake, at least at current levels of wild production, is capable of supporting fry outplants larger than those made to date.

## Trapper and Little Trapper Lakes

Zooplankton data for 1987 through 1994 presented in Figures 23, 24, and 25. Outplants to Trapper Lake began in 1991. Little Trapper Lake is included in these analyses because of the observed premature emigration of outplanted fry from Trapper Lake into Little Trapper Lake and concerns about the impact of this on Little Trapper zooplankton. While total biomass is always lower in Little Trapper Lake, biomass levels for each lake do not exhibit large annual variations. Seasonal variation is also relatively consistent in each lake (Figure 23). Zooplankton densities for both lakes are consistent with those of pre-outplant years (Fig. 24). Mean zooplankton lengths have not declined in recent years, despite the outplants to Trapper Lake (Fig. 25). Species composition data for outplant years (1991 through 1994) are presented in Table 24. Zooplankton species composition differs between lakes: most notable is the scarcity of cladocerans (Bosmina sp., Daphnia sp.) in Trapper Lake, which is typical of highly glacial lakes. However, composition for both lakes has remained quite constant despite large variations in the numbers outplanted (the 1992 outplant was approximately twice that of any other). In summary, these results indicate the zooplankton communities in both Trapper and Little Trapper Lakes have been little impacted by the fry outplants. Trapper Lake could support higher fish densities. However, since the zooplankton forage in Little Trapper is already heavily utilized by wild fry and the impacts of increased predation may not be immediately evident, and in view of the early outmigration of outplanted fry from Trapper Lake in 1991 and 1992, caution should be used if the level of outplants to Trapper Lake are increased.

## Nakina Lake

Results of analyses of Nakina Lake zooplankton data collected in 1993 and 1994 are presented in Figures 26 and 27. Zooplankton is moderate in size and numerically abundant. Biomass is moderate to high with moderate variation over the growing season (Fig. 26). The predominance of calanoid copepods (Heterocopes
and diaptomids, Fig. 27) and lower levels of rotifers, nauplii and cladocerans indicate low cropping by limnetic fish populations. This is consistent with observations from Tuya Lake prior to sockeye fry outplants (see Ancillary Activities, below, for further results of the Nakina Lake surveys).

### 3.5.2 Productivity and carrying capacity estimates

All available Sechii depth data were examined and average depths for the period of observations calculated. New euphotic volume estimates were calculated using these averages and, from these, revised estimates of maximum adult production and spring fry carrying capacity (see Methods Section). Results are tabulated in Table 25, including those for recently surveyed Nakina Lake; additional information on the Nakina surveys is given under Ancillary Activities. There are no dramatic changes in production estimates from those previously reported (PSC 1988). Of most significance are the slightly lower estimates for Tuya and Tatsamenie. Recent experience in S.E. Alaska suggests that maximum spring fry stocking densities derived from the euphotic volume (EV) model are often too high to be stocked on an annual basis. Because of this, a conservative approach of stocking at one-half the EV model prediction has generally been adopted. The Enhancement Sub-committee agrees with this approach; (i.e. it is the Sub-committee's recommendation that a safe level of outplanting is $50 \%$ of that predicted by the EV model; impacts of outplants to this level should be carefully monitored and evaluated before proceeding further).

Other methods of estimating carrying capacity based on zooplankton community structure and abundance, lake nutrients, and phytoplankton production are being examined but these analyses are too preliminary to generate numeric estimates of carrying capacity. However, in terms of relative productivity (i.e. zooplankton production per unit area) for the lakes examined closely, Tuya and Tahitan appear to have similar levels of productivity which are greater than those for Tatsamenie or Trapper lakes.

### 3.6 Ancillary Activities

### 3.6.1 Tahltan Lake

## Broodstock Selectivity

Results of tagging studies conducted in 1990, 1993, and 1994 at Tahltan Lake are inconclusive regarding the potential loss of genetic diversity due to the selection of a narrow temporal component of the run (DFO Whitehorse, unpublished reports 1990, 1993, and 1994). The studies performed in 1990 and 1993 indicate no statistically significant difference in the temporal origin of the broolstock used during those brood years. However, the 1994 tagging study did indicate a difference. This discrepancy, although statistically valid, was not large and perhaps is not unexpected since the egg take operations do not fully span the natural spawning period. Similar tagging studies are scheduled to continue at Tahltan Lake providing a firmer data base to determine the extent that certain temporal and spatial segments of the total escapement are represented in the egg-take brood stock.

### 3.6.2 Tuya Lake

## Studies of Zooplankton Diel Migration Patterns

Studies of changes in diel migration patterns of the zooplankton in Tuya Lake in response to sockeye introductions continued. These studies consisted of sampling simultaneously at 7 different depths at mid-day, twilight, and full darkness within a 24 h period. This has been done once annually in the late summer/early fall of every year since 1990. Outplants began in 1992; prior to this there was virtually no piscivorous predation on the zooplankton. Since this work is not of immediate importance analysis of samples and data have had low priority and have been archived at the Pacific Biological Station in Nanaimo for future reference when needed.

### 3.6.3 Tatsamenie Lake

## Genetic Stock Analysis

During the course of sockeye egg-takes at Little Tatsamenie from 1990 to 1993, there was an overriding concern regarding the collection of broodstock from mixed stocks. Fish collected at L. Tatsamenie could potentially have been from a river spawning population that spawns in the interconnecting creek between Little Tatsamenie and Tatsamenie lakes. Although fry planted in Tatsamenie Lake have not been observed to exhibit this behaviour, the potential exists that fry originating from this creek spawning population (or mixed stock hybrids), may leave the system early. ${ }^{7}$

Tissue samples collected from the creek spawning and shoal spawning populations in 1993 were analyzed by starch gel electrophoresis. Allozyme allele frequencies at seven loci were calculated for both groups. Significant differences (G-test) existed between the two samples at two separate loci. The results of this analysis indicated that there is reproductive isolation of the two stocks, and thus two distinct sub-populations. In 1994, as a result of the genetic analysis, a temporary weir structure was installed at the outlet of Tatsamenie Lake, and a trial egg-collection conducted to deternine if broodstock collection activities could be relocated to this site. Broodstock collection was successful and egg-collection activities were relocated to this site in 1995.

## Enhancement Subcommittee Report to TTC Co-chairs

In June, 1994, the Enhancement Subcommittee reviewed the broodstock collection procedures within the Tatsamenie system with respect to the genetic characteristics of the stock and also broodstock collection options. Results were reported to the TTC co-chairs, together with recommendations for weir operations/broodstock acquisition (Enhancement Sub-committee of the Transboundary Technical Committee, June 16, 1994). Results and recommendations are summarized below:

[^5]Results:

1) The results of genetic analysis of tissue samples (GSI) collected from adult sockeye spawning in the interconnecting creek between L. Tatsamenie and Tatsamenie lakes, and samples collected at a barrier weir located at the outlet to Tatsamenie Lake indicated that there was reproductive isolation, and thus two distinct sub-populations.
2) Although behavioral problems have not been demonstrated for fry planted into Tatsamenie Lake, there is potential for fry from creek spawners or from a creek/lake hybrid to leave Tatsamenie Lake early.
3) Continued use of these two distinct sub-stocks for broodstock could compromise their genotypic discreteness. Specific attributes that have been selected for as a result of the reproductive isolation of these populations, within their specific environments, could be lost or diminished; subsequently reducing the overall fitness of each population, as well as the fitness of enhanced fish produced from the sub-stock crosses.

## Recommendations:

1) The egg-take activities should be entirely relocated from L. Tatsanenie to Tatsamenie Lake in 1995. This would require the construction of a temporary barrier weir. Based on the results of the 1994 program a more permanent structure (i.e. steel tripod weir) may later be established at this site. Operation of the Little Tatsamenie enumeration weir should be continued for one or two years after which its value will be considered.
2) The recommended egg-take goal for Tatsamenie of 2.5 million has been previously established by the TTC for 1994. This goal may be constrained by the escapement level to Tatsamenie Lake and by DFO protocol which limits broodstock collection to $30 \%$ of the spawning population.

Note: In a review of the Trapper Lake Project in February 1996, the Enhancement Subcommittee reported on the suspension of the Trapper Lake project in 1995 and the recommended expansion of the Tatsamenie Lake egg-takes to 5.0 million from 2.5 million, both to occur in 1995. This expansion did not take place due to failure to obtain approval from Province of British Columbia authorities. Details will be presented in a later report.

### 3.6.4 Trapper / Little Trapper Lakes

## Identification of Kokanee Population in Trapper Lake

From the time of first survey in 1987 it was known that Trapper Lake contained a small number of wild sockeye juveniles. The origin of these fish was uncertain, since the falls in the stream connecting Trapper and Little Trapper Lakes appear to be impassable, preventing upstream access of Little Trapper Lake origin anadromous sockeye into Trapper Lake. Two explanations are; either the falls were not impassable, or wild fry were in fact non-anadromous sockeye (kokanee). Since all sockeye fry outplanted to Trapper Lake were thermally marked, wild juveniles in samples collected during evaluation surveys could be distinguished by absence of a thermal mark. Differences in measurements of strontium $(\mathrm{Sr}) /$ calcium ( Ca ) ratios in the marked and unmarked otoliths were used to resolve the origin of the unmarked fish. In 1989, Kalish (1989) showed that trout eggs exposed to salt water during egg formation absorb the elevated levels of Sr present in sea
water into the yolk. The strontium is passed onto the progeny via incorporation into the otoliths, thereby increasing the $\mathrm{Sr} / \mathrm{Ca}$ ratio. Thus, progeny from parents of anadromous origin would have higher ratios than those of non-anadromous origin (e.g. kokanee, in the case of sockeye). A sample of 44 fry collected from Trapper Lake during evaluation surveys was examined for thermal marks; 38 were found to be marked, and therefore had to outplanted fry, the progeny of anadromous parents. The remaining 6 unmarked fish had to be of wild origin, of unknown parentage. The otoliths of the 6 unmarked fish and those of 6 marked fish were examined for $\mathrm{Sr} / \mathrm{Ca}$ ratio. In addition, the otoliths from 6 fish from a known kokanee population were examined. Ratios were lowest in the Trapper Lake unknowns and highest in the marked fry; ratios in the known kokance were intermediate. These results suggest very strongly that the unknown sockeye in Trapper Lake are resident kokanee rather than progeny of anadromous sockeye which might have managed to pass the barrier in the outlet stream.

## Enhancement Sub-committee Report to TTC Co-chairs

In March 1994, the Enhancement Sub-committee of the TTC prepared an extensive review of the Trapper Lake enhancement project for the TTC co-chairs (Enhancement Subcommittee of the Transboundary Technical Committee, March 3, 1994). This included an outline of the project, a summary of outplant evaluation results, a discussion of alternative enhancement projects, and recommendations for the future of the project. Conclusions and recommendations were as follows:

Conclusions:

1) Available outplant evaluation data are inconclusive, are subject to a high degree of uncertainty, and are as yet not sufficiently complete to allow an adequate understanding of the project's success. By the fall of 1994 the first returns of 4 year old enhanced Trapper Lake fish will have occurred and been evaluated, and additional limnological and juvenile stock assessment data will have been compiled. A further review of this project will be made at that time.
2) Review of zooplankton data, including 1993 data, for Little Trapper and Trapper lakes, does not indicate any significant negative impacts on zooplankton abundance or population structure as a result of enhancement activities to date.
3) Smolt production from the Trapper Lake program appears to be much less than anticipated when the project was planned. Returns of enhanced adults will likely be correspondingly low, resulting in a reduced cost-benefit for the project.
4) No alternatives are readily available within the Taku River drainage during the next several years to replace the Trapper Lake enhancement program if it is discontinued. A transfer of enhancement production to ongoing Stikine River enhancement prograns could be undertaken in the short-term, to increase returns of enhanced transboundary sockeye salmon and fully utilize incubation capacity at Snettisham Hatchery, but this would delay reaching the goal of producing annual returns of 100,000 enhanced adults to the Taku River.
5) Snettisham Hatchery has just completed a major retrofit. The retrofit and changes in hatchery procedures such as using volitional emergence, should improve in-hatchery survivals and fiy quality.

Recommendations

1) The Enhancement Subcommittee recommends the continuation of the Trapper Lake program at
the 1.0 million egg-take level in 1994. This level of program will contribute some returns to user groups (although below the agreed upon Enhancement Memorandum of Understanding), free-up money to hasten development of other existing or potential Taku River enhancement projects, and provide additional data on which to evaluate the Trapper Lake program.
2) Also recommend was an egg-take goal of 2.0 million for the Tatsamenie Lake program in 1994. Assessment and refinement of the Tatsamenie Lake program should be accelerated. This lake appears to offer the most potential to achieve the Taku enhancement goal within the shortest time period. The magnitude of egg-takes at this location should be increased as soon as possible, within broodstock limitations, to replace the lost production from reduced egg-takes at Little Trapper. Some of the budgetary savings which occur from reduced egg-takes at Little Trapper will be directed towards additional egg-take (two sites) and assessment efforts at Tatsamenie Lake.
3) There appear to be opportunities within the Taku River which are capable in the long-term of satisfying the current enhancement goal for this drainage. In particular, the potential of King Salmon, Nakina and Kuthai lakes should be investigated to the extent funding allows.
4) The regulatory process and biological assessment requirements that must be satisfied for any new enhancement programs to proceed should be investigated and action initiated where appropriate.

Note: At the recommendation of the Enhancement Subcomınittee egg-takes at Little Trapper were suspended in 1995, mainly because of continued uncertainties regarding success of the project. This decision was the subject of a report from the Enhancement Subcommittee to the TTC co-chairs in February 1996 (Enhancement Subcommittee of the Transboundary Technical Committee, February 5, 1996) and will be reported upon in a later report.

## Pen Rearing Study at Trapper Lake

In addition to the planting of 537,800 un-fed fry directly into Trapper Lake in 1995 (BY 1994), an experiment was conducted on a group of approximately 235,600 fiy to test the effects on fry to smolt survival of short term rearing in a floating net pen in Trapper Lake. These fry were marked with a strontium band on the otolith, in addition to the seven ring thermal band. They were placed in a pen on June 28 and fed for 22 days. Average weight of the firy increased from approximately 0.13 g to 0.39 g , an increase of 0.26 g (200\%).

An estimated 93,000 fry, averaging 0.39 g in weight, were released into the lake from the net pen on July 20. The releases were in lots of approximately 5,000 at several inshore locations around the lake. Release numbers are considered relatively accurate as all released fry were weighed with corresponding sub-samples weighed and the individual fry counted. The large difference between the hatchery estimate of fry placed in the pen and the actual number released is possibly due to the fish being released directly into the lake during the outplant, and/or fry escaping through the net mesh during the first few days of the experiment. Additional details of this experiment will be presented in future enhancement reports when the otoliths from outmigrating smolts and/or returning adults are analyzed.

### 3.6.5 Nakina Lake

Nakina Lake was surveyed for enhancement potential in 1993 and 1994, as described in section 2.5.1. Zooplankton analyses are given above and illustrated in Figure 27; other results can be summarized as follows:

- The relative productivity of Nakina Lake appears to be greater than that of Tatsamenie but slightly lower than that of Tahltan.
- No sockeye or other planktivorous fish have been captured in Nakina Lake; the only species which have been captured or observed are Arctic grayling, cottids, and burbot (freshwater lingcod).
The following observations are presented in Table 25:
- The lake can be classified as clear/stained; the average Secchi depth is 5.6 m and the euphotic depth estimate is 9.41 m .
- The surface area of Nakina Lake is 491 hectares, almost identical to that of Tahltan; the euphotic volume estimate is 46.2 EV units, approximately one-half that of Tahltan, reflecting the reduced light penetration.
- Estimated adult sockeye production potential for Nakina Lake, based on the EV model, is 113,000, approximately one-half that of Tahltan Lake.
- The euphotic volume potential is equivalent to outplants of about $5,667,000$ spring fry; although if the conservative approach previously described is taken, a 'safe' outplant level would be $50 \%$ of this, or $2,834,000$ fry (for a sustained adult production of 56,500 adults).


### 4.0 DISCUSSION

### 4.1 Hatchery Operations

The modifications to the main building of the hatchery, turning it into a permanent CIF, were completed in August, 1993. The new facility has provisions for much better isolation of separate sockeye salmon stocks, greatly reducing the risk of IHNV transmission between groups of fish, as well as several other improvements over the temporary CIF. These improvements include: 1) an oxygen generation system which can be used to reduce super saturation problems and increase dissolved oxygen concentrations; 2) more efficient and higher capacity electric water heaters; and, 3 ) an additional, much more sophisticated water chiller, which when used in concert with the original chiller greatly increases chilling capacity. The new CIF also has improved provisions for water hardness treatment using $\mathrm{CaCl}_{2}$.

Hatchery staff assume an egg to fry survival of $80 \%$. This assumption was exceeded for BY 1992, 1993, and 1994 Tahltan eggs destined for Tahltan Lake. The assumption was met for BY 1993 and 1994 eggs for Tuya and was nominally met for BY 1993 from Little Trapper. Poor hatchery survival for the 1992 brood Tahltan eggs destined for Tuya Lake were attributed to IHNV outbreaks. Eggs from BYs 1992 and 1994 from Little Trapper Lake had poor egg to fry survival, partly due to IHNV for BY 1992. The poor survivals of BY 1992 and 1993 Tatsamenie Lake eggs were primarily attributable to IHNV. BY 1994 Tatsamenie Lake egg survival was improved over the previous two years, but was still below the hatchery target of $80 \%$.

Outbreaks of IHNV at the Snettisham Hatchery have thus far not been obviously related to the prevalence of
the virus in the parental brood stock, which has fluctuated dramatically at each of the three lakes (refer to section 3.3 above, and Appendix 2). The lack of a clear relationship between incidence of the disease in the parents and offspring suggests that as long as the pathogen is present, outbreaks may take place regardless of the parental incidence of the disease. The probability of an outbreak might be better understood if a representative sample of incidence for each incubator was available, however such information would be very costly to obtain.

### 4.2 Egg-take Operations

### 4.2.1 Tahltan/Tuya lakes

All Tahltan Lake egg-takes have come close to or exceeded target levels, largely due to large escapements, and the ease of capture and holding of broodstock.

Concerns had been expressed regarding loss of genetic diversity due to selection of broodstock from a narrow temporal segment of the run. Results from tagging studies are inconclusive, but suggest that the broodstock captured is by and large representative of the entire run (see section 3.7.1). In terms of overall loss of genetic diversity and/or genetic drift, it should be noted there is still the potential that over successive generations of enhancement activity the enhanced component of the run will progressively increase due to the differential survival of the enhanced versus wild offspring.

### 4.2.2 Trapper Lake

There have been no major problems associated with the Little Trapper Lake egg-takes. Brood stock is readily available at recent escapement levels and all egg-takes have come close to or have exceeded the targets. Nevertheless, other concerns regarding this project resulted in the TTC enhancement sub-committee recommending the project be suspended, (see above memo from Enhancement Sub-committee to TTC cochairs, February 5, 1996, under 3.7.4 Ancillary Activities: Trapper/Little Trapper Lakes).

### 4.2.3 Tatsamenie Lake

Escapement levels have been too low in the past several years to allow collection of sufficient broodstock to meet egg-take targets. The possibility of adjusting management strategy to permit larger escapements has been examined, but is complicated by a number of factors such as run timing, stock identification/separation, and annual variations in run strength. Nevertheless, the ability of the contractor to meet expanded future egg take targets is predicated on the availability of sufficient broodstock. A review of egg-take operations, with emphasis on perceived genetic concerns regarding current broodstock collection methods, was done in the spring of 1994. As result, egg-take operations in 1994 were moved from Little Tatsamenie to Tatsamenie Lake (see memo from Enhancement Sub-committee to TTC co-chairs, under 3.7.3 Ancillary Activities: Tatsamenie Lake).

### 4.3 Otolith Marking and Reading

### 4.3.1 Alaska

Otolith marking at Snettisham went well for BY's 1992 through 1994. The hatchery staff worked closely with the otolith lab, providing voucher samples and thermal records for mark validation.

During the period covered by this report, the otolith laboratory gained experience in thermal otolith marking, sampling, and processing. This experience is required in order to successfully recover thermal marks from adult sockeye returning to the Alaskan commercial fisheries. Strategies were developed to identify sample sizes needed to optimize laboratory effort with the objective of minimizing overall uncertainty on the numbers of enhanced fish captured. Other activities included development of an integrated database system which includes an inventory control process using bar-code labeling of samples.

### 4.3.2 Canada

In 1995, all Canadian transboundary juvenile, adult, and smolt otoliths were processed at the newly established otolith Lab in Nanaimo, B.C.. At present the otolith work is funded directly by a small number of projects, however, the eventual goal is to develop the otolith laboratory as a support service. Co-operation between this lab and that at Juneau has been excellent and the paired otoliths processed at one lab are available for independent reading by the other, if requested.

### 4.4 Growth, Survival, and Limnology

### 4.4.1 Talltan Lake

Outplanted fry have grown and survived well, the exception being the poor survival of the 1993 BY (1994 outplant). However, similar poor survivals of outplanted fiy for this brood year were noted at Tatsamenie and Trapper Lakes as well. It is possible there may have been a common factor involved, not specific to Tahltan Lake. The average survival to age $1+$ smolt is $18.7 \%$ or $22.1 \%$, excluding the 1993 brood year. A $20 \%$ fiy to smolt survival biostandard is used in planning the transboundary enhancement projects. Smolt size and zooplankton have shown little impact from the increased fry densities, which in one year exceeded the maximum predicted by the euphotic volume model. Productivity of Tahltan Lake is judged to be very high, similar to that of Tuya Lake, and well in excess of other outplant lakes.

### 4.4.2 Tuya Lake

Outplanted fry have grown exceptionally well, as was predicted from the rich zooplankton forage base, which was characterized by low predation levels by planktivorous fish, prior to the first fry outplant. Smolts from Tuya Lake are approximately twice or more the size of enhanced smolts from any other lake, including Tahltan Lake, the source of the parental stock for the Tuya Lake outplants. Hydroacoustic estimates of fall fry indicate survival to that stage was similar for BY1991 and BY1992 outplants, however, survival for the 1992 brood year, may have been considerably lower. It should be noted that hatchery survival for this group was low as well, as a result of IHNV problems. The poor survival of the 1993 BY outplants noted at other lakes was not apparent at Tuya. Smolt emigrations have not been enumerated at Tuya Lake. However, using
fall fry hydroacoustic estimates as a predictor of smolt numbers, there is no reason to believe survivals to the smolt stage in Tuya have not been as good or better than at Tahltan, since smolt estimates obtained by this method at Tahltan have been shown to be far too low. The zooplankton community structure has been altered by the fry outplants, however this impact has not been negative. The community has now assumed a structure, with a reduced abundance of large predaceous copepods and an increased abundance of nonpredaceous copepods as well as the faster reproducing Cladocerans, (Bosmina sp. and Daphnia sp.), a desirable sockeye forage base. Productivity of Tuya Lake is judged to be very high, similar to that of Tahltan Lake, and results indicate Tuya could support levels of outplanting greater than those adopted thus far.

### 4.4.3 Tatsamenie Lake

Growth of outplanted fry In Tatsamenie Lake has been good, with smolts similar in size to those from Tahltan Lake. Migrating smolts have not been enumerated and survival has been very difficult to estimate. There have been problems obtaining hydroacoustic estimates to calculate survival to the fall fry stage; these include equipment problems and the fact that the number of natural spawners in Tatsamenie Lake was not accurately known in the years involved. The estimated survival to the fall fry stage for BY 1991 through BY 1993, as determined from hydroacoustic estimates and trawl surveys, could be comparable to that for Tahltan Lake. However, estimates of proportions of emigrating enhanced smolts for BY 1990 through BY 1993 are relatively low. Even though the wild sockeye production in Tatsamenie Lake is not known with certainty, the low proportion of emigrating enhanced smolts appears to indicate either a low enhanced fry to smolt survival, or premature emigration of enhanced fish prior to spring smoltification.

The zooplankton in Tatsamenie Lake appears to be lightly impacted by the fry outplants and remains a well balanced community favorable to sockeye production. The productivity of Tatsamenie Lake is judged to be less than that of Tahltan or Tuya but greater than that of Trapper; at current levels of wild production it appears to be capable of supporting levels of outplanting greater than those used to date.

### 4.4.4 Trapper and Little Trapper Lakes

Growth of fry outplanted to Trapper Lake has been good, with smolts similar in size to those from Tahltan and Tatsamenie lakes. This has occurred despite the glacial conditions and less than optimal zooplankton forage; with cladocerans, the preferred food, being almost totally absent. As discussed previously, this fry growth may be associated with the tendency for fry to remain onshore, where benthic organisms may provide a food source superior to the pelagic region zooplankton forage. The zooplankton communities in both Trapper and Little Trapper lakes have been little impacted by the outplants; although it should be noted that impacts may be difficult to detect immediately in Little Trapper Lake, where the forage base is already heavily utilized by wild fry.

It has not been possible to obtain good survival estimates at Trapper Lake. Hydroacoustic estimates of fall fry abundance have been complicated by the tendency for fry to remain onshore where they are inaccessible to sounding gear and the difficulty in obtaining representative trawl samples. Another complicating factor has been the tendency for indeterminable numbers of fry to exit the lake prematurely as indicated by the presence of BY 1990 and BY 1991 enhanced fry in Little Trapper Lake. However, in subsequent brood years the number of enhanced fry rearing in Little Trapper is quite low, likely indicating that premature or passive emigration is no longer occurring. The final complicating factor in determining survival has been the
difficulties in capturing smolts leaving Trapper Lake. The reason for this is uncertain. Aside from possible low survival there are a number of other factors listed below which may influence catches:

1) Since smolt samples consistently show the majority of fish tend to remain in the lake at least two years, only a small number would be expected to have migrated in 1992, following the first outplant in 1991.
2) In 1991 and 1992, the first two years of outplanting, there was substantial early outmigration of fry into Little Trapper Lake which would have affected Trapper Lake smolt emigrations arising from these outplants, most notably in 1992 through 1994.
3) The more advanced fry resulting from volitional emergence, initiated with the 1992 BY (1993 outplants), may have survived at significantly higher levels, resulting in correspondingly greater smolt catches in later years.
4) Since outplants to Trapper have been relatively small ( 0.9 to 1.1 million, the outplant of 1.8 million in 1992 being exceptional) and since there are no wild smolts, total smolt abundance would generally be considerably less than at other lakes, making capture more difficult. The first significant capture ( 38 smolts) was in 1994, the predominant 2 -yr-old age class in this migration being associated with the large outplant in 1992.
5) Trapping methods were modified in 1994, the first year of significant captures. The most notable changes were the use of two nets instead of one and placement of nets in mid-strean as a result of favourable water conditions. Similar techniques were used in 1995.
6) Because of the predominance of older, larger, smolts in the Trapper emigration, as well as the relatively low density of outmigrants, the ability to avoid capture could be greater than at other lakes.

Given the problems in obtaining survival estimates it cannot be said with certainty whether survivals at Trapper Lake differ greatly from those observed at other lakes. However, the most reasonable interpretation of available information is that it has been substantially lower.

In summary, there are questions and uncertainties regarding the Trapper Lake outplants, notably:

1) Strong indications that survival has been low.
2) Possible negative impacts to the wild sockeye fry in Little Trapper Lake through competition for a limited food supply, should early outmigration of outplanted fry into Little Trapper Lake recur.
3) Uncertainties regarding the suitability of the fry forage base of Trapper Lake, even if survivals have been better than estimated.

It was because of these concerns that the Enhancement Subcommittee recommended there be no egg-takes at Little Trapper in 1995, and that this suspension remain in effect until adult return data becomes available to assist in a final evaluation.

### 4.4.5 Nakina Lake

Nakina Lake offers considerable potential for sockeye outplanting and could perhaps be used to replace the production lost through suspension of the outplants to Trapper Lake. Estimated adult sockeye production potential for Nakina Lake, based on the EV model, is 113,000, approximately one-half that of Tahitan Lake,
and very similar to that for Trapper Lake (Table 25). If the recommended conservative $50 \%$ approach is used, the production potential is 56,500 adults, which could be achieved by outplants of approximately $2,834,000$ spring fiy. Issues that would have to examined before proceeding would include a disease profile on resident species, source of broodstock, impact on resident species of outplants, ability to effectively harvest returns including interceptions of co-migrating sockeye stocks and other salmonids, including steelhead, and possibly a full environmental review.

### 5.0 SUMMARY AND RECOMMENDATIONS

### 5.1 Hatchery Operations

The permanent Central Incubation Facility constructed at Snettisham in 1993 is working well, with improved capabilities for treating water and isolating stocks. The occasional occurrence of IHNV indicates the continued need for strict observance of sockeye culture protocols.

### 5.2 Otolith Marking and Reading

The thermal mass marking of otoliths has proved successful and provides a very effective management tool. Efforts should be made to develop an integrated database system, including Canadian data, to increase utility and avoid duplication of marks. The timeliness of results produced by the otolith lab in Nanaino could be improved upon.

### 5.3 Tahltan Lake Outplant Project

There have been no problems meeting egg-take goals and outplanted fiy have grown and survived well. The maximum carrying capacity of Tahltan Lake has not been defined. The lake is capable of supporting current levels of outplanting, at least on an annual basis, but abnormally high wild fiy production could result in fry densities which may not be sustainable on a continued basis. Caution should therefore be used when determining the outplant numbers.

### 5.4 Tuya Lake Outplant Project

Tahltan Lake provides a ready source of broodstock for outplants to Tuya. As expected, outplanted fry have grown extremely well in Tuya Lake. Survival, although not precisely determined, appears to be good. Final confinmation of this depends on adult returns. The lake appears capable of supporting outplants considerably in excess of those to date. However, it would be prudent to proceed cautiously in order to properly consider changes in zooplankton community structure resulting from the outplants.

### 5.5 Tatsamenie Lake Outplant Proiect

Escapement levels in several years have been too low to allow collection of sufficient broodstock to meet objectives. Relocation of the egg-take site from Little Tatsamenie to Tatsamenie Lake, while alleviating genetic concerns, may further reduce broodstock availability. It is recommended that the possibility of special fishery management strategies be considered to allow greater escapements. Outplanted fry have grown well;
survival has been difficult to determine, but indications are that enhanced smolt emigration is less than expected. Confirmation by adult returns is required. Tatsamenie Lake appears capable of supporting outplants considerably in excess of those to date at current escapement levels; however, increased wild production in combination with outplants could conceivably tax the fry carrying capacity of the lake. The number and relative production of natural spawners may have to be considered when determining outplant targets.

### 5.6 Trapper Lake Outplant Proiect

There have been no problems obtaining sufficient eggs to meet objectives, however this stock appears to have a greater susceptibility to IHNV breakouts than others. Outplanted fry have grown well, but it has not been possible to determine survival with any degree of accuracy. Because of this, and concerns about early outmigration of outplanted fry and resulting possible impacts on wild stocks, egg-takes were suspended in 1995. It is recommended this suspension remain in effect until adult returns are assessed. It is also recommended the lost production be replaced by increasing outplants to Tatsamenie, and/or by initiating outplants to other lakes, Nakina Lake being one candidate.

### 5.7 Benefit / Cost Estimates

No attempt was made to estimate benefit / cost ratios in this report. One reason for this is lack of sufficient data on adult returns to estimate adult production, this being the true measure of survival and benefit. It is recommended these analyses should be done in the near future, using recently acquired adult return data and actual costs of the projects.

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FIGURES


Figure 1. Stikine River drainages.


Figure 2. Taku River drainages.


Figure 3. Tahltan Lake surveys, 1992; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ ethanol.


Figure 4. Tahltan Lake surveys, 1993 and 1994; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ or $90 \%$ ethanol, 1993 and 1994 respectively.


Figure 5. Tuya Lake surveys, 1992; mean lengths of juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. All fish are enhanced as there are no wild sockeye in Tuya Lake. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples. Measurements are from fish preserved in denatured $94 \%$ ethanol


Figure 6. Tuya Lake surveys, 1993 and 1994; mean lengths of juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. All fish are enhanced as there are no wild sockeye in Tuya Lake. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples. Measurements are from fish preserved in denatured $94 \%$ or $90 \%$ ethanol, 1993 and 1994, respectively.


Figure 7. Tatsamenie Lake surveys, 1992; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ ethanol.


Figure 8. Tatsamenie Lake surveys, 1993 and 1994; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ or $90 \%$ ethanol, 1993 and 1994, respectively.


Figure 9. Trapper Lake surveys, 1992; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ ethanol.


Figure 10. Trapper Lake surveys, 1993 and 1994; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ or $90 \%$ ethanol, 1993 and 1994, respectively.


Figure 11. Little Trapper Lake surveys, 1992, 1993 and 1994; mean lengths and percentages (in brackets) of enhanced and wild juvenile sockeye in trawl (tr.) and beachseine (bs.) samples. Small length bars ( 1 mm ) indicate 0 fish in the category. $95 \%$ confidence limits are given except for single fish samples and $0 \%$ or $100 \%$ values. Measurements are from fish preserved in denatured $94 \%$ (1992 and 1993) or $90 \%$ (1994) ethanol.


Figure 12. Average weights (g) of enhanced and wild age $1+$ smolts for each lake over all years of observation, 1991 to 1995 (see Tables 16-20 for annual observations for each lake).


Figure 13. Tahltan Lake smolt size versus estimated juvenile abundance in the first year of lake residency. Data is from the 12 year period of 1982 to 1993 (Table 22). A (1) indicates 1 year of rearing under fertilized lake conditions; (2) indicates 2 years of rearing under fertilized lake conditions.


Figure 14. Regression between Tahltan Lake fall hydroacoustic juvenile sockeye estimates and associated smolt estimates. The regression is based on hydroacoustic estimates from 1985, 1987, 1988, 1990, 1991, 1993 and 1994; the 1990 estimate shown was excluded from the regression calculation.


Figure 15. Tahltan Lake smolt emigrations, 1991 (the first year of enhanced smolt production) through 1995; estimated numbers of enhanced and wild smolts, ages combined.


Figure 16. Little Trapper Lake age 1 smolt size in relation to in-lake rearing density, as estimated by the number of wild female spawners for the broodyear. Numbers of fry outplanted to Trapper Lake for the same broodyear are also shown.


Figure 17. Tahltan Lake. Mean zooplankton abundance (density: \#/L), mean length ( mm ), and total biomass (ug/L) over the summer growing season during 1985, 1987 to 1994, contrasted against estimated total sockeye fry densities (mid June) (1994 estimates not available). $95 \%$ Confidence Intervals are indicated on data points and bars.


Figure 18. Tahltan Lake. Proportions of total zooplankton biomass comprising bosminids (Bosmina sp.), daphnids (Daphnia, Ceriodaphnia, Holopedium sp.), cyclopoids (Cyclops sp.), and diaptomids (Dipatomus sp.). Nauplii and rotifers (not shown) comprise less than $2 \%$ of total biomass in all years.


Figure 19. Tuya Lake. Mean zooplankton abundance (density: \#/L), mean length (mm), and total biomass ( $\mathrm{ug} / \mathrm{L}$ ) over the summer growing season during 1987 to 1995, contrasted against estimated total sockeye fry densities (mid June). $95 \%$ Confidence Intervals are indicated on data points and bars.


Figure 20. Tuya Lake. Proportions of total zooplankton biomass comprising bosminids (Bosmina sp.), daphnids (Daphnia, Ceriodaphnia, Holopedium sp.), cyclopoids (Cyclops sp.), and diaptomids (Dipatomus sp.). Nauplii and rotifers (not shown) comprise less than $1 \%$ of total biomass in all years.



*     - Fall Fry Density (Hydroacaustic / Trowl Survey)
$\nabla$ - Sockeye Fry Outplanting Density

Figure 21. Tatsamenie Lake. Mean zooplankton abundance (density: \#/L), mean length ( mm ), and total biomass (ug/L) over the summer growing season during 1987 to 1995, contrasted against estimated total sockeye fry densities (mid June). 95\% Confidence Intervals are indicated on data points and bars.


Figure 22. Tatsamenie Lake. Proportions of total zooplankton biomass comprising bosminids (Bosmina sp.), daphnids (Daphnia, Ceriodaphnia, Holopedium sp.), cyclopoids (Cyclops sp.), and diaptomids (Dipatomus sp.). Nauplii and rotifers (not shown) comprise less than $4 \%$ of total biomass in all years.


Figure 23. Trapper / Little Trapper Lakes. Total zooplankton biomass over the summer growing season for Trapper Lake (1987 to 1994; 2 sampling sites starting 1988) and Little Trapper Lake (1987 to 1991 and 1993 to 1994; 2 sampling sites starting in 1989).


Figure 24. Trapper / Little Trapper Lakes. Total zooplankton abundance over the summer growing season for Trapper Lake (1987 to 1994; 2 sampling sites starting 1988) and Little Trapper Lake (1987 to 1991 and 1993 to 1994; 2 sampling sites starting in 1989).


Figure 25. Trapper / Little Trapper Lakes. Total zooplankton mean length (mm) over the summer growing season for Trapper Lake (1987 to 1994; 2 sampling sites starting 1988) and Little Trapper Lake (1987 to 1991 and 1993 to 1994; 2 sampling sites starting in 1989).


Figure 26. Nakina Lake. Mean zooplankton abundance (density \#/L), mean length (mm), and total biomass (ug/L) over the summer growing season during 1993 and 1994.


Figure 27. Nakina Lake. Proportions of total zooplankton biomass comprising cladocera (Bosmina and Daphnia sp.), nauplii, rotifers, cyclopoids (Cyclops sp.), calanoids (Heterocope sp.) and diaptomids (Dipatomus sp.). Nauplii and rotifers comprise less than $1 \%$ of total biomass in all years.

TABLES

Table 1a. Summary of results of Tahltan Lake outplant projects, egg-take to outplanted fry stage.

| Broodyear | $\begin{gathered} \text { \# eggs taken }{ }^{\mathrm{a}} \\ \quad(\mathrm{x} 1000) \end{gathered}$ | $\begin{aligned} & \text { \# fry planted } \\ & \text { (x1000) } \\ & \hline \end{aligned}$ | percent <br> fertilized | Survival |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | fertilized egg to planted fry | green egg to planted fry |
| $1989{ }^{\text {b }}$ | $\begin{aligned} & 2,955 \\ & (3 \mathrm{M}) \end{aligned}$ | 1,042 | $70 \%$ | $50 \%$ | $35 \%$ |
| 1990 | $\begin{aligned} & 4,511 \\ & (5 \mathrm{M}) \end{aligned}$ | 3,585 | $82 \%$ | $96 \%$ | $79 \%$ |
| 1991 | $\begin{aligned} & 4,246^{\circ} \\ & (5-6 \mathrm{M}) \end{aligned}$ | 1,415 | $95 \%$ | $98 \%$ | $94 \%$ |
| 1992 | $\begin{aligned} & 2,154^{d} \\ & (5.4 \mathrm{M}) \end{aligned}$ | 1,947 | 92\% | $98 \%$ | $90 \%$ |
| 1993 | $\begin{gathered} 969^{\circ} \\ (6.0 \mathrm{M}) \end{gathered}$ | 904 | $n / a$ | n/a | $93 \%$ |
| 1994 | $\begin{aligned} & 1,326^{\mathrm{f}} \\ & \text { (6.0M) } \end{aligned}$ | 1,143 | n/a | n/a | $86 \%$ |
| 1995 | $\begin{aligned} & 3,008^{8} \\ & (6.0 \mathrm{M}) \end{aligned}$ |  | $95 \%$ |  |  |

${ }^{2}$ Egg-take targets in millions (M) are shown in parentheses
${ }^{\mathrm{b}}$ The values given here for BY 1989 differ slightly from those reported previously (PSC 1991) as a result of minor corrections to the data.
${ }^{\text {c }}$ This value includes eggs taken for outplants to both Tahltan and Tuya lakes; eggs are divided at the eyed stage and percent fertilized is therefore the same for both groups.
${ }^{d}$ This value includes eggs taken for Tahltan Lake only; total number of eggs collected in 1992 was 4,901,000.
${ }^{\text {e }}$ This value includes eggs taken for Tahltan Lake only; total number of eggs collected in 1993 was $6,140,000$.
${ }^{i}$ For return to Tahltan Lake only; total number taken was 4,182,000.
${ }^{\mathrm{c}}$ For return to Tahltan Lake only; total number taken was $6,891,000$.

Table 1 b . Summary of results of Tuya Lake outplant projects, egg-take to outplanted fry stage.

| Broodyear | $\begin{aligned} & \text { \# eggs taken }{ }^{a} \\ & (\times 1000) \end{aligned}$ | $\begin{aligned} & \text { \# fry planted } \\ & (\times 1000) \end{aligned}$ | percent fertilized | Survival |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | fertilized egg to planted fry | green egg to planted fry |
| 1991 | $\begin{aligned} & 2,732^{\mathrm{b}} \\ & (5-6 \mathrm{M}) \end{aligned}$ | 1,632 | 95\% | $63 \%$ | $60 \%$ |
| 1992 | $\begin{gathered} 2,747^{c} \\ (5.4 \mathrm{M}) \end{gathered}$ | 1,990 | 92\% | 78\% | 72\% |
| 1993 | $\begin{aligned} & 5,171^{\mathrm{d}} \\ & (6.0 \mathrm{M}) \end{aligned}$ | 4,691 | n/a | n/a | $91 \%$ |
| 1994 | $\begin{aligned} & 2,765^{e} \\ & (6.0 \mathrm{M}) \end{aligned}$ | 2,267 | 87 \% | $94 \%$ | $81 \%$ |
| 1995 | 3,883 ${ }^{\text {f }}$ |  | 95\% |  |  |
|  | (6.0M) |  |  |  |  |

${ }^{2}$ Egg-take targets in millions (M) are shown in parentheses
${ }^{\text {b }}$ This value includes eggs for Tuya only, total number of eggs taken at Tahltan in 1991 was $4,246,000$.
${ }^{c}$ This value includes eggs taken for Tuya planting only; the total number of eggs taken in 1992 was $4,901,000$.
${ }^{\mathrm{d}}$ This value includes eggs taken for Tuya planting only; the total number of eggs collected in 1993 was $6,140,000$.
${ }^{0}$ This value includes eggs taken for Tuya planting only; the total number of eggs taken in 1994 was $4,182,000$.
${ }^{f}$ This value includes eggs taken for Tuya planting only, the total number of eggs taken at Tahltan in 1995 was $6,891,000$.

Table 2. Summary of results of Tatsamenie Lake outplant projects, egg-take to outplanted fry stage.

| Broodyear | $\begin{gathered} \text { \# eggs taken }{ }^{\text {a }} \\ (\times 1000) \end{gathered}$ | $\begin{aligned} & \text { \# fry planted } \\ & (\times 1000) \\ & \hline \end{aligned}$ | percent <br> fertilized | Survival |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | fertilized egg to planted fry | green egg <br> to planted fry |
| 1990 | $\begin{gathered} 985 \\ (2.5 \mathrm{M}) \end{gathered}$ | 673 | $78 \%$ | $88 \%$ | 68\% |
| 1991 | $\begin{aligned} & 1,360 \\ & (1.25- \\ & 1.5 \mathrm{M}) \end{aligned}$ | 1,232 | $93 \%$ | 98\% | $91 \%$ |
| 1992 | $\begin{gathered} 1,486 \\ (1.75 \mathrm{M}) \end{gathered}$ | 909 | 86\% | $71 \%$ | $61 \%$ |
| 1993 | $\begin{gathered} 1,144 \\ (2.5 \mathrm{M}) \end{gathered}$ | 521 | n/a | $\mathrm{n} / \mathrm{a}$ | $45 \%$ |
| 1994 | $\begin{gathered} 1,229 \\ (2.5 \mathrm{M}) \end{gathered}$ | 898 | n/a | n/a | 73\% |
| 1995 | $\begin{gathered} 2,408 \\ (2.5 \mathrm{M}) \end{gathered}$ |  | $84 \%$ |  |  |

[^6]Table 3. Summary of results of Trapper Lake outplant projects, egg-take to outplanted fry stage.

| Broodyear | \# eggs taken ${ }^{\text {a }}$ (x1000) | \# fry planted (x1000) | percent fertilized | Survival |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | fertilized egg to planted fry | green egg <br> to planted fry |
| 1990 | $\begin{array}{r} 2,314 \\ (2.5 \mathrm{M}) \end{array}$ | 934 | 87\% | $47 \%$ | $41 \%$ |
| 1991 | $\begin{aligned} & 2,953 \\ & (3 \mathrm{M}) \end{aligned}$ | 1,811 | 85\% | $72 \%$ | $61 \%$ |
| 1992 | $\begin{gathered} 2,521 \\ (2.75 \mathrm{M}) \end{gathered}$ | 1,113 | $90 \%$ | $49 \%$ | 44\% |
| 1993 | $\begin{gathered} 1,174 \\ (1.0 \mathrm{M}) \end{gathered}$ | 916 | n/a | n/a | $78 \%$ |
| 1994 | $\begin{gathered} 1,062 \\ (1,0 \mathrm{M}) \end{gathered}$ | 773 | n/a | n/a | $72 \%$ |
| 1995 | egg takes discontinued |  |  |  |  |

[^7]Table 4. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Tahltan Lake during the summer and fall of 1992. ${ }^{\text {a }}$


[^8]Table 5. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Tahltan Lake during the summers and falls of 1993 and 1994. There was no juvenile sampling (beachseining or trawling) conducted on Trip 1 in either year. ${ }^{\text {a }}$


[^9]Table 6. Numbers captured, mean lengths and weights, and percentages of fish of different ages in samples of sockeye salmon juveniles from surveys of Tuya Lake during the summer and fall of 1992. There are no wild sockeye in Tuya Lake and the first fry plant was in 1992 (brood year 1991).


[^10]Table 7. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Tuya Lake during the summers and falls of 1993 and 1994. There was no juvenile sampling (beachseining or trawling) conducted on Trip 1 in either year. There are no wild sockeye in Tuya Lake ${ }^{\text {a }}$.


1994 Surveys
Trip 2, July 28

Age $0+$ (BY 1993)
Length
Weight
n

Age 1+ (BY 1992)

Trip 3, Sept. 5

Age 0t (BY 1993)
Length 60.5

Weight 2.02
n 75 n/a
\% 100 n/a

Age $1+$ (BY 1992)
n

| Total Catch: | 131 |
| :---: | ---: |
| Total Sampled: | 75 |
| 60.5 |  |
| 2.02 | n/a |
| 75 | n/a |
| 100 |  |
|  |  |
| 0 | n/a |

Total Index Catch: 1 Total Sampled:
38.0
0.37

1 n/a
100 n/a

0 n/a

0 sockeye captured
${ }^{a}$ Measurements are from specimens preserved in denatured ethanol and are not directly comparable to fresh measurements. Total sanipled refers to no. fish sampled for thenmal marks Total beachseine catches are for index sites only and do not include supplemental catches (see Methods, section 3.4). In some cases, sub-samples were selected from the total sampled, proportional to numbers caplured in individual sets or trawls. Percentages may not agree exaclly with sample sizes for length/weight ( n ) since some fish observed for marks may not have been measurable.

Table 8. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Tatsamenie Lake during the summer and fall of 1992. ${ }^{\text {a }}$

|  | Trawl |  | Beachseine |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Enhanced | Wild | Enhanced wi | ild |
| Trip 1, June 21 |  |  |  |  |
|  | No trawls cond | ted | Total Index Catch: | : 917 |
|  |  |  | Total Sampled: | 310 |
| Age 0+ (BY 1991) |  |  |  |  |
| Length |  |  |  | 33.4 |
| Weight |  |  |  | 0.24 |
| n |  |  | 0 | 44 |
| \% |  |  | 0 | 100 |
| Age 1+ (BY 1990) |  |  |  |  |
| n |  |  | 0 | 0 |
| Trip 2, August 1-2 |  |  |  |  |
|  | Total Catch: | 4 | Total Index Catch: | : 428 |
|  | Total Sampled: | 4 | Total Sampled: | 351 |
| Age 0+ (BY 1991) |  |  |  |  |
| Length |  | 45.0 | 33.4 | 35.9 |
| Weight |  | 1.28 | 0.21 | 0.31 |
| n | 0 | 4 | 9 | 114 |
| \% | 0 | 100 | 7.3 | 92.7 |
| Age 1+ (BY 1990) |  |  |  |  |
| п | 0 | 0 | 0 | 0 |
| Trip 3, August 21 |  |  |  |  |
|  | No trawls cond | ted | Total Index Catch: | : 225 |
|  |  |  | Total Sampled: | *100 |
| Age O+ (BY 1991) |  |  |  |  |
| Length |  |  | 48.5 | 50.2 |
| Weight |  |  | 0.96 | 1.12 |
| n |  |  | 2 | 89 |
| \% |  |  | 2.2 | 96.7 |
| Age 1+ (BY 1990) |  |  |  |  |
| Length |  |  |  | 88.0 |
| Weight |  |  |  | 6.31 |
| n |  |  | 0 | 1 |
| ? |  |  | 0 | 100 |
| Trip 4, September 27 |  |  |  |  |
|  | Total Catch: | 225 | Total Index Catch: | : 22 |
|  | Total Sampled: | 53 | Total Sampled: | 36 |
| Age Ot (BY 1991) |  |  |  |  |
| Length | 48.3 | 50.9 | 30.0 | 35.3 |
| Weight | 0.77 | 1.03 | 0.16 0, | 0.31 |
| n | 4 | 49 | 1 | 32 |
| \% | 7.6 | 92.5 | 3.097 | 97.0 |
| Age 1+ (BY 1990) |  |  |  |  |
| n | 0 | 0 | 0 | 0 |

[^11]Table 9. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Tatsamenie Lake during the summers and falls of 1993 and 1994. No juvenile sampling (beachseining or trawling) was conducted on Trip 1 in either year. ${ }^{\text {a }}$


Table 10. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Trapper Lake during the summer and fall of 1992. ${ }^{\text {a }}$

|  | Trawl |  | Beachseine |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Enhanced | wild | Enhanced Wil |  |
| Trip 1, June 20 |  |  |  |  |
|  | No trawls cond | ted | Total Index Catch: | 19 |
|  |  |  | Total Sampled: | 296 |
| Age $0+(B Y$ 1991) |  |  |  |  |
| Length |  |  | 29.2 |  |
| Weight |  |  | 0.13 |  |
| п |  |  | 75 | 0 |
| \% |  |  | 100 | 0 |
| Age 1+ (BY 1990) |  |  |  |  |
| п |  |  | 0 | 0 |
| Trip 2, August 4 |  |  |  |  |
|  | No trawls condu | ted | Total Index Catch: | 582 |
|  | (equipment fai | re) | Total Sampled: | 327 |
| Age $0+$ (BY 1991) |  |  |  |  |
| Length |  |  | 39.6 |  |
| Weight |  |  | 0.44 |  |
| n |  |  | 89 | 0 |
| \% |  |  | 100 | 0 |
| Age 1+ (BY 1990) |  |  |  |  |
| n |  |  | 0 | 0 |
| Trip 3, August 23 |  |  |  |  |
|  | Total Catch: n | trawls | Total Index Catch: | 314 |
|  |  |  | Total Sampled: | 257 |
| Age O+ (BY 1991) |  |  |  |  |
| Length |  |  | 44.4 |  |
| Weight |  |  | 0.68 |  |
| n |  |  | 100 | 0 |
| \% |  |  | 100 | 0 |
| Age 1+ (BY 1990) |  |  |  |  |
| n |  |  | 0 | 0 |
| Trip 4, Sept. 21-23 |  |  |  |  |
|  | Total Catch: | 49 | Total Index Catch: | 489 |
|  | Total Sampled: | 49 | Total Sampled: | 241 |
| Age 0+ (BY 1991) |  |  |  |  |
| Length | 52.4 | 41.0 | 53.6 |  |
| Weight | 1.03 | 0.51 | 1.13 |  |
| n | 14 | 14 | 100 | 0 |
| \% | 28.6 | 28.6 | 100 | 0 |
| Age 1t (BY 1990) |  |  |  |  |
| Length | 90.3 | 80.2 |  |  |
| Weight | 7.31 | 5.08 |  |  |
| n | 8 | 13 | 0 | 0 |
| \% | 16.3 | 26.5 | 0 | 0 |

Table 11. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Trapper Lake during the summers and falls of 1993 and 1994. There was no juvenile sampling (beachseining or trawling) conducted on Trip 1 in either year. ${ }^{\text {a }}$


Table 12. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in a sample of sockeye salmon juveniles from the single survey of Little Trapper Lake in the summer of 1992 . ${ }^{\text {a }}$

${ }^{\text {a }}$ Measurements are from specimens preserved in denatured ( $94 \%$ ) ethanol and are not directly comparable to fresh measurements. Total beachseine catch is for index sites only and does not include supplemental catches (see Methods, section 3.4). A sub-sample was selected from total beachseine samples proportional to relative abundance at the capture sites. Total sampled refers to no. of juveniles sampled for thermal marks.

Table 13. Numbers captured, mean lengths and weights, and percentages of enhanced and wild fish in samples of sockeye salmon juveniles from surveys of Little Trapper Lake during the summers and falls of 1993 and 1994. No juvenile sampling (beachseining or trawling) was conducted on Trip 1 in either year. ${ }^{\text {a }}$


[^12]Table 14. Hydroacoustic estimates of enhanced and wild juvenile sockeye in transboundary lakes, 1992 through 1994. Total estimates are apportioned to various categories of juveniles based on trawl capture data from Tables 1 through 13. Confidence limits of $95 \%$ are given for total estimates.

| Date of survey | Estimated numbers of juvenile sockeye |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | age 1+ |  | age 1+ |  |
|  |  | enhanced | wild | enhanced | wild |
| Tahltan Lake 1992 |  |  |  |  |  |
| July 29 | no estimate due to technical problems |  |  |  |  |
| Oct. 03 | no estimate due to equipment failure |  |  |  |  |
| 1993 |  |  |  |  |  |
| Sept. 18 | $\begin{gathered} 817,429 \\ +/-158,828 \end{gathered}$ | $294,274$ | 417,706 | 0 | 105,448 |
| 1994 |  |  |  |  |  |
| Sept. 18 | $\begin{gathered} 436,634 \\ +/-150,718 \end{gathered}$ | 25,761 | 410,873 | 0 | 0 |
| Tuya Lake 1992 |  |  |  |  |  |
| July 25 | $\begin{gathered} 147,322 \\ +/-128,268 \end{gathered}$ | 147,322 | $n / \mathrm{a}$ | n/a | n/a |
| Sept. 17 | $\begin{gathered} 596,537^{\mathrm{a}} \\ +/-196,156 \end{gathered}$ | 596,537 | $\mathrm{n} / \mathrm{a}$ | n/a | n/a |
| 1993 |  |  |  |  |  |
| Aug. 30 | $\begin{gathered} 437,304 \\ +/-228,578 \end{gathered}$ | 437,304 | n/a | 0 | n/a |
| 1994 |  |  |  |  |  |
| Sept. 02 | $\begin{gathered} 1,935,265 \\ +/-1,080,984 \end{gathered}$ | 1,935,265 | n/a | 0 | n/a |
| Tatsamenie Lake 1992 |  |  |  |  |  |
| Aug. 02 | $\begin{array}{r} 1,795,965 \\ +/-772,015 \end{array}$ | 0 | 1,795,965 ${ }^{\text {b }}$ | 0 | 0 |
| Sept. 28 | no estimate due to equipment failure |  |  |  |  |

Table 14 cont'd.

| Sept. 14 | 1993 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{array}{r} 1,146,054 \\ +/-409,859 \end{array}$ | 142,111 | 975,292 | 0 | 28,651 |
| 1994 |  |  |  |  |  |
| Sept. 13 | $\begin{array}{r} 1,053,185 \\ +/-358,658 \end{array}$ | 18,957 | 1,034,228 | 0 | 0 |
| Trapper Lake 1992 |  |  |  |  |  |
| Aug. 03 | $\begin{gathered} 196,037 \\ +-55,203 \end{gathered}$ | type estimates not available as no trawls were made due to winch failure |  |  |  |
| Sept. 30 | no estimate due to equipment failure |  |  |  |  |
| 1993 |  |  |  |  |  |
| Sept. 10 | $\begin{gathered} 125,459 \\ +/-64,774 \end{gathered}$ | 94,847 | 15,306 | 12,295 | 3,011 |
| 1994 |  |  |  |  |  |
| Sept. 11 | $\begin{gathered} 64,554 \\ +-25,446 \end{gathered}$ | 64,554 ${ }^{\text {c }}$ | 0 | 0 | 0 |
| Little Trapper Lake 1992 |  |  |  |  |  |
| not surveyed in 1992 |  |  |  |  |  |
| 1993 |  |  |  |  |  |
| Sept. 07 | $\begin{gathered} 296,890^{\mathrm{d}} \\ +/-116,354 \end{gathered}$ | 1,993 | 292,905 | 0 | 1,993 |
| 1994 |  |  |  |  |  |
| Sept. 08 | $\begin{gathered} 554,748^{\mathrm{d}} \\ +/-311,232 \end{gathered}$ | 0 | 543,653 | 0 | 11,095 |

## Nakina Lake

1993 and 1994 no juvenile sockeye detected
${ }^{\text {a }}$ reliability of estimate uncertain due to possible sounder malfunction
${ }^{\text {b }}$ based on exceptionally small trawl sample ( 4 fish)
${ }^{\text {c }}$ based on exceptionally small sample ( 3 fish)
${ }^{\text {d }}$ probable underestimate due to very high density

Table 15. Estimated survivals of outplanted sockeye fry based on fall hydroacoustic/trawl surveys. Reliability of these estimates is discussed in the text.

| Broodyear <br> (BY) | Year of outplanting (BY+1) | Number outplanted | Survival to fall, BY+1 (age 0+) |  | Survival to fall, BY+2 (age $1+)^{2}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Estimated fall fry | \% | Estimated fall fry | \% |
|  |  |  | Tabltan Lake |  |  |  |
| 1989 | 1990 | 1,042,000 | no est. | n/a | 6,224 | 0.1 |
| 1990 | 1991 | 3,585,000 | 479,286 | 13.4 | no est. | n/a |
| 1991 | 1992 | 1,415,000 | no est. | n/a | 0 | 0.0 |
| 1992 | 1993 | 1,947,000 | 294,274 | 15.1 | 0 | 0.0 |
| 1993 | 1994 | 904,000 | 25,761 | 2.8 | n/a | n/a |
| Average |  |  |  | 10.4 |  | . 03 |
|  |  |  | Tuya Lake |  |  |  |
| 1991 | 1992 | 1,632,000 | 596,537 | 36.6 | 0 | 0.0 |
| 1992 | 1993 | 1,990,000 | 437,304 | 22.0 | 0 | 0.0 |
| 1993 | 1994 | 4,691,000 | 1,935,265 | 41.3 | $\mathrm{n} / \mathrm{a}$ | n/a |
| Average |  |  |  | 33.3 |  | 0.0 |
|  |  |  | Tatsamenic Lake |  |  |  |
| 1990 | 1991 | 673,000 | 201,563 | 29.9 | no est. | n/a |
| 1991 | 1992 | 1,232,000 | no est. | n/a | 0 | 0.0 |
| 1992 | 1993 | 909,000 | 142,111 | 15.6 | 0 | 0.0 |
| 1993 | 1994 | 521,000 | 18,957 | 3.6 | n/a | n/a |
| Average |  |  |  | 16.4 |  | 0.0 |
| Trapper Lake |  |  |  |  |  |  |
| 1990 | 1991 | 934,000 | 231,689 | 24.8 | no est. | n/a |
| 1991 | 1992 | 1,811,000 | no est. | $\mathrm{n} / \mathrm{a}$ | 12,295 | 0.1 |
| 1992 | 1993 | 1,113,000 | 94,847 | 8.5 | 0 | 0.0 |
| 1993 | 1994 | 916,000 | $0^{\text {b }}$ | 0.0 | $\mathrm{n} / \mathrm{a}$ | n/a |
| Average |  |  |  | $11.1{ }^{\text {c }}$ |  | 0.1 |

[^13]Table 16. Estimations of total emigration, percent by age class, mean lengths and weights of Tahltan Lake smolts; 1991 through 1995. ${ }^{\text {a }}$

|  |  | BY | Percent | Estimated number | length (mm) | weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Wild |  |  |  | 1991 |  |  |
|  | 1 | 1989 | 74.36 | 1,105,882 | 90.6 | 5.82 |
|  | 2 | 1988 | 7.48 | 111,243 | 112 | 11.80 |
|  | 3 | 1987 | 0.02 | 3,272 | 153 | 27.3 |
| Enhanced | 1 | 1989 | 17.94 | 266,868 | 88.6 | 5.40 |
|  | 2 | 1988 | n/a | (outplants began in 1990, BY 1989) |  |  |
| Wild |  |  |  | 1992 |  |  |
|  | 1 | 1990 | 43.81 | 681,310 | 84.8 | 4.77 |
|  | 2 | 1989 | 4.26 | 66,238 | 110 | 10.20 |
| Enhanced | 3 | 1988 | 0.02 | 3,154 | 177 | 45.80 |
|  | 1 | 1990 | 49.70 | 772,782 | 84.3 | 4.63 |
|  | 2 | 1989 | 2.03 | 31,542 | 115 | 12.00 |
|  | 3 | 1988 | $n / \mathrm{a}$ |  |  |  |
| 1993 |  |  |  |  |  |  |
| Wild | 1 | 1991 | 86.01 | 2,799,607 | 80.1 | 3.94 |
|  | 2 | 1990 | 1.72 | 55,955 | 105.3 | 10.03 |
| Enhanced | 1 | 1991 | 11.36 | 369,892 | 79.6 | 3.85 |
|  | 2 | 1990 | 0.91 | 29,591 | 116.5 | 12.85 |
| 1994 |  |  |  |  |  |  |
| Wild | 1 | 1992 | 59.30 | 542,633 | 84.3 | 5.00 |
|  | 2 | 1991 | 8.54 | 78,176 | 101.8 | 8.41 |
| Enhanced | 1 | 1992 | 32.16 | 294,310 | 83.4 | 4.74 |
|  | 2 | 1991 | 0 | - | - | - |
| 1995 |  |  |  |  |  |  |
| Wild | 1 | 1993 | 90.44 | 743,674 | 83.4 | 4.71 |
|  | 2 | 1992 | 2.84 | 23,353 | 116.7 | 13.45 |
| Enhanced | 1 | 1993 | 5.43 | 44,650 | 81.7 | 4.37 |
|  | 2 | 1992 | 1.29 | 10,607 | 113.0 | 11.98 |

${ }^{\text {a }}$ Measurements are from fresh (unpreserved) fish.

Table 17. Estimated percentages by age class, total emigration, and mean lengths and weights for smolts observed at Tuya Lake in 1993, 1994, and 1995. ${ }^{\text {a }}$

| Origin | Age | BY | Percent | Estimated number | length (mm) | weight <br> (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 1993 |  |  |
| Wild | There are no wild sockeye in Tuya Lake |  |  |  |  |  |
| Enhanced | 1 | 1991 | 100 | no estimate | 99.7 | 8.76 |
|  | 2 | 1990 | n/a | (the first outplant was in 1992, BY 1991) |  |  |
|  |  |  |  | 1994 |  |  |
| Wild | There are no wild sockeye in Tuya Lake |  |  |  |  |  |
| Enhanced | 1 | 1992 | 96.00 | no estimate | 99.0 | 8.99 |
|  | 2 | 1991 | 4.00 | no estimate | 135.3 | 22.34 |
|  |  |  |  | 1995 |  |  |
| Wild | There are no wild sockeye in Tuya Lake |  |  |  |  |  |
| Enhanced | 1 | 1993 | 97.07 | no estimate | 95.58 | 9.64 |
|  | 2 | 1992 | 2.93 | no estimate | 137.0 | 27.35 |

[^14]Table 18. Estimated percentage by age class, total emigration, numbers, and mean lengths and weights for smolts observed at Tatsamenie Lake; 1992 through $1995{ }^{\text {a }}$


[^15]Table 19. Estimated percentage by age class, total emigration, and mean lengths and weights for smolts observed at Trapper Lake; 1992 through 1995. ${ }^{\text {a }}$

| Origin | Age | BY | No. in <br> Sample | Per- <br> cent | Total emigration | Mean length <br> $(\mathrm{mm})$ | Mean weight <br> $(\mathrm{g})$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

There were no smolts captured during trapping conducted in 1992


[^16]Table 20. Estimated percentages of total emigration, numbers, and mean lengths and weights for smolts observed at Little Trapper Lake; 1992 through 1995. ${ }^{\text {a }}$

| Origin | Age | BY | Percent | Estimated number | $\begin{aligned} & \text { Mean length } \\ & (\mathrm{mm}) \end{aligned}$ | Mean weight <br> (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 |  |  |  |  |  |  |
| Wild | 1 | 1990 | 70.75 | no estimate | 69.8 | 2.90 |
|  | 2 | 1989 | 22.13 | no estimate | 95.3 | 7.75 |
| Enhanced | 1 | 1990 | 7.11 | no estimate | 69.9 | 2.93 |
|  | 2 | 1989 | n/a (there were no outplants until 1991, BY 1990) |  |  |  |
| 1993 |  |  |  |  |  |  |
| Wild | 1 | 1991 | 71.0 | no estimate | 59.0 | 2.09 |
|  | 2 | 1990 | 25.0 | no estimate | 82.3 | 4.98 |
| Enhanced | 1 | 1991 | 1.0 | no estimate | 54.0 | 1.3 |
|  | 2 | 1990 | 3.0 | no estimate | 86.0 | 5.37 |
| 1994 |  |  |  |  |  |  |
| Wild | 1 | 1992 | 88.81 | no estimate | 59.9 | 1.73 |
|  | 2 | 1991 | 8.39 | no estimate | 81.3 | 4.03 |
| Enhanced | 1 | 1992 | 2.10 | no estimate | 54.7 | 1.43 |
|  | 2 | 1991 | 0.70 | no estimate | 70.0 | 2.5 |
| 1995 |  |  |  |  |  |  |
| Wild | 1 | 1993 | 64.08 | no estimate | 66.5 | 2.78 |
|  | 2 | 1992 | 33.06 | no estimate | 80.9 | 4.80 |
| Enhanced | 1 | 1993 | 0.41 | no estimate | 69.0 | 3.20 |
|  | 2 | 1992 | 2.45 | no estimate | 110.7 | 13.42 |

[^17]Table 21. Survival of sockeye fry outplanted to Tahltan Lake, brood-years 1989 to 1993; from outplanted fry to smolt.

| Abundance |  |  |  |  | Survival from planted fry |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Juveniles ${ }^{\text {a }}$ |  | Smolts |  | Juveniles ${ }^{\text {a }}$ |  | Smolts |  |  |
| Fry planted (year) | fall, age $0+$ (year) | $\begin{gathered} \text { fall, } \\ \text { age } 1+ \\ \text { (year) } \end{gathered}$ | age $1+$ smolts (year) | age $2+$ smolts (year) | to fall age 0+ | $\text { to fall }{ }^{\mathrm{b}} \text { age } 1+$ | to age 1+ | $\begin{array}{r} \text { to age } \\ 2+ \end{array}$ | $\begin{aligned} \text { total } \begin{array}{c} \text { (age } \\ 1+2) \end{array} \end{aligned}$ |
| Brood-year 1989 |  |  |  |  |  |  |  |  |  |
| $\begin{array}{r} 1,042,000 \\ (1990) \end{array}$ | no est. (1990) | $\begin{array}{r} 6,224 \\ (1991) \end{array}$ | $\begin{array}{r} 266,868 \\ (1991) \end{array}$ | $\begin{aligned} & 31,542 \\ & (1992) \end{aligned}$ | $n / \mathrm{a}$ | 0.1 | 25.6\% | 3.03\% | 28.6\% |
| Brood-year 1990 |  |  |  |  |  |  |  |  |  |
| $\begin{array}{r} 3,585,000 \\ (1991) \end{array}$ | $\begin{array}{r} 479,286 \\ (1991) \end{array}$ | $\begin{aligned} & \text { no est. } \\ & \text { (1992) } \end{aligned}$ | $\begin{array}{r} 772,782 \\ (1992) \end{array}$ | $\begin{aligned} & 29,591 \\ & (1993) \end{aligned}$ | 13.4\% | $n / 2$ | 21.6\% | 0.82\% | 22.4\% |
| Brood-year 1991 |  |  |  |  |  |  |  |  |  |
| $\begin{array}{r} 1,415,000 \\ (1992) \end{array}$ | $\begin{aligned} & \text { no est. } \\ & \text { (1992) } \end{aligned}$ | $\begin{array}{r} 0 \\ (1993) \end{array}$ | $\begin{array}{r} 369,892 \\ (1993) \end{array}$ | $\begin{array}{r} 0 \\ (1994) \end{array}$ | $n / \mathrm{a}$ | 0.0\% | 26.2\% | 0.00\% | 26.2\% |
| Brood-year 1992 |  |  |  |  |  |  |  |  |  |
| $\begin{array}{r} 1,947,000 \\ (1993) \end{array}$ | $\begin{array}{r} 294,274 \\ (1993) \end{array}$ | $\begin{array}{r} 0 \\ (1994) \end{array}$ | $\begin{array}{r} 294,310 \\ (1994) \end{array}$ | $\begin{aligned} & 10,607 \\ & (1995) \end{aligned}$ | 15.1\% | 0.0\% | 15.1\% | 0.54\% | 15.7\% |
| Brood-year 1993 |  |  |  |  |  |  |  |  |  |
| $\begin{array}{r} 904,000 \\ (1994) \end{array}$ | $\begin{aligned} & 25,761 \\ & (1994) \end{aligned}$ | $\begin{aligned} & \text { no est. }{ }^{\text {c }} \\ & (1995) \end{aligned}$ | $\begin{aligned} & 44,650 \\ & (1995) \end{aligned}$ | $\begin{aligned} & 10,607 \\ & (1996) \end{aligned}$ | 2.8\% | n/a | 4.9\% | n/a | n/a |
| Average survivals |  |  |  |  | 10.4\% | 0.03\% | 18.7\% | 1.1\% | $23.2 \%^{\text {d }}$ |

${ }^{2}$ see comments on accuracy of hydroacoustic estimates in text.
${ }^{\mathrm{b}}$ this age class likely under-represented in samples because of greater net avoidance associated with larger size.
${ }^{c}$ this data collected after period covered by this report, analysis not yet done.
${ }^{d}$ average for brood years 1989-1992.

Table 22. Summary of sockeye salmon smolt size in Tahltan Lake relative to spring juvenile abundance in first year of lake rearing. Years when lake enrichment (fertilization) took place are indicated.

| Brood- <br> year <br> (BY) <br> t | Estimated total juvenile abundance ( 1000 's) in spring of year $t+1^{c}$ |  |  | Mean weight (g) of smolts originating from BY t |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Adult Weir Counts | wild |  | enhanced |  |  |  |
|  | age $0+$ (BY t) | $\begin{aligned} & \text { age } 1+ \\ & \text { (BY } \mathrm{t}-1 \text { ) } \end{aligned}$ | total |  | age 1+ | age $2+$ | age 1+ | age $2+$ | year t+1 | year t+2 |
| 1982 | 1195 | 19 | 1214 | 28,257 | 4.81 | 9.31 |  |  | no | no |
| 1983 | 1786 | 154 | 1940 | 21,256 | 3.75 | 8.45 |  |  | no | по |
| 1984 | 1350 | 35 | 1385 | 32,777 | 4.71 | 10.77 |  |  | no | yes |
| 1985 | 2831 | 184 | 3015 | 67,326 | 6.34 | 12.11 |  |  | yes | yes |
| 1986 | 3791 | 165 | 3956 | 20,280 | 5.75 | 15.85 |  |  | yes | no |
| 1987 | 2191 | 81 | 2272 | 6,958 | 6.02 | 8.97 |  |  | no | no |
| 1988 | 2243 | 133 | 2376 | 2,536 | 5.93 | 11.68 |  |  | no | no |
| 1989 | 5048 | 161 | 5209 | 8,316 | 5.82 | 10.17 | 5.40 | 12.0 | no | no |
| 1990 | 6256 | 142 | 6398 | 14,927 | 4.77 | 10.03 | 4.63 | 12.85 | no | no |
| 1991 | 12398 | 213 | 12611 | 50,135 | 3.94 | 8.41 | 3.85 | $-^{\text {a }}$ | no | no |
| 1992 | 4027 | 113 | 4140 | 59,907 | 5.00 | 13.45 | 4.74 | 11.98 | no | no |
| 1993 | 3755 | 49 | 3804 | 53,362 | 4.71 | $7.2^{\text {b }}$ | 4.37 | $8.7{ }^{\text {b }}$ | no | no |

${ }^{\text {a }}$ none observed in sample
${ }^{\mathrm{b}}$ this age class emigrated in 1996
${ }^{c}$ estimates of spring juvenile abundance were calculated by extrapolating backwards using actual smolt counts beginning in 1984 and fry mortality rates that were presented in a previous report (TTC, 1994; Appendix 3).

Table 23. Percentages of enhanced fry from Trapper Lake outplants in samples of age $0+$ fry collected the same year in Little Trapper Lake. Standardized percentages are observed percentages adjusted for differences in the numbers outplanted and in the numbers of wild spawners in Little Trapper Lake (see Results Section 3.7.4 for details).

|  | Little Trapper sample details | Observed Percent enbanced | Standardized Percent enbanced |
| :---: | :---: | :---: | :---: |
| $\begin{gathered} 1991 \\ (1990) \end{gathered}$ | July 10 beachseine | 12.0 | 12.0 |
| $\begin{gathered} 1992 \\ (1991) \end{gathered}$ | July 14 beachseine | 18.0 | 27.4 |
| $\begin{gathered} 1993 \\ (1992) \end{gathered}$ | July 28 beachseine | 0 | 0 |
| " | Sept. 7 beachseine | 0 | 0 |
| " | Sept. 7 Irawl | 0.7 | 0.9 |
| $\begin{gathered} 1994 \\ (1993) \end{gathered}$ | July 23 beachseine | 0.7 | 1.5 |
| " | Sept. 9 beachseine | $3.1{ }^{\text {a }}$ | 6.8 |
| " | Sept. 9 trawl | 0 | 0 |

[^18]Table 24. Zooplankton species composition (\% of total biomass) of Trapper and Little Trapper Lakes, 1991 through 1994.

| Lake, year | Season | Bosminids | Daphnids | Cyclopoids | Nauplii and rotifers |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Trapper, 1991 | summer | 0 | 0 | 100 | $<0.1$ |
|  | fall | 1.5 | 0 | 98.5 | $<0.1$ |
| Trapper, 1992 | summer | 0.2 | 0 | 99.8 | $<0.1$ |
|  | fall | 0.1 | 0 | 99.9 | $<0.1$ |
| Trapper, 1993 | summer | 0.4 | 0 | 99.6 | $<0.1$ |
|  | fall | 0.7 | 0 | 99.3 | $<0.1$ |
| Trapper, 1994 | summer | 0 | 0.8 | 98.4 | 0.8 |
|  | fall | 1.2 | 2.4 | 92.2 | 4.2 |
| Little Trapper, 1991 | summer | 2.1 | 0 | 83.0 | 14.9 |
|  | fall | 69.2 | 0.1 | 17.4 | 13.3 |
| Little Trapper, 1993 | summer | 10.4 | 0.2 | 81.0 | 8.4 |
|  | fall | 86.0 | 0.3 | 3.3 | 10.4 |
| Little Trapper, 1994 | summer | 12.2 | 2.0 | 83.9 | 1.9 |
|  | fall | 33.2 | 5.1 | 20.9 | 40.8 |

Table 25. Revised estimates of Euphotic Volume (EV), adult production capacities, and spring fry capacities for all Transboundary lakes studied. Based on all available Secchi depth data.

| Lake <br> (number <br> years observed) | Average Secchi depth | $\begin{gathered} \text { Lake }^{2} \\ \text { type } \end{gathered}$ | Euphotic ${ }^{\text {b }}$ depth estimate (m) | Surface area (hectares) | $\begin{aligned} & \hline \mathrm{EV} \\ & \text { units } \end{aligned}$ | Previous ${ }^{\text {d }}$ EV unit estimate | Adult ${ }^{e}$ production estimate ( K 's) | previous adult product. estimate | estimated ${ }^{\text {f }}$ spring fry capacity (K's) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Tahltan (10) | 11.23 | C | 19.21 | 492 | 94.53 | 81 | 201 | 176 | 10,040 |
| Tuya (9) | 5.29 | C/s | 9.07 | 3,127 | 283.50 | 407 | 543 | 764 | 27,142 |
| Tatsamenie (10) | 5.23 | C/G | 10.03 | 1,679 | 168.41 | 202 | 335 | 394 | 16,726 |
| Little <br> Tatsamenie (3) | 4.8 | C/G | 9.39 | 76 | 7.14 | n/a | 43 | n/a | 2,131 |
| Trapper (8) | 1.7 | G | 6.79 | 557 | 37.81 | 35 | 98 | 93 | 4,907 |
| Little Trapper (7) | 2.34 | G | 8.90 | 199 | 17.71 | 26 | 62 |  |  |
| Nakina (2) | 5.6 | cs | 9.41 | 491 | 46.21 | n/a | 113 | n/a | 5,667 |
| Chutine (2) | 1.05 | G | 4.64 | 615 | 28.56 | 12 | 81 | 51 | 4,070 |
| Christina (2) | 0.55 | G | 2.99 | 141 | 4.22 | 8 | 37 | 44 | 1,867 |
| Kennicott (2) | 1.65 | C/s | 5.01 | 128 | 6.41 | 5 | 41 | 39 | 2,066 |
| King Salmon (3) | 10.08 | C | 17.79 | 227 | 40.37 | 28 | 103 | 80 | 5139 |
| Kuthai (1) | 9.9 | C | 17.56 | 157 | 27.57 | 39 | 80 | 100 | 3,980 |
| Klukshu (3) | 2.83 | C/G | 6.45 | 135 | 8.71 | 12 | 45 | 51 | 2,273 |

[^19]Table 26. Transboundary Lakes sockeye brood stock disease histories, BY's 1988-1995. Results are discussed in Appendix 2.

Tahltan

| $\begin{gathered} \text { BROOD } \\ \text { YEAR } \end{gathered}$ | BKD |  | $\mathrm{IHNV}^{\text {a }}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sample <br> Size | Percent <br> Positive | IHNV Positive |  | Positives greater or equal to $10^{4} \mathrm{pfu}$ |  |
|  |  |  | Sample <br> Size | Percent |  |  |
|  |  |  |  |  | Number | Percent |
| 1988 | 19/60 | 31.7\% | 54/60 | 90.0\% | 28/54 | 51.9\% |
| 1989 | 7/151 | 4.6\% | 3/159 | 1.9\% | 1/3 | 33.3\% |
| 1990 | $9 / 150$ | 6.0\% | 5/150 | 3.3\% | $0 / 5$ | 0.0\% |
| 1991 | 11/148 | 7.4\% | 144/152 | 94.7\% | 65/144 | 45.1\% |
| 1992 | 9/154 | 5.8\% | 141/154 | 91.6\% | 82/141 | 58.2\% |
| 1993 | 11/150 | 7.3\% | 107/149 | 71.8\% | 45/107 | 42.1\% |
| 1994 | 4/150 | 2.7\% | 75/150 | 50.0\% | 21/75 | 28.0\% |
| 1995 | $7 / 150$ | 4.7\% | 93/150 | 62.0\% | 45/93 | 48.4\% |

Tatsamenle

| BROODYEAR | BKD |  | 1HNV |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sample <br> Size | Percent <br> Positive | IHNV Positive |  | Positives greater or equal to $10^{4} \mathrm{pfu}$ |  |
|  |  |  | Sample <br> Size | Percent |  |  |
|  |  |  |  |  | Number | Percent |
| 1988 | 3/67 | 4.5\% | 25/65 | 38.5\% | 4/25 | 16.0\% |
| 1989 |  |  |  | o egg tak |  |  |
| 1990 | 12/150 | 8.0\% | 96/150 | 64.0\% | 50/96 | 52.1\% |
| 1991 | 9/150 | 6.0\% | 5/150 | 3.3\% | $0 / 5$ | 0.0\% |
| 1992 | 5/151 | 3.3\% | 95/150 | 63.3\% | $49 / 95$ | 51.6\% |
| 1993 | 24/111 | 21.6\% | 94/149 | 63.1\% | 57/94 | 60.6\% |
| 1994 | 10/150 | 6.7\% | 1/103 | 1.0\% | $0 / 1$ | 0.0\% |
| 1995 | 15/150 | 10.0\% | 1/149 | 0.7\% | 1/1 | 100.0\% |

Little Trapper

| $\begin{gathered} \text { BROOD } \\ \text { YEAR } \end{gathered}$ | BKD |  | IHNV |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \text { Sample } \\ & \text { Size } \end{aligned}$ | Percent <br> Positive | IHNV Positive |  | Positives greater or equal to $10^{4} \mathrm{pfu}$ |  |
|  |  |  | Sample <br> Size | Percent |  |  |
|  |  |  |  |  | Number | Percent |
| 1988 | 2/60 | 3.3\% | 52/60 | 86.7\% | 23/52 | 44.2\% |
| 1989 |  |  |  | no egg take |  |  |
| 1990 | 20/150 | 13.3\% | 146/152 | 96.1\% | 113/1461 | 77.4\% |
| 1991 | 9/150 | 6.0\% | 20/150 | 13.3\% | 5/20 | 25.0\% |
| 1992 | 1/153 | 0.7\% | 146/150 | 97.3\% | 126/146 | 86.3\% |
| 1993 | 10/150 | 6.7\% | 90/150 | 60.0\% | 47/90 | 52.2\% |
| 1994 | 10/150 | 6.7\% | 50/148 | 33.8\% | 16/50 | 32.0\% |
| 1995 |  |  |  | no egg take |  |  |

${ }^{\text {a }}$ For 1 HNV , a titer $\geq 10^{4}$ plague forming units (pfu), is the point at which the probability of vertical (parent to offspring) transmission of IHNV is felt to greatly increase.

APPENDICES

Appendix 1. Summary of thermal marks applied at Snettisham Hatchery to transboundary river sockeye salmon, Broodyears 1989 to 1995.

| Stock | Release Site | Brood year | Mark ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: |
| Tahltan | Tahltan Lake | 89 | 4 |
|  |  | 90 | 3 |
|  |  | 91 | 4 |
|  |  | 92 | 7 |
|  |  | 93 | $6+5$ |
|  |  | 94 | 6 |
|  |  | 95 | 6 |
| Tahltan | Tuya Lake | 91 | 6 |
|  |  | 92 | 5 |
|  |  | 93 | 4+5 |
|  |  | 94 | 4 |
|  |  | 95 | 4+4 |
| Little Trapper | Trapper Lake | 90 | 5 |
|  |  | 91 | 6 |
|  |  | 92 | 7+3 |
|  |  | 93 | 4+5 |
|  |  | 94 | 7 |
| Litle Tatsamenie | Tatsamenie Lake | 90 | 3 |
|  |  | 91 | 4 |
|  |  | 92 | 4+3 |
|  |  | 93 | 5+5 |
|  |  | 94 | 5 |
|  |  | 95 | 5 |

${ }^{2}$ Each mark is comprised of a single band containing the listed number of rings. Where 2 bands of rings are denoted (i.e. $6+5$ ), the $1^{\text {st }}$ band is applied before the sockeye embryo hatches, and the $2^{\text {nd }}$ band is applied after hatching.

## Appendix 2. Brood stock disease screening results for Tahltan, Little Trapper, and Tatsamenie lakes. Data summary is located in Table 26.

note: For IHNV, a titer $\geq 10^{4}$ plague forming units (pfu) is the point at which the likelihood of vertical (parent to offspring) transmission of IHNV is felt to greatly increase.

## Tahtan Lake

## 1990 Brood year

Disease testing of brood stock showed 9/150 fish positive for BKD (6.0\%). Five out of 150 (3.3\%) tested pasitive for IHNV; however, none of these had titers $>10^{4}$ pfu. The incidences of both diseases were very comparable to those seen in 1989 (TTC 1991) and considerably lower than those seen in natural spawners in the preliminary survey year in 1988 (TTC, 1989). This was likely attributable to the 1990 samples having been captured sooner after entering the spawning grounds, thereby reducing stress and the likelihood of horizontal (fish to fish) transmission. Risk of vertical transmission was considered to be very low.

## 1991 Brood year

Disease testing of brood stock showed 11/148 fish positive for BKD (7.4\%). Of 152 fish tested, 144 ( $94.7 \%$ ) tested positive for IHNV; 65 of these had titers $>10^{4}$ pfu ( $42.8 \%$ of the total sample). The high incidence and high titers of IHNV were comparable to those seen in 1988. It was felt the high incidence of IHNV was due to the cyclical nature of the virus and/or a result of factors favoring horizontal transmission. The latter appears quite probable since the high escapement in 1991 resulted in large numbers of fish congregating near and competing for the limited spawning grounds, where brood stock is captured and held. It is unlikely the higher incidence was due to differences in methods of capture and/or holding of brood stock since these were very similar to those of previous years.

## 1992 Brood year

Disease testing of brood stock showed $9 / 154$ fish positive for BKD (5.8\%). Of 154 fish tested, $141(91.6 \%)$ tested positive for IHNV; 82 of these had titers $\geq 10^{4}$ pfu ( $58.2 \%$ of the positive fish). The risk of vertical transmission of IHNV this year may again be high.

## 1993 Brood year

Disease testing of brood stock showed 11/150 fish positive for BKD (7.3\%). Of 149 fish tested, 107 (71.8\%) tested positive for IHNV; 45 of these had titers $\geq 10^{4} \mathrm{pfu}$ ( $42.1 \%$ of the positive fish). The risk of vertical transmission of IHNV this year could be considered moderate.

## 1994 Brood year

Disease testing of brood stock showed $4 / 150$ fish positive for BKD ( $2.7 \%$ ). Of 150 fish tested, 75 ( $50.0 \%$ ) tested positive for IHNV; 21 of these had titers $\geq 10^{4}$ pfu ( $28.0 \%$ of the positive fish). The risk of vertical transmission of $I H N V$ this year could be considered moderate.

## Little Tatsamenie Lake

## 1990 Brood year

Disease testing of brood stock showed $12 / 150$ fish positive for BKD (8.0\%). Incidence of IHNV was $96 / 150(64.0 \%), 50$ of these having titers $>10^{4} \mathrm{pfu}$ ( $33.3 \%$ of the total sample). The incidence of both diseases was higher than those seen in the 1988 preliminary survey ( $0 \% \mathrm{BKD} ; 38.5 \% \mathrm{IHNV}$ with $16 \%$ tier $>10^{4}$ ). Tatsamenie brood stock is quite green when captured and the higher disease incidences may have been attributable to the requirement to hold brood stock fish for several weeks to ripen.

## 1991 Brood year

Disease testing of brood stock showed $9 / 150$ fish positive for BKD (6.0\%). Incidence of IHNV was only $5 / 150(3.3 \%)$, with none of these having titers $>10^{4}$ pfu Incidence of IHNV was considerably lower than in 1990 , possibly a result of lower holding densities and reduced stress. Likelihood of vertical transmission was considered to be very low.

## 1992 Brood year

Disease testing of brood stock showed $5 / 151$ fish positive for BKD (3.3\%). Of 150 fish tested, 95 (63.3\%) tested positive for IHNV; 49 of these had titers $\geq 10^{4} \mathrm{pfu}$ ( $51.6 \%$ of the positive fish). The risk of vertical transmission of IHNV this year was considered moderate.

## 1993 Brood year

Disease testing of brood stock showed 24/111 fish positive for BKD (21.6\%). Of 149 fish tested, 94 ( $63.3 \%$ ) tested positive for IHNV; 57 of these had titers $\geq 10^{4} \mathrm{pfu}$ ( $60.6 \%$ of the positive fish). The risk of vertical transmission of IHNV this year was considered moderate.

## 1994 Brood year

Disease testing of brood stock showed 10/150 fish positive for BKD (6.7\%). Of 103 fish tested, 1 ( $1.0 \%$ ) tested positive for IINV; the titer was below $10^{4} \mathrm{pfu}$. The risk of vertical transmission of IHNV this year could be considered low.

## Little Trapper Lake

## 1990 Brood year

Disease testing of brood stock showed 20/150 fish positive for BKD (13.3\%). Incidence of IHNV was very high, 146/152 (96.1\%); 113 of these had titer's $\geq 10^{4}$ pfu ( $73.0 \%$ of the total sample). BKD level was much bigher than the $3.3 \%$ seen in the initial disease survey in 1988. Incidence of IHNV was somewhat higher than the 1988 incidence of $86.7 \% \mathrm{IHNV}$ with $44.2 \%$ having $\geq 10^{4}$ pfu. It was felt the higher disease incidences were likely due to horizontal transmission resulting from holding large numbers of broodstock under stressful conditions for long periods. The possibility of vertical transmission of IHNV was considered to be high.

## 1991 Brood year

Disease testing of brood stock showed $9 / 150$ fish positive for BKD ( $6.0 \%$ ). Incidence of IHNV was much lower than in 1990 , $20 / 150(13.3 \%) 5$ of these had titer's $\geq 10^{4} \mathrm{pfu}(3.3 \%$ of the total sample). The risk of vertical transmission of IHNV was considered to be minimal. The lower incidence of IHNV could have been attributable in part to the reduced need for sorting of fish due to the availability of more ripe fish from the larger escapement. However, capture and holding techniques employed in 1990 and 1991 were similar and it is also possible the lower incidence may have been attributable simply to an inherent annual variability in virus levels. If so, the similar reduced incidence of IHNV observed in the Little Tatsamenie brood stock in 1991 suggests annual fluctuations might be widespread rather than stock specific.

## 1992 Brood year

Disease testing of brood stock showed 1/153 fish positive for BKD ( $0.7 \%$ ). Of 150 fish tested for viruses, 146 ( $97.3 \%$ ) tested positive for IHNV; 126 of these had titers $\geq 10^{4} \mathrm{pfu}$ ( $86.3 \%$ of the positive fish). The increase in incidence of IHNV virus again suggests annual variability in virus levels. The risk of vertical transmission of IHNV this year was considered high.

## 1993 Brood year

Disease testing of brood stock showed $10 / 150$ fish positive for BKD (6.7\%). Of 150 fish tested, $90(60.0 \%)$ tested positive for IHNV; 47 of these had titers $\geq 10^{4}$ pfu ( $52.2 \%$ of the positive fish). The risk of vertical transmission of IHNV this year was considered high.

## 1994 Brood year

Disease testing of brood stock showed 10/150 fish positive for BKD (6.7\%). Of 148 fish tested, $50(33.8 \%)$ tested positive for IHNV; 16 of these had titers $\geq 10^{4}$ pfu ( $32.0 \%$ of the positive fish). The risk of vertical transmission of IHNV this year was considered moderate.

Appendix 3. Summary of fry stocking dates for brood years 1989 through 1994 at Tahltan, Tuya, Trapper, and Tatsamenie lakes.

| Brood <br> Year | Release Site | Stocking Dates |
| :--- | :--- | :--- |
|  |  |  |
| 1989 | Tahltan | June 6,13,14,20,25 |
| 1990 | Tahltan | June 2,3,4,5,7,12,13,18,19,20,21 |
| 1991 | Tahltan | June 9,10 |
| 1992 | Tahltan | June 23,26; July 2 |
| 1993 | Tahltan | June 24,28 |
| 1994 | Tahltan | June 26; July 3 |
|  |  |  |
| 1991 | Tuya | June 17,20,21 |
| 1992 | Tuya | June 16,25; July 7 |
| 1993 | Tuya | June 24,28,30; July 1,12,13 |
| 1994 | Tuya | June 21, 25; July 3 |
|  |  |  |
| 1990 | Trapper | June 5,8,20,22 |
| 1991 | Trapper | June 4,5,9,10,11 |
| 1992 | Trapper | June 25; July 2 |
| 1993 | Trapper | June 16,24 |
| 1994 | Trapper | June 21,28; July 3 |
|  |  |  |
| 1990 | Tatsamenie | June 22 |
| 1991 | Tatsamenie | June 22,24,26 |
| 1992 | Tatsamenie | July 9,14 |
| 1993 | Tatsamenie | July 14 |
| 1994 | Tatsamenie | July 18, 21 |
|  |  |  |

Appendix 4. Zooplankton biomass, density, and mean size for six transboundary lakes, 1987-1995.

## Glossary of Codes for

Zooplankton Data

| Code | Name |
| :---: | :--- |
| 20 | Epischura nevadensis |
| 60 | Harpacticoid |
| 100 | Bosmina longispina |
| 141 | Chydorid |
| 180 | Ceriodaphnia |
|  | quadrangula |
| 450 | Ostracod |
| 900 | Chironomid Larva |
| 970 | Mite |
| BOS | Bosmina |
| CAL | Calanoid |
| CYC | Cyclopoid |
| DIA | Diaphanosoma |
| DPH | Daphnia |
| DPT | Dioptera |
| HOL | Holopedium |
| LGB | Large Beast |
| NP | Nauplii |
| ROT | Rotifer |
| TOT | Total |
|  |  |
| LTPR | Little Trapper Lake |
| TPR | Trapper Lake |
| TLTN | Tahltan Lake |
| TATS | Tatsamenie Lake |
| TUYA | Tuya Lake |
| LTATS | Little Tatsamenie Lake |


| BIOMASS, DENSITY AND MEAN SIZE FOR ZOOPLANKTON IN 6 TRANSBOUNDARY LAKES, 1987-1995 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| A GLOSSARY OF SHORTENED TERMS IS PROVIDED IN THE ABOVE CODES. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| Biomass ( $\mathrm{mg} / \mathrm{m}^{3}$ ) | LTATS | 6/8/90 | 1 | 13:20 | 0.0067 | 0.6087 | 0.0428 |  | 20.6358 | 0.0605 |  | 0.664 | 0.2551 | 50.8278 | 22.2736 |  |
| Density (\#/m ${ }^{3}$ ) | LTATS | 6/8/90 | 1 | 13:20 | 2.6667 | 33.3333 | 2.2222 |  | 1088.8900 | 4.4444 |  | 2022.22 | 311.11 | 34.2222 | 3499.11 | $\mathrm{DLA}=141 . \mathrm{CAL}=060 \mathrm{LGB}=900$ |
| Mean size (mm) | LTATS | 6/8/90 | 1 | 13:20 | 0.3468 | 0.4135 | 0.7076 |  | 0.6635 | 0.5898 |  | 0.1501 | 0.1254 | 0.6198 | 0.3159 |  |
| Biomass ( $\mathrm{mg/m} \mathrm{~m}^{3}$ ) | LTATS | 6/8/90 | 1 | 19:00 | 0.003 | 0.0731 | 13.8898 |  | 1.7442 | 23.0149 | 5.2085 | 0.0103 | 11.0427 | 76.975 | 54.985 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTATS | 7/14/90 | 1 | 19:00 | 1.7778 | 3.5556 | 32 |  | 55.1111 | 33.7778 | 179.56 | 44.4444 | 13466.6632 | 5.3333 | 13821.9717 | DIA $=141 . \mathrm{CAL}=020 . \mathrm{LGB}=900$ |
| Mean size (mm) | LTATS | 7/14/90 | 1 | 19:00 | 0.3197 | 0.4246 | 1.5999 |  | 0.7980 | 2.4823 | 0.7831 | 0.135 | 0.1433 | 1.7089 | 0.164 |  |
| Biomass (mg/m3) | LTATS | 9/6/90 | 1 | 9:30 |  | 0.2323 | 0.2954 |  | 5.6250 |  | 0.0609 | 0.0816 | 1.4965 | 0.317 | 7.7917 |  |
| Density (\#/m ${ }^{3}$ ) | LTATS | 9/6/90 | 1 | 9:30 |  | 9 | 5 |  | 325.0000 |  | 0.5 | 350 | 1825 | 1 | 2515.5 | LGB=OSTRACOD |
| Mean size (mm) | LTATS | 9/6/90 | 1 | 9:30 |  | 0.4766 | 0.9299 |  | 0.6370 |  | 1.3377 | 0.129 | 0.1057 | 0.488 | 0.1809 |  |
| Biomass (mg/m3) | LTATS | 10/2/90 | 1 | 12:00 |  | 0.1622 | 0.1556 |  | 24.0322 |  |  | 0.2034 | 3.3073 | 5.5648 | 27.8607 |  |
| Density (\#/m $\mathrm{m}^{3}$ ) | LTATS | 10/2/90 | 1 | 12:00 |  | 4.5 | 4.5 |  | 1683.3300 |  |  | 1133.33 | 4033.33 | 0.5 | 6859.5 | $\underline{L G B}=900$ |
| Mean size (mm) | LTATS | 10/2/90 | 1 | 12:00 |  | 0.561 | 0.8307 |  | 0.5945 |  |  | 0.1202 | 0.1004 | 1.6237 | 0.2258 |  |
| Biomass (mg/m3) | LTPR | 9/2/87 |  |  |  | 3.4334 | 0.0956 |  | 6.6842 |  |  | 0.3676 | 25.5184 |  | 36.0993 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 9/2/87 |  |  |  | 320 | 1.0667 |  | 253.8700 |  |  | 800 | 31119.3129 |  | 32494.5106 |  |
| Mean size (mm) | LTPR | 9/2/87 |  |  |  | 0.3159 | 1.1793 |  | 0.7477 |  |  | 0.171 | 4.49303 |  | 0.1857 |  |
| Biomass (mg/m3) | LTPR | 8/20/88 | South side |  | 0.0137 | 259.52 |  |  | 51.9163 |  |  |  | 11.1083 |  | 322.56 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 8/20/88 | South side |  | 312 | 4.34189 |  |  | 1813.3300 |  |  |  | 13546.5904 |  | 37336.1902 | DIA $=141$ |
| Mean size (mm) | LTPR | 8/20/88 | South side |  | 0.4236 | 0.3362 |  |  | 0.7790 |  |  |  | 0.1798 |  | 0.301 |  |
| Biomass (mg/m3) | LTPR | 6/8/89 | 1 |  |  | 1.1923 | 0.081 |  | 35.4935 |  |  | 10.0394 | 2.9739 | 128.94 | 49.78 |  |
| Density (H/m ${ }^{3}$ ) | LTPR | 6/8/89 | 1 |  |  | 108.8 | 3.2 |  | 3768.8900 |  |  | 12870.9321 | 3626.67 | 8.5333 | 20386.844 | LGB $=900$ |
| Mean size (mm) | LTPR | 6/8/89 | 1 |  |  | 0.3217 | 0.7681 |  | 0.4756 |  |  | 0.2007 | 0.0964 | 1.6177 | 0.2343 |  |
| Biomass (mg/m3) | LTPR | 7/14/89 | 1 | 14:00 | 0.0027 | 14.2918 | 0.4532 |  | 10.5062 |  | 0.2417 | 0.0765 | 19.3126 | 0.889 | 44.8848 |  |
| Biomass (mg/m3) | LTPR | 7/14/89 | 2 | 14:00 | 0.002 | 2.8681 | 1.2019 |  | 23.8975 | 0.2242 | 0.1383 | 0.4313 | 8.4318 |  | 37.195 |  |
| Density (\#/m3) | LTPR | 7/14/89 | 1 | 14:00 | 0.2667 | 1621.33 | 13.0667 |  | 1450.6700 |  | 4.2667 | 256 | 23551.5774 | 0.2667 | 26897.3813 | DLA $=$ chydoridac. $\mathrm{LGB}=$ araneida |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 7/14/89 | 2 | 14:00 | 2.1333 | 230.4 | 34.1333 |  | 2005.3300 | 4.2667 | 2.1333 | 512 | 10282.5303 |  | 13072.8416 | DIA ${ }^{\text {I }}$ [1 |
| Mean size (mm) | LTPR | 7/14/89 | 1 | 14:00 | 0.5806 | 0.2897 | 0.8246 |  | 0.4574 |  | 1.0067 | 0.1471 | 0.0841 | 1.0864 | 0.1178 |  |
| Mean size (mm) | LTPR | 7/14/89 | 2 | 14:00 | 0.2636 | 0.3389 | 0.7866 |  | 0.5413 | 0.9795 | 1.062 | 0.2098 | 0.1088 |  | 0.1854 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | LTPR | 8/30/89 | 1 | 16:00 |  | 17.1439 | 0.0532 |  | 5.2196 |  | 0.0449 | 0.0706 | 11.4144 | 0.0382 | 33.9465 |  |
| Biomass (mg/m3) | LTPR | 8/30/89 | 2 | 16:30 |  | 4.791 | 0.0221 |  | 8.3167 |  |  | 0.1176 | 2.46 | 0.1132 | 15.7074 |  |
| Delssity (\#/m³) | LTPR | 8/30/89 | 1 | 16:00 |  | 2933.33 | 0.5333 |  | 640.0000 |  | 1.0667 | 266.67 | 13919.704 | 0.5333 | 17761.9237 | LGB=450 |
| Density (\#/m) | LTPR | 8/30/89 | 2 | 16:30 |  | 546.67 | 0.5333 |  | 506.6700 |  |  | 226.67 | 3000 | 0.2667 | 4280.8 | LGB=450 |
| Mean size (mm) | LTPR | 8/30/89 | 1 | 16:00 |  | 0.2422 | 1.2412 |  | 0.4864 |  | 0.9006 | 0.1389 | 0.0881 | 0.302 | 0.1287 |  |
| Mean size (mm) | LTPR | 8/30/89 | 2 | 16:30 |  | 0.2871 | 0.8907 |  | 0.6134 |  |  | 0.1776 | 0.1308 | 0.5465 | 0.2105 |  |
| Biomass (mg/m3) | LTPR | 10/2/89 | 1 | 16:30 |  | 1.2859 | 0.2506 |  | 1.2970 |  | 0.1131 | 0.0153 | 0.6997 | 4.8288 | 3.6616 |  |
| Biomass (mg/m3) | LTPR | 10/2/89 | 2 | 16:15 | 0.0021 | 5.2504 | 0.0398 |  | 5.7597 |  | 0.1522 | 0.1108 | 3.7173 | 0.2904 | 15.0323 |  |
| Density ( $\% / \mathrm{m}^{3}$ ) | LTPR | 10/2/89 | 1 | 16:30 |  | 140.95 | 1.3333 |  | 83.8095 |  | 3.4667 | 68.5714 | 853.33 | 0.2667 | 1151.73 | LGB=900 |
| Density (\#/m) | LTPR | 10/2/89 | 2 | 16:15 | 0.2667 | 480 | 0.2667 |  | 266.6700 |  | 2.6667 | 231.11 | 4533.33 | 0.5333 | 5514.84 | LGB $=900$ |
| Mean size (mm) | LTPR | 10/2/89 | 1 | 16:30 |  | 0.2992 | 1.3976 |  | 0.5983 |  | 0.8208 | 0.1308 | 0.0919 | 1.9079 | 0.1605 |  |
| Mean size (mm) | LTPR | 10/2/89 | 2 | 16:15 | 0.5313 | 0.3268 | 1.4196 |  | 0.6890 |  | 1.0023 | 0.1745 | 0.1448 | 0.5927 | 0.1888 |  |
| Biomass (mg/m3) | LTPR | 6/10/90 | 1 | 9:00 | 0.0088 | 0.6477 |  |  | 7.1704 |  |  | 0.7713 | 0.0875 | 22.1156 | 8.6857 |  |
| Biomass (mg/m3) | LTPR | 6/10/90 | 2 | 9:30 |  | 0.0391 | 0.152 |  | 6.3789 |  |  | 0.4287 | 0.123 | 41.3774 | 7.1217 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 6/10/90 | 1 | 9:00 | 0.2667 | 73.3333 |  |  | 693.3300 |  |  | 1220 | 106.67 | 1.6 | 2095.2 | DIA ${ }^{\text {arancida (spider), LGB }}$ L 900 |
| Density (\#/m) | LTPR | 6/10/90 | 2 | 9:30 |  | 0.8 | 2.1333 |  | 330.0000 |  |  | 710 | 150 | 0.5333 | 1193.47 | LGB $=900$ |
| Mean size (mm) | LTPR | 6/10/90 | 1 | 9:00 | 0.8602 | 0.2967 |  |  | 0.4948 |  |  | 0.1905 | 0.1097 | 1.6379 | 0.292 |  |
| Mean size (mm) | LTPR | 6/10/90 | 2 | 9:30 |  | 0.6508 | 1.0647 |  | 0.6124 |  |  | 0.1876 | 0.1063 | 3.0567 | 0.298 |  |
| Biomass (mg/m3) | LTPR | 7/17/90 | 1 | 14:30 |  | 3.0171 |  |  | 72.4953 |  |  | 0.1825 | 0.1166 |  | 75.8115 |  |
| Biomass (mg/m3) | LTPR | 7/17/90 | 2 | 14:00 |  | 3.5781 |  |  | 72.0216 |  |  | 1.0917 | 0.5029 | 13.8463 | 22.19 |  |
| Density (\#/m) | LTPR | 7/17/90 | 1 | 14:30 |  | 337.78 |  |  | 4960.0000 |  |  | 248.89 | 142.22 |  | 5688.89 |  |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Density (\#/m ${ }^{3}$ ) | LTPR | 7/17/90 | 2 | 14:00 |  | 293.33 |  |  | 5813.3300 |  |  | 906.67 | 613.33 | 0.5333 | 7627.2 |  |
| Mean size (mm) | LTPR | 7/17/90 | 1 | 14:30 |  | 0.2923 |  |  | 0.5822 |  |  | 0.2027 | 0.101 |  | 0.5364 |  |
| Mean size (mm) | LTPR | 7/17/90 | 2 | 14:00 |  | 0.3373 |  |  | 0.5408 |  |  | 0.2232 | 0.1082 | 2.1534 | 0.46 |  |
| Biomass (mg/m3) | LTPR | 9/4/90 | 1 | 14:50 | 0.0022 | 8.9589 | 0.0571 | 0.0009 | 3.8287 |  |  | 0.0567 | 1.0204 | 0.2678 | 13.925 |  |
| Biomass (mg/m3) | LTPR | 9/4/90 | 2 | 13:50 | 0.0014 | 8.658 | 0.12 |  | 3.8113 |  |  | 0.068 | 1.2318 | 0.7512 | 13.8905 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 9/4/90 | 1 | 14:50 | 0.5333 | 951.11 | 1.6 | 0.2667 | 204.4400 |  |  | 222.22 | 1244.44 | 0.5333 | 2625.16 |  |
| Density (\#/m) | LTPR | 9/4/90 | 2 | 13:50 | 0.8 | 942.22 | 3.2 |  | 186.6700 |  |  | 160 | 1502.22 | 0.5333 | 2795.64 | DIA $=141, \mathrm{LGB}=900$ |
| Mean size (mm) | LTPR | 9/4/90 | 1 | 14:50 | 0.4321 | 0.3042 | 0.8772 | 0.4103 | 0.6028 |  |  | 0.1341 | 0.1008 | 0.5409 | 0.2171 |  |
| Mean size (mm) | LTPR | 9/4/90 | 2 | 13:50 | 0.319 | 0.302 | 0.8827 |  | 0.6727 |  |  | 0.1615 | 0.1021 | 0.8078 | 0.2121 |  |
| Biomass (mg/m3) | LTPR | 10/2/90 | 1 |  |  | 2.37 | 0.0394 |  | 0.6671 |  |  | 0.0696 | 0.6524 | 2.0479 | 3.7985 |  |
| Biomass (mg/m3) | LTPR | 10/2/90 | 2 | 13:30 |  | 1.7829 | 0.2636 |  | 1.8044 |  |  | 0.0254 | 0.2962 | 4.7691 | 4.1725 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 10/2/90 | 1 |  |  | 248.89 | 1.0667 |  | 44.4444 |  |  | 240 | 795.56 | 0.8 | 1330.76 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | LTPR | 10/2/90 | 2 | 13:30 |  | 172.12 | 8.2667 |  | 92.1212 |  |  | 101.82 | 361.21 | 0.5333 | 736.07 | $L G B=900$ |
| Mean size (mm) | LTPR | 10/2/90 | 1 |  |  | 0.3069 | 0.8858 |  | 0.6055 |  |  | 0.1422 | 0.1016 | 0.8191 | 0.1652 |  |
| Mean size (mm) | LTPR | 10/2/90 | 2 | 13:30 |  | 0.3165 | 0.8281 |  | 0.6608 |  |  | 0.1341 | 0.0933 | 1.3627 | 0.2313 |  |
| Biomass (mg/m3) | LTPR | 6/19/91 | 1 | 8:00 |  | 0.5257 |  |  | 61.1518 |  |  |  | 10.7584 | 0.8379 | 72.4358 |  |
| Biomass (mg/m3) | LTPR | 6/19/91 | 2 | 8:00 |  | 0.1077 |  |  | 30.0741 |  | 0.0211 | 0.217 | 5.7145 | 6.2896 | 36.1345 |  |
| Density (\#/m) | LTPR | 6/19/91 | 1 | 8:00 |  | 53.3333 |  |  | 4480.0000 |  |  |  | 13119.8841 | 1.0667 | 17654.2796 | LGB $=900$ |
| Density (\#/m) | LTPR | 6/19/91 | 2 | 8:00 |  | 35.5556 |  |  | 3448.8900 |  | 0.5333 | 320 | 6968.89 | 1.0667 | 10774.8194 | LGB $=900$ |
| Mean size (mm) | LTPR | 6/19/91 | 1 | 8:00 |  | 0.3132 |  |  | 0.5601 |  |  |  | 0.0994 | 0.671 | 0.217 |  |
| Mean size (mm) | LTPR | 6/19/91 | 2 | 8:00 |  | 0.1823 |  |  | 0.4930 |  | 0.8868 | 0.1891 | 0.1378 | 1.3137 | 0.2533 |  |
| Biomass (mg/m3) | LTPR | 7/22/91 | 1 | 9:45 |  | 1.9603 |  |  | 31.5551 |  |  | 0.0144 | 2.0603 |  | 35.5902 |  |
| Biomass (mg/m3) | LTPR | 7/22/91 | 2 | 9:00 |  | 0.6276 |  |  | 24.5511 |  |  | 0.0439 | 7.4638 |  | 32.6863 |  |
| Density (\#/m) | LTPR | 7/22/91 | 1 | 9:45 |  | 237.04 |  |  | 2275.5600 |  |  | 23.7037 | 2512.59 |  | 5048 |  |
| Density (\#/m) | LTPR | 7/22/91 | 2 | 9:00 |  | 71.1111 |  |  | 1635.5600 |  |  | 106.67 | 9102.22 |  | 10915.4087 |  |
| Mean size (mm) | LTPR | 7/22/91 | 1 | 9:45 |  | 0.283 |  |  | 0.5778 |  |  | 0.1921 | 0.1074 |  | 0.328 |  |
| Mean size (mm) | LTPR | 7/22/91 | 2 | 9:00 |  | 0.2956 |  |  | 0.5879 |  |  | 0.1622 | 0.129 |  | 0.1992 |  |
| Biomass (mg/m3) | LTPR | 8/28/91 | 1 | 19:00 |  | 69.2361 | 0.003 | 0.003 | 53.5657 |  |  | 0.1774 | 10.8634 | 0.1682 | 134.01 |  |
| Biomass (m/m3) | LTPR | 8/28/91 | 2 | 19:30 |  | 16.4119 | , |  | 1.4105 |  |  | 0.0086 | 4.3873 |  | 22.2184 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 8/28/91 | 1 | 19:00 |  | 11455.9207 | 0.32 | 0.32 | 2432.0000 |  |  | 512 | 13247.9903 | 0.32 | 27648.3997 | DPH $=141$. LGB $=180, \mathrm{HOL}=180$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 8/28/91 | 2 | 19:30 |  | 3148.8 |  |  | 81.2800 |  |  | 51.2 | 5350.4 |  | 8631.68 |  |
| Mean size (mm) | LTPR | 8/28/91 | 1 | 19:00 |  | 0.2454 | 0.5693 | 0.2773 | 0.6991 |  |  | 0.1538 | 0.1099 | 0.5869 | 0.2187 |  |
| Mean size (mm) | LTPR | 8/28/91 | 2 | 19:30 |  | 0.2274 |  |  | 0.6389 |  |  | 0.1194 | 0.0958 |  | 0.149 |  |
| Biomass (mg/m3) | LTPR | 10/6/91 | 1 | 12:30 |  | 4.5028 |  |  | 0.5384 |  |  | 0.0267 | 0.4155 |  | 5.4833 |  |
| Biomass (mg/m3) | LTPR | 10/6/91 | 2 |  |  | 3.2505 | 0.0039 |  | 0.6025 |  |  | 0.0405 | 0.6305 | 2.358 | 4.5278 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 10/6/91 | 1 | 12:30 |  | 613.33 |  |  | 46.6667 |  |  | 120 | 506.67 |  | 1286.67 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | LTPR | 10/6/91 | 2 |  |  | 426.67 | 0.5333 |  | 45.3333 |  |  | 133.33 | 786.89 | 0.5333 | 1375.29 | LGB=900. DPH ${ }^{\text {P }} 80$ |
| Mean size (mm) | LTPR | 10/6/91 | 2 | 12:30 |  | 0.2705 |  |  | 0.5540 |  |  | 0.1273 | 0.0728 |  | 0.1896 |  |
| Mean size (mm) | LTPR | 10/6/91 | 2 |  |  | 0.2754 | 0.5195 |  | 0.5746 |  |  | 0.1431 | 0.1031 | 1.1936 | 0.1765 |  |
| Biomass ( $\mathrm{m} / \mathrm{m} \mathrm{m}$ ) | LTPR | 6/19/93 | 1 | 14:00 | 0.0006 | 0.7894 | 0.0226 |  | 3.8629 |  | 0.0436 | 0.2703 | 0.6706 | 23.3964 | 5.66 |  |
| Biomass (mg/m3) | LTPR | 6/19/93 | 2 | 14:30 | 0.0022 | 0.2403 | 0.0393 |  | 17.7434 |  | 0.0757 | 0.4073 | 0.8965 | 32.6699 | 19.4046 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 6/19/93 | 1 | 14:00 | 0.2667 | 44.4444 | 0.8 |  | 248.8900 |  | 0.5333 | 453.33 | 817.78 | 0.8 | 1566.84 | LGB $=900$, DLA $=141$ |
| Density (\#/m ${ }^{3}$ ) | LTPR | 6/19/93 | 2 | 14:30 | 0.2667 | 26.6667 | 1.6 |  | 1253.3300 |  | 0.5333 | 533.33 | 1093.33 | 1.8667 | 2910.93 | LGB $=900$. DLA $=141$ |
| Mean size (mm) | LTPR | 6/19/93 | 1 | 14:00 | 0.353 | 0.4019 | 0.7848 |  | 0.5663 |  | 1.155 | 0.1852 | 0.1173 | 1.8016 | 0.2179 |  |
| Mean size (mm) | LTPR | 6/19/93 | 2 | 14:30 | 0.5432 | 0.3005 | 0.7534 |  | 0.5445 |  | 1.4034 | 0.2042 | 0.1187 | 1.6578 | 0.321 |  |
| Biomass (mg/m3) | LTPR | 7/29/93 | 1 |  |  | 4.2258 | 0.043 |  | 13.0216 | 0.963 | 0.1636 |  | 0.0984 | 0.6735 | 18.5155 | Sample full of rotifers |
| Biomass (mg/m3) | LTPR | 7/29/93 | 2 |  |  | 1.3023 | 0.0025 |  | 31.2311 |  |  | 0.1032 | 0.328 | 8.9571 | 32.9671 |  |
| Density (\#/m) | LTPR | 7/29/93 | 1 |  |  | 680 | 1.3333 |  | 600.0000 | 1.3333 | 3.4667 |  | 120 | 0.2667 | 1406.4 | LGB $=900, \mathrm{CAL}=020$, |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 7/29/93 | 2 |  |  | 186.67 | 0.2667 |  | 2213.3300 |  |  | 240 | 400 | 0.8 | 3041.07 | LGB $=900$ |
| Mean size (mm) | LTPR | 7/29/93 | 1 |  |  | 0.2497 | 0.8389 |  | 0.6888 | 2.5533 | 0.9366 |  | 0.1327 | 0.9904 | 0.4316 |  |
| Mean size (mm) | LTPR | 7/29/93 | 2 |  |  | 0.2626 | 0.5672 |  | 0.5805 |  |  | 0.1658 | 0.1029 | 1.5839 | 0.4657 |  |
| Biomass (mg/m3) | LTPR | 9/8/93 | 1 |  |  | 58.814 | 0.2392 |  | 1.2530 |  | 0.117 | 0.0494 | 4.3733 | 10.2885 | 64.846 |  |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass (mg/m3) | LTPR | 9/8/93 | 2 |  |  | 49.3471 | 0.0586 |  | 2.8258 |  | 0.1721 | 0.4225 | 7.872 | 19.5341 | 60.6981 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 9/8/93 | 1 |  |  | 8747.38 | 21.3333 |  | 91.7333 |  | 4.9778 | 213.33 | 5333.33 | 2.8444 | 14414.8458 | LGB $=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 9/8/93 | 2 |  |  | 7253.33 | 4.2667 |  | 168.5300 |  | 8.5333 | 853.33 | 9600 | 4.2667 | 17892.0458 | LGB 900 |
| Mean size (mm) | LTPR | 9/8/93 | 1 |  |  | 0.2596 | 0.5863 |  | 0.5577 |  | 0.7307 | 0.1348 | 0.1144 | 0.9909 | 0.2067 |  |
| Mean size (mm) | L.TPR | 9/8/93 | 2 |  |  | 0.2606 | 0.6297 |  | 0.6297 |  | 0.6867 | 0.1752 | 0.1435 | 1.1652 | 0.1975 |  |
| Biomass (mg/m3) | LTPR | 6/19/94 | 1 | 7:00 |  | 0.2755 | 0.0111 |  | 35.9837 |  |  | 0.3642 | 5.77728 | 285.3 | 42.41 |  |
| Biomass (mg/m3) | LTPR | 6/19/94 | 2 | 8:00 |  | 0.8742 | 0.0481 |  | 31.3093 |  |  | 0.2088 | 3.7611 | 57.8744 | 36.2015 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 6/19/94 | 1 | 7:00 |  | 17.0667 | 2.1333 |  | 1920.0000 |  |  | 213.33 | 7040 | 3.2 | 9195.73 | $\underline{L G B}=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 6/19/94 | 2 | 8:00 |  | 53.3333 | 2.1333 |  | 2453.3300 |  |  | 213.33 | 4586.67 | 4.2667 | 7313.07 | LGB=900 |
| Mean size (mm) | LTPR | 6/19/94 | 1 | 7:00 |  | 0.3896 | 0.4651 |  | 0.6388 |  |  | 0.2589 | 0.1357 | 2.8712 | 0.2451 |  |
| Mean size (mm) | LTPR | 6/19/94 | 2 | 8:00 |  | 0.3954 | 0.7382 |  | 0.5455 |  |  | 0.2269 | 0.1248 | 1.6682 | 0.272 |  |
| Biomass (mg/m3) | LTPR | 7/23/94 | 1 | 14:45 |  | 4.3284 | 0.6663 |  | 44.2791 |  |  | 0.0119 | 0.8528 |  | 50.1385 |  |
| Biomass (mg/m3) | LTPR | 7/23/94 | 2 | 16:00 |  | 35.4347 | 5.0869 |  | 178.8700 |  |  | 0.1802 | 6.1227 |  | 225.69 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 7/23/94 | 1 | 14:45 |  | 213.33 | 13.8667 |  | 1813.3300 |  |  | 26.6667 | 1040 |  | 3107.2 |  |
| Density (\#/m ${ }^{3}$ ) | L.TPR | 7/23/94 | 2 | 16:00 |  | 2880 | 145.07 |  | 5866.6700 |  |  | 213.33 | 7466.67 |  | 16571.4305 |  |
| Mean size (mm) | LTPR | 7/23/94 | 1 | 14:45 |  | 0.4303 | 0.9403 |  | 71.9300 |  |  | 0.1716 | 0.2159 |  | 0.5273 |  |
| Mean size (mm) | LTPR | 7/23/94 | 2 | 16:00 |  | 0.3315 | 0.8428 |  | 0.7732 |  |  | 0.2129 | 0.1861 |  | 0.4253 |  |
| Biomass (mg/m3) | LTPR | 9/9/94 | 1 | 14:00 |  | 91.6536 | 2.1069 |  | 7.1230 |  | 5.8486 | 0.1324 | 12.1579 |  | 119.02 |  |
| Biomass (mg/m3) | LTPR | 9/9/94 | 2 | 14:20 | 0.0256 | 12.4204 | 1.9183 | 0.0577 | 7.8278 |  |  | 0.0771 | 0.7081 |  | 23.0351 |  |
| Density ( ${\mathrm{H} / \mathrm{m}^{3} \text { ) }}^{\text {a }}$ | LTPR | 9/9/94 | 1 | 14:00 |  | 8853.33 | 179.2 |  | 0.7256 |  | 166.4 | 213.33 | 14826.5464 |  | 24532.9628 |  |
| Density (\#/m ${ }^{3}$ ) | LTPR | 9/9/94 | 2 | 14:20 | 4.2667 | 1015.87 | 162.13 | 4.2667 | 223.4900 |  |  | 101.59 | 863.49 |  | 2375,11 |  |
| Mean size (mm) | LTPR | 9/9/94 | 1 | 14:00 |  | 0.3162 | 0.5971 |  | 0.7256 |  | 0.8102 | 0.1921 | 0.1474 |  | 0.2234 |  |
| Mean size (mm) | LTPR | 9/9/94 | 2 | 14:20 | 0.4884 | 0.34 | 0.5887 | 0.6394 | 0.8174 |  |  | 0.2072 | 0.1406 |  | 0.3245 |  |
| Biomass (mg/m3) | LTPR | 7/24/95 | 1 | 9:50 |  | 0.0122 | 0.0559 |  | 29.1352 |  |  | 0.3469 | 0.3717 |  | 29.922 |  |
| Biomass (mg/m3) | LTPR | 7/24/95 | 2 |  |  | 0.0164 | 0.0309 |  | 1.6497 |  |  | 0.0191 | 0.0276 |  | 1.7438 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | LTPR | 7/24/95 | 1 | 9:50 |  | 1.0667 | 4.2667 |  | 1520.0000 |  |  | 280 | 453.33 |  | 2258.67 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | LTPR | 7/24/95 | 2 |  |  | 1.6 | 2.6667 |  | 88.0000 |  |  | 25.6 | 33.6 |  | 151.47 |  |
| Mean size (mm) | LTPR | 7/24/95 | 1 | 9:50 |  | 0.3354 | 0.6166 |  | 0.6552 |  |  | 0.2466 | 0.1602 |  | 0.505 |  |
| Mean size (mm) | LTPR | 7/24/95 | 2 |  |  | 0.3135 | 0.5911 |  | 0.6491 |  |  | 0.2019 | 0.2142 |  | 0.4725 |  |
| Biomass (mg/m3) | LTPR | 9/14/95 | 1 | 14:37 |  | 10.1035 | 2.0261 |  | 18.5735 |  |  | 0.9644 | 1.5416 |  | 33.2091 |  |
| Biomass (mg/m3) | LTPR | 9/14/95 | 2 |  |  | 2.7816 | 0.7681 |  | 0.5510 |  |  | 0.4401 | 3.3784 |  | 7.9191 |  |
| Density (\#/m) | LTPR | 9/14/95 | 1 | 14:37 |  | 880 | 90 |  | 880.0000 |  |  | 1100 | 1880 |  | 4830 | haul depth 15 m |
| Density (\#/m) | LTPR | 9/14/95 | 2 |  |  | 240 | 61.6 |  | 32.8000 |  |  | 640 | 4120 |  | 5094.4 | haul depth 15m |
| Mean size (mm) | LTPR | 9/14/95 | 1 | 14:37 |  | 0.3294 | 0.7315 |  | 0.6706 |  |  | 0.2151 | 0.155 |  | 0.3051 |  |
| Mean size (mm) | LTPR | 9/14/95 | 2 |  |  | 0.3318 | 0.6033 |  | 0.6226 |  |  | 0.1959 | 0.1357 |  | 0.1613 |  |
| Biomass (mg/m3) | TATS | 8/31/87 | Trans 3 |  |  | 52.638 | 101.47 |  | 352.2400 |  |  | 11.0202 | 3.6526 |  | 521.02 |  |
| Biomass (mg/m3) | TATS | 8/31/87 | Trans 6 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/31/87 | Trans 3 |  |  | 1152 | 1938.77 |  | 6502.4000 |  |  | 14438.0985 | 4454.4 |  | 28485.9036 |  |
| Density ( $1 / \mathrm{m}^{3}$ ) | TATS | 8/31/87 | Trans 6 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mean size (mm) | TATS | 8/31/87 | Trans 3 |  |  | 0.6229 | 0.9719 |  | 0.9108 |  |  | 0.2039 | 0.171 |  | 0.4293 |  |
| Mean size (mm) | TATS | 8/31/87 | Trans 6 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Biomass (mg/m3) | TATS | 7/22/88 | 1 | 9:30 |  | 104 | 43.4572 |  | 311.4000 |  |  | 2.9555 | 4.9661 | 207.37 | 466.75 |  |
| Biomass (mg/m3) | TATS | 7/22/88 | 2 | 10:30 |  | 94.5558 | 12.7863 |  | 151.7400 |  |  | 1.0868 | 1.968 | 87.9957 | 262.1343 |  |
| Density (\#/m) | TATS | 7/22/88 | 1 | 9:30 |  | 3136 | 1408. |  | 11903.9200 |  |  | 2560 | 6016 | 1.28 | 25024.7645 | LGB=900 |
| Density (\#/m) | TATS | 7/22/88 | 2 | 10:30 |  | 2848 | 448 |  | 6464,0000 |  |  | 896 | 2400 | 0.64 | 13056.597 | $L G B=900$ |
| Mean size (mm) | TATS | 7/22/88 | 1 | 9:30 |  | 0.5357 | 0.7926 |  | 0.7213 |  |  | 0.2415 | 0.1759 | 3.7255 | 0.522 |  |
| Mean size (mm) | TATS | 7/22/88 | 2 | 10:30 |  | 0.537 | 0.6784 |  | 0.6962 |  |  | 0.2454 | 0.1742 | 3.7446 | 0.5341 |  |
| Biomass (mg/m3) | TATS | 8/22/88 | 1 | 12:20 |  | 60.432 | 13.2027 |  | 98.2797 |  |  | 2.2928 | 1.277 | 62.3783 | 175.4817 |  |
| Biomass (mg/m3) | TATS | 8/22/88 | 2 | 11:30 |  | 52.4971 | 22.5871 |  | 95.0346 |  |  | 2.6754 | 0.9971 |  | 173.79 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/22/88 | 1 | 12:20 |  | 1557.33 | 725.33 |  | 3648.0000 |  |  | 1834 | 1557.33 | 0.32 | 9322.99 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/22/88 | 2 | 11:30 |  | 1408 | 746.67 |  | 3733.3300 |  |  | 2026.67 | 1216 |  | 9130.67 |  |
| Mean size (mm) | TATS | 8/22/88 | 1 | 12:20 |  | 0.5739 | 0.6756 |  | 0.7407 |  |  | 0.2486 | 0.1759 | 4.2167 | 0.5167 |  |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean size (mm) | TATS | 8/22/88 | 2 | 11:30 | ! | . 0.5654 | 0.7733 |  | 0.7175 |  |  | 0.2536 | 0.1782 |  | 0.5238 |  |
| Biomass (mg/m3) | TATS | 9/24/88 | 1 | 19:00 |  | 40.346 | 64.7678 | 0.0055 | 210.6300 | 0.8921 | 1.8927 | 2.1042 | 1.0496 | 44.5739 | 321.6861 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 9/24/88 | 1 | 19:00 |  | 1152 | 1600 | 0.32 | 7264.0000 | 32.32 | 32 | 1440 | 1280 | 0.32 | 12800.8851 | LGB=900 |
| Mean size (mm) | TATS | 9/24/88 | 1 | 19:00 |  | 0.5508 | 0.8965 | 0.6904 | 0.7507 | 0.6601 | 1.027 | 0.2632 | 0.1803 | 3.7698 | 0.6396 |  |
| Biomass (mg/m3) | TATS | 6/8/89 | 1 | 11:00 |  | 2.4167 | 9.2954 |  | 249.0200 |  |  | 1.8949 | 2.2042 |  | 264.83 |  |
| Biomass (mg/mis) | TATS | 6/8/89 | 2 | 11:30 |  | 11.9213 | 21.4037 |  | 413.3300 |  |  | 2.7401 | 2.4491 | 203.53 | 451.84 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 6/8/89 | 1 | 11:00 |  | 192 | 192 |  | 12735.9105 |  |  | 5632 | 2688 |  | 21439.764 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 6/8/89 | 2 | 11:30 |  | 597.33 | 597.33 |  | 16981.2635 |  |  | 7168 | 2986.67 | 2.56 | 28332.8329 | LGB $=900$ |
| Mean size (mm) | TATS | 6/3/89 | 1 | 11:00 |  | 0.3363 | 0.8886 |  | 0.6414 |  |  | 0.1496 | 0.0902 |  | 0.4426 |  |
| Mean size (mm) | TATS | 6/8/89 | 2 | 11:30 |  | 0.4167 | 0.7872 |  | 0.7081 |  |  | 0.1587 | 0.1203 | 3.1223 | 0.5029 |  |
| Biomass (mg/m3) | TATS | 7/13/89 | 1 | 11:40 |  | 232.63 | 84.9644 |  | 254.0300 |  |  | 4.4563 | 17.4234 |  | 593.5 |  |
| Biomass (mg/m3) | TATS | 7/13/89 | 2 | 11:20 |  | 7.9011 | 17.6661 |  | 85.9659 |  |  | 3.0146 | 6.1716 |  | 120.72 |  |
| Densiv. (\#/m ${ }^{3}$ ) | TATS | 7/13/89 | 1 | 11:40 |  | 7168 | 2944 |  | 8192.0000 |  |  | 11903.92 | 21247.6058 |  | 51455.2864 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 7/13/89 | 2 | 11:20 |  | 307.2 | 512 |  | 2662.4000 |  |  | 5580.8 | 7526.4 |  | 16588.6101 |  |
| Mean size (mm) | TATS | 7/13/89 | 1 | 11:40 |  | 0.523 | 0.7474 |  | 0.4574 |  |  | 0.1555 | 0.0896 |  | 0.3091 |  |
| Mean size (mm) | TATS | 7/13/89 | 2 | 11:20 |  | 0.4812 | 0.732 |  | 0.7610 |  |  | 0.18 | 0.1172 |  | 0.2674 |  |
| Biomass (mg/m3) | TATS | 8/30/89 | 1 | 11:50 |  | 33.9275 | 94.9818 |  | 174.3500 |  |  | 3.2493 | 6.4375 |  | 312.94 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/30/89 | 2 | 11:20 |  | 127.65 | 112.63 |  | 86.0279 | 0.525 |  | 4.113 | 8.4668 |  | 339.41 |  |
| Density ( $\# / m^{3}$ ) | TATS | 8/30/89 | 1 | 11:50 |  | 1024 | 2389.33 |  | 5034.6700 |  |  | 9898.67 | 7850.67 |  | 34068.5714 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/30/89 | 2 | 11:20 |  | 3413.33 | 3242.67 |  | 2304.0000 | 1.28 |  | 11775.7886 | 10325.2363 |  | 31062.0418 | CAL $=020$ |
| Mean size (mm) | TATS | 8/30/89 | 1 | 11:50 |  | 0.5381 | 0.8792 |  | 0.7936 |  |  | 0.1469 | 0.0965 |  | 0.3382 |  |
| Mean size (mm) | TATS | 8/30/89 | 2 | 11:20 |  | 0.5632 | 0.8454 |  | 0.7949 | 2.0847 |  | 0.1528 | 0.1247 |  | 0.3086 |  |
| Biomass (mg/m3) | TATS | 10/4/89 | 1 | 15:00 |  | 5.4015 | 28.2782 |  | 44.1322 |  |  | 0.9815 | 2.54 |  | 81.3334 |  |
| Biomass (mg/m3) | TATS | 10/4/89 | 2 | 14:45 |  | 38.8386 | 198.55 |  | 107.0000 |  |  | 2.0393 | 3.6946 |  | 350.12 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 10/4/89 | 1 | 15:00 |  | 147.2 | 486.4 |  | 1408.0000 |  |  | 2432 | 3097.6 |  | 7571.2 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 10/4/89 | 2 | 14:45 |  | 921.6 | 3132.2 |  | 2560.0000 |  |  | 4403.2 | 4505.6 |  | 15513.5075 |  |
| Mean size (mm) | TATS | 10/4/89 | 1 | 15:00 |  | 0.5641 | 1.0229 |  | 0.7471 |  |  | 0.1586 | 0.095 |  | 0.3054 |  |
| Mean size (mm) | TATS | 10/4/89 | 2 | 14:45 |  | 0.5994 | 1.0567 |  | 0.8326 |  |  | 0.1697 | 0.1276 |  | 0,471 |  |
| Biomass (mg/m3) | TATS | 6/8/90 | 1 | 13:50 |  | 1.5433 | 4.7243 |  | 92.2893 |  |  | 2.2629 | 1.207 |  | 102.03 |  |
| Biomass (mg/m3) | TATS | 6/8/90 | 2 | 13:30 |  | 5.5715 | 4.1399 |  | 30.8883 |  |  | 1.934 | 1.9243 | 5.2663 | 44.4581 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 6/8/90 | 1 | 13:50 |  | 51.2 | 212.48 |  | 3200.0000 |  |  | 7072 | 1472 |  | 12007.4282 |  |
| Density ( ${ }^{\left(1 / \mathrm{m}^{3} \text { ) }\right.}$ | TATS | 6/8/90 | 2 | 13:30 |  | 192 | 116.48 |  | 1173.3300 |  |  | 3477.33 | 2346.67 | 2.56 | 7308.37 |  |
| Mean size (mm) | TATS | 6/8/90 | 1 | 13:50 |  | 0.5012 | 0.6611 |  | 0.7602 |  |  | 0.1476 | 0.0924 |  | 0.3147 |  |
| Mean size (mm) | TATS | 6/8/90 | 2 | 13:30 |  | 0.493 | 0.7331 |  | 0.7202 |  |  | 0.1804 | 0.0987 | 0.8301 | 0.2581 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | TATS | 7/15/90 | 1 | 10:15 |  | 143.63 | 37.4739 |  | 84.8700 |  |  | 2.5302 | 13.183 |  | 281.68 |  |
| Biomass (mg/m3) | TATS | 7/15/90 | 2 | 10:00 |  | 59.3212 | 11.1126 |  | 48.3389 |  |  | 2.5228 | 7.3997 |  | 128.7 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 7/15/90 | 1 | 10:15 |  | 5222.4 | 1241.03 |  | 6348.8000 |  |  | 6144 | 16076.4443 |  | 35032.4088 |  |
| Density ( $/ 1 / \mathrm{m}^{3}$ ) | TATS | 7/15/90 | 2 | 10:00 |  | 1856 | 448 |  | 3008.0000 |  |  | 5312 | 9024 |  | 19647.6223 |  |
| Mean size (mm) | TATS | 7/15/90 | 1 | 10:15 |  | 0.4874 | 0.7717 |  | 52.9900 |  |  | 0.1589 | 0.101 |  | 0.2702 |  |
| Mean size (mm) | TATS | 7/15/90 | 2 | 10:00 |  | 0.5209 | 0.7126 |  | 0.5811 |  |  | 0.1691 | 0.0998 |  | 0.246 |  |
| Biomass (mg/m3) | TATS | 9/8/90 | 2 | 14:17 |  | 13.9349 | 64.7186 |  | 64.7186 |  |  | 3.6857 | 1.5114 | 0.0723 | 138.99 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 9/8/90 | 2 | 14:17 |  | 409.6 | 1382.4 |  | 2355.2000 |  |  | 10854.2545 | 1843.2 | 0.64 | 16845.3485 | LGB=UNK INSECT (HEAD WIDTH |
| Mean size (mm) | TATS | 9/8/90 | 2 | 14:17 |  | 0.5383 | 0.701 |  | 0.7010 |  |  | 0.1504 | 0.097 | 0.3516 | 0.2889 |  |
| Biomass (mg/m3) | TATS | 10/2/90 | 1 | 12:40 |  | 28.8685 | 125.09 |  | 126.0900 |  |  | 0.6529 | 2.3791 |  | 283.08 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 10/2/90 | 1 | 12:40 |  | 896 | 3200 |  | 4949.3300 |  |  | 1877.33 | 2901.33 |  | 13823.8814 |  |
| Mean size (mm) | TATS | 10/2/90 | 1 | 12:40 |  | 0.5311 | 0.8907 |  | 0.6916 |  |  | 0.1517 | 0.1095 |  | 0.5318 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | TATS | 6/22/91 | 1 |  |  | 0.8155 | 3.1401 |  | 293.4100 |  |  | 1.0572 | 0.035 | 95.5878 | 298.4622 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 6/22/91 | 1 |  |  | 58.3333 | 82.56 |  | 8661.3300 |  |  | 4266.67 | 42.6667 | 1.28 | 13139.8377 | LGB $=900$ |
| Mean size (mm) | TATS | 6/22/91 | 1 |  |  | 0.3084 | 0.7565 |  | 0.8092 |  |  | 0.1331 | 0.0874 | 3.0591 | 0.584 |  |
| Biomass (mg/m3) | TATS | 7/15/91 | 1 |  |  | 45.0595 | 33.7417 |  | 270.5100 |  | 207.6 | 0.943 | 4.3783 |  | 562.24 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 7/15/91 | 1 |  |  | 2267.43 | 768 |  | 7972.5700 |  | 3218.29 | 2779.43 | 5339.43 |  | 22344.9815 |  |
| Mean size ( mm ) | TATS | 7/15/91 | 1 |  |  | 0.4221 | 0.9127 |  | 0.8021 |  | 1.0515 | 0.1518 | 0.1015 |  | 0.555 |  |


| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | roT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass (mg/m3) | TATS | 8/17/91 | 1 | 17:00 |  | 499.51 | 12.9444 |  | 138.3700 |  |  | 1.5172 | 5.6678 |  | 658.01 |  |
| Density (\#/m) | TATS | 8/17/91 | 1 | 17:00 |  | 15999.6322 | 576 |  | 6784.0000 |  |  | 8448 | 6912 |  | 38719.5231 |  |
| Mean size (mm) | TATS | 8/17/91 | 1 | 17:00 |  | 0.5197 | 0.7088 |  | 0.6350 |  |  | 0.1185 | 0.097 |  | 0.3797 |  |
| Biomass (mg/m3) | TATS | 9/15/91 | 1 |  |  | 13.9574 | 12.8683 |  | 250.3500 |  |  | 1.4212 | 2.2042 |  | 280.8 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 9/15/91 | 1 |  |  | 448 | 393.61 |  | 6144.0000 |  |  | 6784 | 2688 |  | 16457.3536 |  |
| Mean size (mm) | TATS | 9/15/91 | 1 |  |  | 0.5241 | 0.8015 |  | 0.8273 |  |  | 0.1273 | 0.0777 |  | 0.4075 |  |
| Biomass (mg/m3) | TATS | 6/22/92 | 1 |  |  | 69.7674 | 77.7684 |  | 132.9600 |  |  | 4.9503 | 1.0196 |  | 286.47 |  |
| Biomass (mg/m3) | TATS | 6/22/92 | 2 |  |  | 20.4519 | 1.424 |  | 44.5886 |  |  | 1.2995 | 0.3849 | 58.2157 | 68.1443 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 6/22/92 | 1 |  |  | 2852.57 | 1536 |  | 6509.7100 |  |  | 10971.3482 | 1243.43 |  | 23112.6637 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 6/22/92 | 2 |  |  | 896 | 83.2 |  | 1962.6700 |  |  | 2005.33 | 469.33 | 1.28 | 5417.81 | LGB $=900$ |
| Mean size (mm) | TATS | 6/22/92 | 1 |  |  | 0.455 | 0.8493 |  | 0.6574 |  |  | 0.1667 | 0.0882 |  | 0.3816 |  |
| Mean size (mm) | TATS | 6/22/92 | 2 |  |  | 0.4493 | 0.6079 |  | 0.6846 |  |  | 0.1885 | 0.1017 | 2.5959 | 0.4108 |  |
| Biomass (mg/m3) | rats | 8/2/92 | 1 |  |  | 39.5791 | 35.163 |  | 225.8900 |  |  | 2.694 | 0.5598 |  | 303.89 |  |
| Biomass (mg/m3) | TATS | 8/2/92 | 2 |  |  | 194.51 | 92.4346 |  | 163.8300 |  |  | 4.8911 | 1.9942 | 664.28 | 457.66 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/2/92 | 1 |  |  | 1792 | 2474.67 |  | 6656.0000 |  |  | 3584 | 682.67 |  | 15189.0286 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 8/2/92 | 2 |  |  | 7552 | 2816 |  | 5632.0000 |  |  | 9984 | 2432 | 3.36 | 28419.079 | LGB $=900$ |
| Mean size (mm) | TATS | 8/2/92 | 1 |  |  | 0.4429 | 0.6098 |  | 0.7878 |  |  | 0.1964 | 0.1052 |  | 0.5479 |  |
| Mean size (mm) | TATS | 8/2/92 | 2 |  |  | 0.4738 | 0.7831 |  | 0.7279 |  |  | 0.1709 | 0.1094 | 4.0191 | 0.4176 |  |
| Biomass (mg/m3) | TATS | 8/22/92 | 1 |  |  | 44.1607 | 56.1898 |  | 116.8200 |  |  | 1.7573 | 2.0467 | 269.91 | 220.97 |  |
| Biomass (mg/m3) | TATS | 8/22/92 | 2 |  |  | 129.24 | 129.9 |  | 112.5600 |  |  | 3.1742 | 3.1068 | 468.23 | 377.98 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 8/22/92 | 1 |  |  | 1280 | 1920 |  | 3328.0000 |  |  | 4544 | 2496 | 1.28 | 13569.0674 | LGB $=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 8/22/92 | 2 |  |  | 3379.2 | 3891.2 |  | 5017.6000 |  |  | 8601.6 | 3788.8 | 2.56 | 24680.8436 |  |
| Mean size (mm) | TATS | 8/22/92 | 1 |  |  | 0.549 | 0.8076 |  | 0.7977 |  |  | 0.1578 | 0.0926 | 4.3286 | 0.432 |  |
| Mean size (mm) | TATS | 8/22/92 | 2 |  |  | 0.5745 | 0.8334 |  | 0.6616 |  |  | 0.1563 | 0.0912 | 4.1281 | 0.4135 |  |
| Biomass (mg/m3) | TATS | 9/28/92 | 1 |  |  | 143.48 | 142.17 |  | 274.8400 |  |  | 2.8973 | 3.0438 |  | 566.42 |  |
| Biomass (mg/m3) | TATS | 9/28/92 | 2 |  |  | 51.8297 | 45.2997 |  | 95.760 I |  |  | 1.1245 | 1.4095 |  | 195.42 |  |
| Density (\#/m ${ }^{3}$ ) | Tars | 9/28/92 | 1 |  |  | 3712 | 2688 |  | 6144.0000 |  |  | 8192 | 3712 |  | 24447.8122 |  |
|  | TATS | 9/28/92 | 2 |  |  | 1316.57 | 877.71 |  | 2596.5700 |  |  | 2157.71 | 1718.86 |  | 8667.43 |  |
| Mean size (mm) | TATS | 9/28/92 | 1 |  |  | 0.5702 | 0.9795 |  | 0.8732 |  |  | 0.1512 | 0.0909 |  | 0.4782 |  |
| Mean size (mm) | TATS | 9/28/92 | 2 |  |  | 0.5771 | 0.9738 |  | 0.8015 |  |  | 0.1753 | 0.0997 |  | 0.4898 |  |
| Biomass (mg/m3) | TATS | 6/19/93 | 1 | 20:30 |  | 91.2747 | 2.3042 |  | 205.7800 |  |  | 4.0126 | 10.496 | 88.1341 | 313.8659 |  |
| Density (\#/m) | TATS | 6/19/93 | 1 | 20:30 |  | 4224 | 76.8 |  | 11263.9315 |  |  | 6400 | 12799.7062 | 1.28 | 34765.6217 | LGB $=900$ |
| Mean size (mm) | TATS | 6/19/93 | 1 | 20:30 |  | 0.4308 | 0.7757 |  | 0.6165 |  |  | 0.1893 | 0.0929 | 2.9807 | 0.323 |  |
| Biomass (mg/m3) | TATS | 8/2/93 | 1 |  |  | 130.72 | 99.8252 |  | 165.4800 |  |  | 1.4945 | 5.5629 | 119.83 | 403.08 |  |
| Biomass (mg/m3) | TATS | 8/2/93 | 2 |  |  | 87.283 | 51.6054 |  | 74.0545 |  |  | 0.9447 | 6.2276 |  | 220.12 |  |
| Density ( $\left(1 / \mathrm{m}^{3}\right.$ ) | TATS | 8/2/93 | 1 |  |  | 4736 | 2432 |  | 7296.0000 |  |  | 5120 | 6784 | 1.28 | 26368.7778 | LGB $=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 8/2/93 | 2 |  |  | 3498.67 | 2730.67 |  | 4266.6700 |  |  | 3328 | 7594.67 |  | 21418.5467 |  |
| Mean size (mm) | TATS | 8/2/93 | 1 |  |  | 0.493 | 0.843 |  | 0.6814 |  |  | 0.1426 | 0.0892 | 3.3021 | 0.4056 |  |
| Mean size (mm) | TATS | 8/2/93 | 2 |  |  | 0.4714 | 0.6672 |  | 0.6143 |  |  | 0.1402 | 0.0883 |  | 0.3375 |  |
| Biomass (mg/m3) | TATS | 9/15/93 | 1 |  |  | 7.0765 | 42.3251 |  | 133.1000 |  |  | 1.7884 | 0.5878 |  | 184.88 |  |
| Biomass (mg/m3) | TATS | 9/15/93 | 2 |  |  | 32.4053 | 141.81 |  | 105.8800 |  |  | 1.7959 | 2.4491 |  | 284.35 |  |
| Density (\#/m) | TATS | 9/15/93 | 1 |  |  | 256 | 1484.8 |  | 3840.0000 |  |  | 4505.6 | 716.8 |  | 10803.1399 |  |
| Density (\#/mi ${ }^{\text {a }}$ | TATS | 9/15/93 | 2 |  |  | 1024 | 4693.33 |  | 3498.6700 |  |  | 7680 | 2986.67 |  | 19882.4546 |  |
| Mean size (mm) | TATS | 9/15/93 | 1 |  |  | 0.4944 | 0.7911 |  | 0.7819 |  |  | 0.1596 | 0.0903 |  | 0.4709 |  |
| Mean size (mm) | TATS | 9/15/93 | 2 |  |  | 0.5291 | 0.7985 |  | 0.7459 |  |  | 0.1305 | 0.0856 |  | 0.4103 |  |
| Biomass (mg/m3) | TATS | 6/17/94 | 1 |  |  | 1.3517 | 0.0646 |  | 7.5935 |  |  | 0.2403 | 0.2617 | 124.76 | 9.51 |  |
| Biomass (mg/m3) | TATS | 6/17/94 | 2 |  |  | 115.97 | 4.1209 |  | 122.7900 |  |  | 1.2067 | 0.2099 | 11.9359 | 256.23 |  |
| Density ( $\mathrm{\#} / \mathrm{m}^{3}$ ) | TATS | 6/17/94 | 1 |  |  | 52.5253 | 1.3333 |  | 210.1000 |  |  | 222.22 | 319.19 | 3.5556 | 808.93 | LGB $=900$ |
| Density ( $/ / \mathrm{m}^{3}$ ) | TATS | 6/17/94 | 2 |  |  | 4032 | 51.2 |  | 5184.0000 |  |  | 1216 | 256 | 1.28 | 10740.3887 | LGB=900 |
| Mean size (mm) | TATS | 6/17/94 |  |  |  | 0.4772 | 0.9581 |  | 0.7903 |  |  | 0.2307 | 0.1537 | 2.1007 | 0.3711 |  |
| Meant size (mm) | TATS | 6/17/94 | 2 |  |  | 0.4917 | 1.0915 |  | 0.6935 |  |  | 0.2269 | 0.115 | 1.5307 | 0.5531 |  |
| Biomass (mg/m3) | TATS | 6/18/94 | 1 |  |  |  | 0.0571 |  | 448.3500 |  | 0.1219 | 1.3786 | 2.3091 |  | 452.22 |  |


| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Density (\#/m ${ }^{3}$ ) | TATS | 6/18/94 | 1 | \% |  | : | 1.28 |  | 24319.8001 |  | 1.28 | 1280 | 2816 |  | 28418.4247 |  |
| Mean size (mm) | TATS | 6/18/94 | 1 |  |  | 1 | 0.9502 |  | 0.6303 |  | 1.2227 | 0.2323 | 0.1371 |  | 0.5636 |  |
| Biomass (mg/m3) | TATS | 7/26/94 | 1 | 16:20 |  | 165.49 | 22.0104 |  | 240.9900 |  |  | 3.1095 | 16.3738 |  | 447.97 |  |
| Biomass (mg/m3) | TATS | 7/26/94 | 2 | 18:00 |  | 238.36 | 8.434 |  | 261.2000 |  |  | 3.2125 | 18.8928 |  | 530.12 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 7/26/94 | 1 | 16:20 |  | 3840 | 451.28 |  | 66.5600 |  |  | 5504 | 19967.79 |  | 36419.1663 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 7/26/94 | 2 | 18:00 |  | 5760 | 225.64 |  | 6912.0000 |  |  | 5376 | 23039.8687 |  | 41313.3108 |  |
| Mean size (mm) | TATS | 7/26/94 | 1 | 16:20 |  | 0.6048 | 0.8935 |  | 0.7920 |  |  | 0.1795 | 0.1183 |  | 0.3116 |  |
| Mean size (mm) | TATS | 7/26/94. | 2 | 18:00 |  | 0.5929 | 0.832 |  | 0.8100 |  |  | 0.1865 | 0.1255 |  | 0.317 |  |
| Biomass (mg/m3) | TATS | 9/14/94 | 1 |  |  | 7.0053 | 20.1878 |  | 171.4000 |  |  | 12.0058 | 1.5219 |  | 212.12 |  |
| Biomass (mg/m3) | TATS | 9/14/94 | 2 |  |  | 47.1271 | 96.596 |  | 143.0800 |  |  | 8.3754 | 1.4694 |  | 296.65 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TATS | 9/14/94 | 1 |  |  | 128 | 432 |  | 3712.0000 |  |  | 16575.6283 | 1856 |  | 22703.8757 |  |
| Density (\#/m^3)=10 | TATS | 9/14/94 | 2 |  |  | 938.67 | 2304 |  | 3925.3300 |  |  | 11263.9315 | 1792 |  | 20223.6714 |  |
| Mean size (mm) | TATS | 9/14/94 | 1 |  |  | 0.6836 | 0.9267 |  | 0.8854 |  |  | 0.2005 | 0.1168 |  | 0.3222 |  |
| Mean size (mm) | TATS | 9/14/94 | 2 |  |  | 0.6519 | 0.8892 |  | 0.7761 |  |  | 0.202 | 0.113 |  | 0.4047 |  |
| Biomass (mg/m3) | TATS | 7/18/95 | 1 | 13:00 |  | 94.4762 | 46.0219 |  | 335.2900 |  |  | 2.2723 | 9.6563 |  | 487.71 |  |
| Biomass (mg/m3) | TATS | 7/18/95 | 2 |  |  | 72.993 | 26.7999 |  | 0.1037 |  |  | 1.0364 | 1.312 | 1144.59 | 178.24 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TaTS | 7/28/95 | 1 | 13:00 |  | 2304 | 2432 |  | 7424.0000 |  |  | 4992 | 11775.7886 |  | 28927.4399 |  |
| Density (\#/m ${ }^{3}$ ) | TATS | 7/28/95 | 2 |  |  | 1920 | 1216 |  | 2304.0000 |  |  | 2624 | 1600 | 1.28 | 9665.28 | LGB $=900$ |
| Mean size (mm) | TATS | 7/28/95 | 1 | 13:00 |  | 0.5775 | 0.6822 |  | 0.8621 |  |  | 0.1652 | 0.129 |  | 0.4056 |  |
| Mean size (mm) | TATS | 7/28/95 | 2 |  |  | 0.5727 | 0.696 |  | 0.7662 |  |  | 0.1594 | 0.1405 | 6.4081 | 0.4514 |  |
| Biomass (mg/m3) | TATS | 9/19/95 | 1 |  |  | 26.8365 | 55.1778 |  | 167.7400 |  |  | 2.926 | 1.207 |  | 253.89 |  |
| Biomass (mg/m3) | TATS | 9/19/95 | 2 |  |  | 24.7829 | 110.41 |  | 54.4639 |  |  | 1.6699 | 1.0496 |  | 192.37 |  |
| Density ( $\left(1 / \mathrm{m}^{3}\right.$ ) | TATS | 9/19/95 | 1 |  |  | 640 | 768 |  | 2560.0000 |  |  | 5568 | 1472 |  | 11007.7866 |  |
| Density (\#/m) | TATS | 9/19/95 | 2 |  |  | 576 | 1664 |  | 1792.0000 |  |  | 2944 | 1280 |  | 8256 |  |
| Mean size (mm) | TATS | 9/19/95 | 1 |  |  | 0.5983 | 1.092 |  | 1.0000 |  |  | 0.1759 | 0.128 |  | 0.4496 |  |
| Mean size (mm) | TATS | 9/19/95 | 2 |  |  | 0.6073 | 1.0625 |  | 0.7390 |  |  | 0.1823 | 0.1297 |  | 0.502 |  |
| Biomass (mg/m3) | TLTN | 6/18/85 |  | 10:20 | 1.6206 | 17.0669 | 7.6542 |  | 313.7900 |  | 169.04 | 17.9679 | 5.1168 |  | 532.26 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 6/18/85 |  | 10:20 | 160 | 480 | 96 |  | 7968.0000 |  | 3872 | 19839.9238 | 6240 |  | 38655.3847 |  |
| Mean size (mm) | TLTN | 6/18/85 |  | 10:20 | 0.5809 | 0.5602 | 1.1456 |  | 0.8237 |  | 0.9103 | 0.216 | 0.1073 |  | 0.4014 |  |
| Biomass (mg/m3) | TLTN | 7/16/85 |  | 10:40 | 1.1442 | 26.643 | 48.6846 |  | 250.8500 |  | 178.58 | 13.305 | 32.144 |  | 551.35 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/16/85 |  | 10:40 | 160 | 480 | 880 |  | 5040.0000 |  | 2240 | 13919.704 | 39199.4524 |  | 61919.8649 |  |
| Mean size (mm) | TLTN | 7/16/85 |  | 10:40 | 0.5175 | 0.684 | 0.8993 |  | 0.9039 |  | 1.1392 | 0.2208 | 0.1085 |  | 0.2525 |  |
| Biomass (mg/m3) | TLIN | 8/13/85 |  | 14:15 | 1.1442 | 13.7906 | 35.5531 |  | 63.5492 |  | 14.304 | 1.5427 | 49.5936 |  | 179.48 |  |
| Density (\#/m $\mathrm{m}^{3}$ ) | TLTN | 8/13/85 |  | 14:15 | 160 | 320 | 1178.67 |  | 2218.6700 |  | 181.33 | 1920 | 60479.7518 |  | 66458.4182 |  |
| Mean size (mm) | TLTN | 8/13/85 |  | 14:15 | 0.5175 | 0.6135 | 0.8213 |  | 0.7179 |  | 1.1316 | 0.2013 | 0.0871 |  | 0.1309 |  |
| Biomass (mg/m3) | TLTN | 9/10/85 |  | 12:30 |  | 28.0633 | 271.1 |  | 275.5700 |  | 258.88 | 29.3848 | 61.0867 |  | 924.09 |  |
| Density (\#/mi ${ }^{3}$ ) | TLTN | 9/10/85 |  | 12:30 |  | 512 | 3456 |  | 6976.0000 |  | 3008 | 33302.7425 | 74495.4932 |  | 121727.864 |  |
| Mean size (mm) | TLTN | 9/10/85 |  | 12:30 |  | 0.6877 | I. 1251 |  | 0.7974 |  | 1.1643 | 0.2116 | 0.075 |  | 0.2131 |  |
| Biomass (mg/m3) | TLTN | 8/26/87 |  |  |  | 155.9 | 606.71 |  | 647.6400 |  | 645.25 | 33.9519 | 57.5315 |  | 2146.98 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/26/87 |  |  |  | 5120 | 11788.268 |  | 14843.6259 |  | 10018.6683 | 63998.5302 | 70160.0678 |  | 175930.039 |  |
| Mean size (mm) | TLTN | 8/26/87 |  |  |  | 0.5162 | 0.9414 |  | 0.8477 |  | 1.0452 | 0.1766 | 0.1215 |  | 0.3219 |  |
| Biomass (mg/m3) | TLTN | 8/28/88 | 1 |  |  | 28.8155 | 204.86 |  | 115.3900 |  | 222.66 | 6.447 | 1.7318 | 3.2136 | 574.9164 |  |
| Biomass (mg/m3) | TLTN | 8/28/88 | 2 |  |  | 32.3784 | 151.25 |  | 95.2246 |  | 136.6 | 9.7348 | 1.2595 | 6.4995 | 425.4505 |  |
| Density (\#/m ${ }^{3}$ ) | TLIN | 8/28/88 | 1 |  |  | 704 | 5120 |  | 5632.0000 |  | 4736 | 6912 | 2112 | 1.28 | 25216.8015 | LGB=970 |
| Density (\#/m) | TLTN | 8/28/88 | 2 |  |  | 1152 | 4096 |  | 5312.0000 |  | 3200 | 7680 | 1536 | 1.92 | 22977.8823 | LGB $=970$ |
| Mean size (mm) | TLTN | 8/28/88 | 1 |  |  | 0.5473 | 0.8798 |  | 0.6192 |  | 0.9314 | 0.2204 | 0.2653 | 0.8902 | 0.5898 |  |
| Mean size (mm) | TL. N | 8/28/88 | 2 |  |  | 0.5036 | 0.8599 |  | 0.5942 |  | 0.8898 | 0.2386 | 0.2943 | 0.9758 | 0.5393 |  |
| Biomass (mg/m3) | TLTN | 6/8/89. | 1 | 14:10 |  | 0.1935 | 8.5337 |  | 89.2521 |  | 54.039 | 4.7418 | 0.6822 | 50.1733 | 157.4467 |  |
| Biomass (mg/m3) | TLTN | 6/8/89 | 2 | 13:30 |  | 0.7249 | 66.4683 |  | 637.4500 |  | 471.35 | 15.1796 | 8.3968 |  | 1199.57 |  |
| Density (\#/m) | TLTN | 6/8/89 | 1 | 14:10 |  | 8.64 | 192 |  | 3712.0000 |  | 2176 | 12863.8213 | 832 | 0.32 | 19784.7242 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | TLIN | 6/8/89 | 2 | 13:30 |  | 24 | 1706.67 |  | 20053.0298 |  | 11733.2953 | 27733.201 | 10239.7652 |  | 71489.128 |  |
| Mean size (mm) | TLTN | 6/8/89 | 1 | 14:10 |  | 0.4445 | 0.935 |  | 0.6859 |  | 0.7326 | 0.1546 | 0.1023 | 3.9215 | 0.3234 |  |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | Tor | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean size (mm) | TLTN | 6/8/89 | 2 | 13:30 |  | 0.513 | 0.8082 |  | 0.7498 |  | 0.8749 | 0.1788 | 0.1143 |  | 0.4591 |  |
| Biomass (mg/m3) | TLTN | 7/13/89 | 1 | 12:40 |  | 4.8205 | 165.44 |  | 242.3800 |  | 328.41 | 3.7205 | 6.9274 |  | 751.7 |  |
| Biomass (mg/m3) | TLTN | 7/13/89 | 2 | 13:00 |  | 24.6277 | 29.41 |  | 250.2600 |  | 250.07 | 7.7937 | 15.674 |  | 846.83 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/13/89 | 1 | 12:40 |  | 184.32 | 4480 |  | 8192.0000 |  | 7552 | 11263.9315 | 8448 |  | 40119.9146 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/13/89 | 2 | 13:00 |  | 877.53 | 9386.67 |  | 7509.3300 |  | 3413.33 | 17919.6697 | 19114.3705 |  | 58221.0455 |  |
| Mean size (mm) | TLTN | 7/13/89 | 1 | 12:40 |  | 0.4803 | 0.8229 |  | 0.7574 |  | 0.9015 | 0.1488 | 0.0921 |  | 0.4796 |  |
| Mean sizc (mm) | TLTN | 7/13/89 | 2 | 13:00 |  | 0.4969 | 0.7823 |  | 0.7879 |  | 1.0965 | 0.1633 | 0.1231 |  | 0.3902 |  |
| Biomass (mg/m3) | TLTN | 8/30/89 | 1 | 14:30 |  | 7.2334 | 265.67 |  | 178.5400 |  | 100.86 | 9.7762 | 5.5979 |  | 567.68 |  |
| Biomass (mg/m3) | TLTN | 8/30/89 | 2 | 14:50 |  | 85.5732 | 293.38 |  | 104.2100 |  | 73.0852 | 8.3548 | 6.4142 | 2.6516 | 271.0184 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/30/89 | 1 | 14:30 |  | 261.12 | 6485.33 |  | 6314,6700 |  | 2560 | 26112.593 | 6826.67 |  | 48559.0297 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/30/89 | 2 | 14:50 |  | 2560 | 6400 |  | 3555.5600 |  | 1280 | 22897.6019 | 7822.22 | 2.1333 | 44517.42 1 3 | ****NOTE:time is 15:50 on the fieid : |
| Mean size (mm) | TLTN | 8/30/89 | 1 | 14:30 |  | 0.4962 | 0.884 |  | 0.6982 |  | 0.8793 | 0.1557 | 0.0992 |  | 0.3556 |  |
| Mean size (mm) | TLTN | 8/30/89 | 2 | 14:50 |  | 0.5349 | 0.9091 |  | 0.7093 |  | 0.9999 | 0.1539 | 0.1383 | 0.7294 | 0.3503 |  |
| Biomass (mg/m3) | TLTN | 10/4/89 | 1 | 16:30 |  | 72.6202 | 284.71 |  | 118.8600 |  | 47.8606 | 7.9167 | 4.6182 | 3.3452 | 536.5948 |  |
| Biomass (mg/m3) | TLTN | 10/4/89 | 2 | 16:45 |  | 75.9906 | 658.94 |  | 186.9500 |  | 38.5027 | 2.5952 | 2.599 | 2.3754 | 967.95 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | TLTN | 10/4/89 | 1 | 16:30 |  | 2560 | 5120 |  | 11775.7886 |  | 1024 | 37375.7572 | 5632 | 1.28 | 63489.2212 | LGB=ARANEIDA |
| Density (\#/m ${ }^{3}$ ) | TLTN | 10/4/89 | 2 | 16:45 |  | 2438.1 | 10239.7652 |  | 11946.4811 |  | 609.52 | 8045.71 | 3169.52 | 1.0667 | 36450.2071 | LGB=ARANEIDA |
| Mean size (mm) | TLTN | 10/4/89 | 1 | 16:30 |  | 0.5043 | 1.008 |  | 0.4678 |  | 0.9281 | 0.1264 | 0.1016 | 0.9858 | 0.2868 |  |
| Mean size (mm) | TLTN | 10/4/89 | 2 | 16:45 |  | 0.5231 | 1.0635 |  | 0.5576 |  | 1.0504 | 0.1472 | 0.1213 | 0.9497 | 0.5771 |  |
| Biomass (mg/m3) | TLTN | 6/10/90 |  |  |  | 2.2171 | 6.3491 |  | 395.2700 |  | 154.59 | 20.1748 | 0.2099 |  | 578.81 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | TLTN | 6/10/90 |  |  |  | 69.12 | 166.4 |  | 13311.8986 |  | 3754.67 | 22100.9021 | 256 |  | 39658.8393 |  |
| Mean size (mm) | TLTN | 6/10/90 |  |  |  | 0.5264 | 0.8651 |  | 0.7759 |  | 0.8934 | 0.2217 | 0.1017 |  | 0.4737 |  |
| Biomass (mg/m3) | TLTN | 7/17/90 |  |  |  | 32.2025 | 171.99 |  | 171.5700 |  | 137.82 | 16.7763 | 12.0704 | 2.6491 | 542.4209 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 7/17/90 |  |  |  | 1152 | 6656 |  | 5248.0000 |  | 2432 | 16511.632 | 14719.7353 | 0.96 | 46720.7709 | LGB $=$ ACARINA |
| Meansize (mm) | TLTN | 7/17/90 |  |  |  | 0.5003 | 0.7671 |  | 0.7906 |  | 1.0063 | 0.2304 | 0.1056 | 0.884 | 0.3775 |  |
| Biomass (mg/m3) | TLTN | 8/25/90 | 1 |  |  | 47.9986 | 306.29 |  | 247,5800 |  | 78.5976 | 40.2323 | 4.6671 | 4.6671 | 747.9929 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 8/25/90 | 1 |  |  | 1792 | 9984 |  | 8704.0000 |  | 1536 | 41215.4456 | 33279.7458 | 0.64 | 96511.7077 | LGB=ACARNA |
| Meant size (mm) | TLTN | 8/25/90 | 1 |  |  | 0.4915 | 0.8095 |  | 0.7244 |  | 0.9724 | 0.2265 | 1.4098 | 1.4098 | 0.3137 |  |
| Biomass (mg/m3) | TLTN | 10/1/90 | 1 |  |  | 41.9183 | 81.4373 |  | 200.2000 |  | 34.1286 | 6.7595 | 3.4987 | 6.1019 | 367.9381 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 10/1/90 | 1 |  |  | 1621.33 | 1962.67 |  | 14933.1009 |  | 682.67 | 7594.67 | 4266.67 | 0.64 | 31061.3266 | LGB=ACARINA |
| Mean size (mm) | TLTN | 10/1/90 | 1 |  |  | 0.4828 | 0.9077 |  | 0.5534 |  | 0.9516 | 0.2482 | 0.1227 | 1.5154 | 0.4398 |  |
| Biomass (mg/m3) | TLTN | 6/24/91 | 1 | 10:00 |  | 20.7607 | 17.5667 |  | 251.6400 |  | 88.4477 | 16.0968 | 9.7963 | 14.8541 | 419.16 |  |
| Biomass (mg/m3) | TLTN | 6/24/91 | 2 | 12:15 |  | 28.4574 | 6.3748 |  | 458.8600 |  | 253.22 | 9.7316 | 11.4756 |  | 768.12 |  |
| Density (\#/m) | TLTN | 6/24/91 | 1 | 10:00 |  | 711.11 | 477.87 |  | 9102.2200 |  | 1280 | 13498.9504 | 11946.4811 | 4.2667 | 49406.0348 | LGB $=$ ARANEIDA |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 6/24/91 | 2 | 12:15 |  | 1194.67 | 187.23 |  | 4.2189 |  | 3754.67 | 25599.4126 | 13994.5842 |  | 61285.8198 |  |
| Mean size (mm) | TLTN | 6/24/91 | 1 | 10:00 |  | 0.4959 | 0.8265 |  | 0.6965 |  | 1.038 | 0.19 | 0.123 | 1. 1022 | 0.2997 |  |
| Mean size (mm) | TLTN | 6/24/91 | 2 | 12:15 |  | 0.4616 | 0.8132 |  | 0.7145 |  | 1.0495 | 0.1577 | 0.1163 |  | 0.3612 |  |
| Biomass (mg/m3) | TLTN | 7/12/91 | 1 |  |  | 67.5008 | 140.11 |  | 181.6500 |  | 134,41 | 9.8672 | 4.4608 |  | 538 |  |
| Biomass (mg/m3) | TLTN | 7/12/91 | 2 |  |  | 5.4126 | 7.9296 |  | 54.3531 |  | 60.1331 | 2.0788 | 1.5114 |  | 131.42 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TL.TN | 7/12/91 | 1 |  |  | 2453.33 | 3733.33 |  | 5013.3300 |  | 2026.67 | 18666.3756 | 5440 |  | 37332.7515 |  |
| Density (\#/m $\mathrm{m}^{3}$ ) | TLTN | 7/12/91 | 2 |  |  | 256 | 160 |  | 1996.8000 |  | 819.2 | 3609.6 | 1843.2 |  | 8684.8 |  |
| Mean size (mm) | TLTN | 7/12/91 | 1 |  |  | 0.4912 | 0.8135 |  | 0.7761 |  | 1.0373 | 0.178 | 0.1206 |  | 0.3807 |  |
| Mean size (mm) | TLTN | 7/12/91 | 2 |  |  | 0.4334 | 0.9306 |  | 0.7147 |  | 1.0789 | 0.1831 | 0.1261 |  | 0.3989 |  |
| Biomass (mg/m3) | TLTN | 8/18/91 | 1 | 20:00 |  | 38.7578 | 65.5066 |  | 82.1760 |  | 19.9369 | 1.424 | 12.1287 |  | 219.93 |  |
| Biomass (mg/m3) | TLTN | 8/18/91 | 2 |  |  | 30.9726 | 540.88 |  | 241.3400 |  | 68.199 | 16.651 | 17.6333 |  | 915.68 |  |
| Biomass (mg/m3) | TLTN | 8/18/91 | 2 |  |  | 41.2969 | 540.88 |  | 241.3400 |  | 90.932 | 16.651 | 17.6333 |  | 915.68 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/18/91 | 1 | 20:00 |  | 1493.33 | 2062.22 |  | 3413.3300 |  | 392.73 | 2488.89 | 14791.0839 |  | 24641.0948 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 8/18/91 | 2 |  |  | 864 | 9386.67 |  | 7680.0000 |  | 944 | 30549.2111 | 21503.5419 |  | 70926.7402 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/18/91 | 2 |  |  | 1152 | 9386.67 |  | 7680.0000 |  | 1258.67 | 30549.2111 | 21503.5419 |  | 70926.7402 |  |
| Mean size (mm) | TLTN | 8/18/91 | 1 | 20:00 |  | 0.4788 | 0.7943 |  | 0.6563 |  | 0.9251 | 0.1727 | 0.2242 |  | 0.3532 |  |
| Mean size ( mm ) | TLTN | 8/18/91 | 2 |  |  | 0.5563 | 0.9974 |  | 0.7383 |  | 1.0791 | 0.1743 | 0.133 |  | 0.3474 |  |
| Mean size (mm) | TLTN | 8/18/91 | 2 |  |  | 0.5563 | 0.9974 |  | 0.7283 |  | 1.0791 | 0.1743 | 0.133 |  | 0.3474 |  |

Appendix 4 cont'd

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | Tor | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass (mg/m3) | TLTN | 9/11/91 | 1 |  | ! | 44.0711 | 130 |  | 269.1500 |  | 94.221 | 16.3773 | 2.3091 |  | 556.12 |  |
| Biomass (mg/m3) | TLTN | 9/11/91 | 2 |  | 0.0122 | 33.7819 | 106.67 |  | 105.8100 |  | 22.1165 | 1.5882 | 1.1895 |  | 271.17 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 9/11/91 | 1 |  |  | 1280 | 2048 |  | 9856.0000 |  | 1267.2 | 33279.7458 | 2816 |  | 50546.3732 |  |
| Density ( ${\text { / } / \mathrm{m}^{3} \text { ) }}^{\text {d }}$ | TLTN | 9/11/91 | 2 |  | 2.56 | 1152 | 2436 |  | 5504.0000 |  | 268.8 | 4565.33 | 1450.67 |  | 15375.1726 |  |
| Mean size (mm) | KLIN | 9/11/91 | 1 |  |  | 0.5468 | 9856 |  | 0.6654 |  | 1.1022 | 0.1701 | 0.131 |  | 0.3325 |  |
| Mean size (mm) | TLTN | 9/11/91 | 2 |  | 0.4517 | 0.5106 | 0.9071 |  | 0.6114 |  | 1.1476 | 0.1498 | 0.3388 |  | 0.4972 |  |
| Biomass (mg m 3 ) | TLTN | 6/23/92 | 1 |  |  | 184.13 | 53.5402 |  | 370.6900 |  | 661.6 | 9.5722 | 4.8981 |  | 1284.43 |  |
| Biomass (mg/m3) | TLTN | 6/23/92 | 2 |  |  | 21.0119 | 20.4616 |  | 63.3538 |  | 142.98 | 1.2248 | 1.8893 |  | 250.92 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 6/23/92 | 1 |  |  | 7253.33 | 2773.33 |  | 9386.6700 |  | 7253.33 | 23892.9624 | 5973.33 |  | 1064.7523 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 6/23/92 | 2 |  |  | 621.71 | 621.71 |  | 1901.7100 |  | 1718.86 | 2889.14 | 2304 |  | 10057.0359 |  |
| Mean size (mm) | TLTN | 6/23/92 | 1 |  |  | 0.4708 | 0.6951 |  | 0.8454 |  | 1.1815 | 0.1586 | 0.0968 |  | 0.4637 |  |
| Mean size (mm) | TLTN | 6/23/92 | 2 |  |  | 0.5343 | 0.8006 |  | 0.7993 |  | 1.1516 | 0.1647 | 0.1021 |  | 0.5012 |  |
| Biomass (mg/m3) | TLTN | 7/29/92 | 1 |  |  | 100.26 | 366.8 |  | 72.9004 |  | 14.0238 | 3.1382 | 0.6717 |  | 557.79 |  |
| Biomass (mg/m3) | TLTN | 7/29/92 | 2 |  |  | 102.19 | 324.5 |  | 140.3800 |  | 222.7 | 3.8561 | 4.8981 |  | 798.53 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/29/92 | 1 |  |  | 3379.2 | 8192 |  | 2764.8000 |  | 230.03 | 5120 | 819.2 |  | 20505.0107 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/29/92 | 2 |  |  | 3413.33 | 8533.33 |  | 4949.3300 |  | 4266.67 | 8362.67 | 5973.33 |  | 35498.4998 |  |
| Mean size (mm) | TLTN | 7/29/92 | 1 |  |  | 0.5042 | 0.9115 |  | 0.7220 |  | 1.0083 | 0.1877 | 0.1053 |  | 0.607 |  |
| Mean size (mm) | TLIN | 7/29/92 | 2 |  |  | 0.509 | 0.8277 |  | 0.7127 |  | 0.9217 | 0.1659 | 0.0969 |  | 0.5134 |  |
| Biomass (mg/m3) | TLTN | 8/20/92 | 2 |  |  | 169.85 | 737.76 |  | 151.3100 |  | 180.42 | 13.6541 | 5.6678 |  | 1258.66 |  |
| Density (\#/m) | TLTN | 8/20/92 | 2 |  |  | 5376 | 15615.6476 |  | 5120.0000 |  | 2560 | 36607.494 | 6912 |  | 72190.4917 |  |
| Mean size (mm) | TLTN | 8/20/92 | 2 |  |  | 0.5234 | 0.9274 |  | 0.7330 |  | 1.0803 | 0.1522 | 0.099 |  | 0.4165 |  |
| Biomass (mg/m3) | TLTN | 8/21/92 | 1 |  |  | 47.3896 | 399.24 |  | 126.0200 |  | 91.3971 | 1.1639 | 1.244 |  | 666.46 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | TLTN | 8/21/92 | 1 |  |  | 2180.74 | 9671.11 |  | 3223.7000 |  | 1706.67 | 6352.59 | 1517.04 |  | 24651.3098 |  |
| Mean size (mm) | TLTN | 8/21/92 | 1 |  |  | 0.4329 | 0.8865 |  | 0.7994 |  | 0.9521 | 0.1206 | 0.1207 |  | 0.5943 |  |
| Biomass (mg/m3) | TLTN | 10/3/92 | 1 |  |  | 16.7349 | 84.6621 |  | 83.5834 |  | 9.6899 | 1.4887 | 0.933 |  | 197.09 |  |
| Biomass (mg/m3) | TLTN | 10/3/92 | 2 |  |  | 38.6586 | 134.2 |  | 56.6878 |  | 84.7129 | 1.1469 | 1.1196 |  | 316.53 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 10/3/92 | 1 |  |  | 568.89 | 2133.33 |  | 8626.6700 |  | 247.07 | 5973.33 | 1137.78 |  | 16886.9027 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TLTN | 10/3/92 | 2 |  |  | 1706.67 | 3072 |  | 5034.6700 |  | 1706.67 | 4522.67 | 1365.33 |  | 17407.6441 |  |
| Mean size (mrn) | TLTN | 10/3/92 | 1 |  |  | 0.512 | 0.8567 |  | 0.5341 |  | 0.8342 | 0.1361 | 0.1046 |  | 0.4088 |  |
| Mean size (mm) | TLTN | 10/3/92 | 2 |  |  | 0.4526 | 0.9239 |  | 0.5161 |  | 0.9286 | 0.1352 | 0.1225 |  | 0.4925 |  |
| Biomass (mg/m3) | TLTN | 6/17/93 | 1 | 12:00 |  | 38.5713 | 67.7575 |  | 128.2200 |  | 109.85 | 1.0364 | 1.3995 |  | 346.83 |  |
| Biomass (mg/m3) | TLTN | 6/17/93 | 2 | 13:00 |  | 17.7111 | 99.799 |  | 370.6200 |  | 165.41 | 5.2103 | 5.6678 | 98.9948 | 664.4152 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 6/17/93 | 1 | 12:00 |  | 2005.33 | 2090.67 |  | 4864.0000 |  | 1834.67 | 1706.67 | 1706.67 |  | 14207.9035 |  |
| Density (\#/ ${ }^{3}$ ) | TLTN | 6/17/93 | 2 | 13:00 |  | 1152 | 2816 |  | 16895.8482 |  | 3072 | 11903.92 | 6912 | 1.28 | 42752.3508 | LGB $=900$ |
| Mean size (mm) | TLTN | 6/17/93 | 1 | 12:00 |  | 0.4105 | 0.7933 |  | 0.7203 |  | 1 | 0.1876 | 0.1154 |  | 0.8568 |  |
| Mean size (mm) | TLTN | 6/17/93 | 2 | 13:00 |  | 0.3762 | 0.8193 |  | 0.6379 |  | 0.9526 | 0.1654 | 0.1094 | 3.0127 | 0.4485 |  |
| Biomass (mg/m3) | TLTN | 8/3/93 | 1 | 19:30 |  | 165.64 | 473.21 |  | 335.1200 |  | 205.76 | 3.9082 | 13.4349 |  | 1197.07 |  |
| Biomass (mg/m3) | TLTN | 8/3/93 | 2 |  |  | 68.442 | 124.8 |  | 94.0042 |  | 97.298 | 4.8551 | 3.0905 |  | 392.49 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/3/93 | 1 | 19:30 |  | 5632 | 9216 |  | 11946.4811 |  | 3925,33 | 5802.67 | 16383.625 |  | 52906.6179 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 8/3/93 | 2 |  |  | 2702.22 | 3840 |  | 2773.3300 |  | 2560 | 9600 | 3768.89 |  | 25244.1063 |  |
| Mean size (mm) | TLTN | 8/3/93 | 1 | 19:30 |  | 0.5119 | 0.9647 |  | 0.7253 |  | 0.9304 | 0.1897 | 0.1493 |  | 0.5224 |  |
| Mean size ( mm ) | TLTN | 8/3/93 | 2 |  |  | 0.4739 | 0.8319 |  | 0.7833 |  | 0.8062 | 0.1703 | 0.2383 |  | 0.4454 |  |
| Biomass (mg/m3) | TLTN | 9/19/93 | 1 |  |  | 6.1164 | 324.37 |  | 255.4700 |  | 8.6381 | 8.0413 | 3.7786 |  | 606.41 |  |
| Biomass (mg/m3) | TLTN | 9/19/93 | 2 |  |  | 12.9729 | 136.66 |  | 59.7758 |  | 47.295 | 2.8419 | 2.1342 |  | 261.68 |  |
| Density (\#/m) | TLTN | 9/19/93 | 1 |  |  | 256 | 8192 |  | 11007.7866 |  | 130.56 | 18431.8955 | 4608 |  | 42626.5319 |  |
| Density (\#/m) | TLTN | 9/19/93 | 2 |  |  | 682.67 | 3669.33 |  | 2858.6700 |  | 810.67 | 4864 | 2602.67 |  | 15487.8096 |  |
| Mean size (mm) | TLTN | 9/19/93 | 1 |  |  | 0.47 | 0.8865 |  | 0.6359 |  | 1.0629 | 0.1674 | 0.1472 |  | 0.429 |  |
| Mcan size (mm) | TLTN | 9/19/93. | 2 |  |  | 0.416 | 2858.67 |  | 0.6278 |  | 1.0041 | 0.1824 | 0.1871 |  | 0.4811 |  |
| Biomass (mg/m3) | TLTN | 6/17/94 | 1 |  |  | 38.2352 | 71.5169 |  | 251.3400 |  | 154.96 | 0.9236 | 1.6181 |  | 518.59 |  |
| Biomass (mg/m3) | TLTN | 6/17/94 | 2 | 13:00 |  | 38.131 | 147.75 |  | 742.8600 |  | 653.82 | 2.2319 | 3.4637 | 568.29 | 1588.27 |  |
| Density (\#/m') | TLTN | 6/17/94 | 1 |  |  | 1120 | 1120 |  | 4480.0000 |  | 1706.67 | 1600 | 1973.33 |  | 11999.9656 |  |
| Density ( $/ / \mathrm{m}^{3}$ ) | TLTN | 6/17/94 | 2 | 13:00 |  | 1280 | 2304 |  | 13183.7816 |  | 6784 | 4480 | 4224 | 2.56 | 32258.1903 | LGB $=900$ |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean size (mm) | TLTN | 6/17/94 | 1 |  |  | 0.5346 | 0.9962 |  | 0.9331 |  | 1.1755 | 0.1839 | 0.1178 |  | 0.7023 |  |
| Mean size (mm) | TLTN | 6/17/94 | 2 | 13:00 |  | 0.5038 | 0.9584 |  | 0.9421 |  | 1.2133 | 0.1761 | 0.1209 | 4.4034 | 0.7693 |  |
| Biomass (m/m3) | TLTN | 7/27/94 | 1 | 17:00 |  | 175.24 | 572.29 |  | 73.3258 |  | 174.7 | 4.3329 | 8.0469 |  | 1007.94 |  |
| Biomass (mg/m3) | TLTN | 7/27/94 | 2 | 15:30 |  | 196.73 | 1706.23 |  | 509.1400 |  | 1579.07 | 21.1005 | 4024.7706 |  | 4037.04 |  |
| Density (\#/m) | TLIN | 7/27/94 | 1 | 17:00 |  | 4693.33 | 9671.11 |  | 1991.1100 |  | 1280 | 9102.22 | 9813.33 |  | 36551.062 |  |
| Density (\#/m) | TLTN | 7/27/94 | 2 | 15:30 |  | 4608 | 18943.9265 |  | 6656.0000 |  | 13311.8986 | 33791.6968 | 30207.8628 |  | 107517.709 |  |
| Mean size (mm) | TLTN | 7/27/94 | 1 | 17:00 |  | 0.5688 | 0.9893 |  | 0.8012 |  | 1.3775 | 0.1689 | 0.102 |  | 0.4961 |  |
| Mean size (mm) | TLTN | 7/27/94 | 2 | 15:30 |  | 0.6069 | 1.1439 |  | 1.0686 |  | 1.3055 | 0.1863 | 0.0967 |  | 0.5411 |  |
| Biomass (mg/m3) | TLTN | 9/18/94 | 1 | 11:20 |  | 13.3137 | 229.51 |  | 113.9100 |  | 63.6442 | 0.7457 | 1.6794 |  | 422.81 |  |
| Biomass (mg/m3) | TLTN | 9/18/94 | 2 | 13:15 |  | 28.489 | 821.23 |  | 303.5800 |  | 461.84 | 3.4159 | 1.2595 |  | 1619.76 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | TLTN | 9/18/94 | 1 | 11:20 |  | 614.4 | 3020.8 |  | 6195.2000 |  | 665.6 | 1587.2 | 2048 |  | 14130,9058 |  |
| Density ( $\mathrm{\#} / \mathrm{m}^{3}$ ) | TLTN | 9/18/94 | 2 | 13:15 |  | 1024 | 10751.7709 |  | 11434.5743 |  | 4437.33 | 5632 | 1536 |  | 34815.2886 |  |
| Mean size (mm) | TLTN | 9/18/94 | 1 | 11:20 |  | 0.4442 | 1.0949 |  | 0.5841 |  | 1.1982 | 0.1665 | 0.1962 |  | 0.613 |  |
| Mean size (mm) | TLTN | 9/18/94 | 2 | 13:15 |  | 0.4869 | 1.0807 |  | 0.6473 |  | 1.2577 | 0.1823 | 0.2445 |  | 0.7612 |  |
| Biomass (mg/m3) | TLTN | 7/30/95 | 1 | 11:00 |  | 195.15 | 974.86 |  | 256.6300 |  | 79.1832 | 12.5031 | 8.1636 |  | 1535.49 |  |
| Biomass (mg/m3) | TLTN | 7/30/95 | 2 | 13:30 |  | 198.36 | 1065.98 |  | 335.6200 |  | 255.28 | 43.2965 | 7.5571 |  | 1906.1 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/30/95 | 1 | 11:00 |  | 5973.33 | 19910.8591 |  | 7395.5600 |  | 940.17 | 13937.6643 | 9955.56 |  | 58112.559 |  |
| Density (\#/m ${ }^{3}$ ) | TLTN | 7/30/95 | 2 | 13:30 |  | 5376 | 21247.6058 |  | 7680.0000 |  | 3200 | 30207.8628 | 9216 |  | 76927.2132 |  |
| Mean size (mm) | TLTN | 7/30/95 | 1 | 11:00 |  | 5361 | 0.9457 |  | 0.7771 |  | 1.1456 | 0.2154 | 0.3042 |  | 0.6003 |  |
| Mean size (mm) | TLTN | 7/30/95 | 2 | 13:30 |  | 0.568 | 0.9565 |  | 0.8556 |  | 1.1374 | 0.2585 | 0.1908 |  | 0.561 |  |
| Biomass (mg/m3) | TPR | 9/3/87 |  |  |  |  |  |  | 185.2400 |  |  | 0.7945 | 19.0502 |  | 205.09 |  |
| Denisity ( $\# / \mathrm{m}^{3}$ ) | TPR | 9/3/87 |  |  |  |  |  |  | 8829.4400 |  |  | 1920 | 23231.647 |  | 33981.3008 |  |
| Mean size (mm) | TPR | 9/3/87 |  |  |  |  |  |  | 0.6393 |  |  | 0.1625 | 0.1637 |  | 0.2872 |  |
| Biomass (mg/m3) | TPR | 7/21/88 | 1 | 15:00 |  |  |  |  | 176.3900 |  | 0.0351 | 0.0611 | 0.2362 | 124.2 | 176.73 |  |
| Biomass (mg/m3) | TPR | 7/21/88 | 2 | 15:30 |  | 0.003 |  |  | 90.5404 |  |  | 0.0029 | 0.2624 |  | 90.8087 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 7/21/88 | 1 | 15:00 |  |  |  |  | 11071.8451 |  | 0.64 | 32 | 288 | 0.32 | 11392.7876 | LGB $=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 7/21/88 | 2 | 15:30 |  | 0.32 |  |  | 4912.0000 |  |  | 1.6 | 320 |  | 5233.92 |  |
| Mean size (mm) | TPR | 7/21/88 | 1 | 15:00 |  |  |  |  | 0.5997 |  | 0.9991 | 0.292 | 0.1836 | 5.3048 | 0.5885 |  |
| Mean size (mm) | TPR | 7/21/88 | 2 | 15:30 |  | 0.3062 |  |  | 0.6313 |  |  | 0.286 | 0.1855 |  | 0.6039 |  |
| Biomass (mg/m3) | TPR | 8/22/88 | 1 | 9:45 |  | 0.2092 |  |  | 217.7800 |  |  | 0.021 | 1.1808 |  | 219.18 |  |
| Biomass (mg/m3) | TPR | 8/22/88 | 2 | 9:00 |  | 9.5018 |  |  | 46.9341 |  |  |  | 2.073 | 47.4311 | 58.5089 |  |
| Density (\#/m²) | TPR | 8/22/88 | 1 | 9:45 |  | 10.56 |  |  | 4.0663 |  |  | 32 | 1440 |  | 13130.4618 |  |
| Density (\#/m²) | TPR | 8/22/88 | 2 | 9:00 |  | 720 |  |  | 3024.0000 |  |  |  | 2528 | 0.32 | 6272.32 | LGB=90S? ? ? quastion marks on datas |
| Mean size (mm) | TPR | 8/22/88 | 1 | 9:45 |  | 0.4144 |  |  | 0.6545 |  |  | 0.1976 | 0.2045 |  | 0.6039 |  |
| Mean size (mm) | TPR | 8/22/88 | 2 | 9:00 |  | 0.3546 |  |  | 0.6115 |  |  |  | 0.2769 | 3.8487 | 0.4473 |  |
| Biomass (mg/m3) | TPR | 9/24/88 | 1 | 18:00 |  | 0.019 | 0.0576 |  | 62.5461 | 0.1594 | 0.0109 | 0.003 | 2.3354 |  | 65.1314 |  |
| Biomass (mg/m3) | TPR | 9/24/88 | 2 | 17:30 |  |  | 0.0315 |  | 145.3100 |  |  | 0.0013 | 4.2246 |  | 149.57 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 9/24/88 | 1 | 18:00 |  | 0.8 | 2.72 |  | 2720.0000 | 0.32 | 0.16 | 3.04 | 2848 |  | 5575.04 | CAL $=020$ |
| Density (\#/m ${ }^{3}$ ) | TPR | 9/24/88 | 2 | 17:30 |  |  | 1.28 |  | 6272.0000 |  |  | 1.28 | 5152 |  | 11426.4152 |  |
| Mean size (mm) | TPR | 9/24/88 | 1 | 18:00 |  | 0.4475 | 0.7166 |  | 0.7121 | 2.0746 | 1.083 | 0.228 | 0.1812 |  | 0.4407 |  |
| Mean size (mm) | TPR | 9/24/88 | 2 | 17:30 |  |  | 0.7575 |  | 0.7154 |  |  | 0.2323 | 0.1769 |  | 0.4726 |  |
| Biomass (mg/m3) | TPR | 6/8/89 | 1 | 10:30 |  | 0.0887 | 0.1305 |  | 76.6787 |  |  | 2.7923 | 2.0467 | ???? | 81.88 |  |
| Biomass (mg/m3) | TPR | 6/8/89 | 2 | 10:00 |  | 0.2904 | 0.1931 |  | 47.1156 |  |  | 3.0734 | 1.9155 | 3.1747 | 52.5881 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 6/8/89 | 1 | 10:30 |  | 2.4 | 3.04 |  | 2272.0000 |  |  | 5632 | 2496 | 0.16 | 10405.4292 | $L G B=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 6/8/89 | 2 | 10.00 |  | 8.64 | 4.16 |  | 1984.0000 |  |  | 8320 | 2336 | 0.16 | 12652.8978 | LGB=900 |
| Mean size (mm) | TPR | 6/8/89 | 1 | 10:30 |  | 0.5627 | 0.9192 |  | 0.8202 |  |  | 0.1759 | 0.1145 | 2.8778 | 0.3022 |  |
| Mean size (mm) | TPR | 6/8/89 | 2 | 10:00 |  | 0.5331 | 0,9366 |  | 0.7154 |  |  | 0.1547 | 0.0958 | 1.9688 | 0.2323 |  |
| Biomass (mg/m3) | TPR | 7/13/89 | 1 | 10:50 |  |  |  |  | 221.2700 |  |  | 4.472 | 11.9654 |  | 239.9 |  |
| Biomass (mg/m3) | TPR | 7/13/89 | 2 | 10:30 |  |  |  |  | 81.7489 |  |  | 11.5508 | 2.1692 | 69.8688 | 95.4712 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 7/13/89 | 1 | 10:50 |  |  |  |  | 10367.8809 |  |  | 12799.7062 | 14591.838 |  | 37759.8274 |  |
| Density (\#/m) | TPR | 7/13/89 | 2 | 10:30 |  |  |  |  | 6741.3300 |  |  | 15445.0742 | 2645.33 | 0.64 | 24832.4746 | LGB $=900$ |
| Mean size (mm) | TPR | 7/13/89 | 1 | 10:50 |  |  |  |  | 0.6363 |  |  | 0.1774 | 0.0893 |  | 0.2694 |  |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean size (mm) | TPR | 7/13/89 | 2 | 10:30 | , |  |  |  | 0.5003 |  |  | 0.2026 | 0.1131 | 3.4757 | 0.274 |  |
| Biomass (mg/m3) | TPR | 8/30/89 | 1 | 10:00 |  | 0.0457 | 0.0311 |  | 69.5030 |  |  | 0.4401 | 6.9498 | 122.3 | 77 |  |
| Biomass (mg/m3) | TPR | 8/30/89 | 2 | 10:30 | 0.0014 | 0.1602 | 0.4866 |  | 152.1900 |  | 0.0409 | 0.7646 | 3.8205 |  | 157.47 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 8/30/89 | 1 | 10:00 |  | 2.56 | 1.28 |  | 7168.0000 |  |  | 2112 | 8512 | 0.32 | 17795.9018 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | TPR | 8/30/89 | 2 | 10:30 | 0.16 | 5.28 | 11.84 |  | 9984.0000 |  | 0.64 | 2713.6 | 4659.2 |  | 16978.9176 | DIA $=$ DPH |
| Mean size (mm) | TPR | 8/30/89 | 1 | 10:00 |  | 0.3961 | 0.7574 |  | 0.5061 |  |  | 0.1252 | 0.0793 | 5.2775 | 0.2568 |  |
| Mean size (mm) | TPR | 8/30/89 | 2 | 10:30 | 0.5582 | 0.5126 | 0.8853 |  | 0.6007 |  | 1.0181 | 0.1434 | 0.1067 |  | 0.397 |  |
| Biomass (mg/m3) | TPR | 10/4/89 | 1 | 14:30 |  | 0.0068 | 58.371 |  | 58.3710 |  |  | 0.2202 | 1.866 | 42.0208 | 60.4592 |  |
| Biomass (mg/m3) | TPR | 10/4/89 | 2 | 13:00 |  | 0.7731 |  |  | 59.0654 |  | 0.0512 | 0.8286 | 1.8106 | 67.086 | 62.524 |  |
| Density (\#/ $\mathrm{m}^{3}$ ) | TPR | 10/4/89 | 1 | 14:30 |  | 0.64 | 5233.78 |  | 5233.7800 |  |  | 938.67 | 2275.56 | 0.32 | 8448:96 | LGB $=900$ |
| Density (\#/ $\mathrm{m}^{3}$ ) | TPR | 10/4/89 | 2 | 13:00 |  | 64 |  |  | 3904.0000 |  | 0.64 | 2624 | 2208 | 0.64 | 8801.28 | LGB 9900 |
| Mean size (mm) | TPR | 10/4/89 | 1 | 14:30 |  | 0.3252 | 0.5424 |  | 0.5424 |  |  | 0.1299 | 0.0859 | 3.6964 | 0.3737 |  |
| Mean size (mm) | TPR | 10/4/89 | 2 | 13:00 |  | 0.3348 |  |  | 0.6087 |  | 1.1461 | 0.1481 | 0.1138 | 3.3802 | 0.3455 |  |
| Biomass (mg/m3) | TPR | 6/8/90 | 1 | 12:50 |  |  |  |  | 109.0200 |  |  | 0.6289 | 0.0175 |  | 109.67 |  |
| Biomass (mg/m3) | TPR | 6/8/90 | 2 |  |  | 0.0052 |  |  | 18.6672 |  | 0.0125 | 0.2592 | 0.0656 |  | 19.0097 |  |
| Density (\#/ $\mathrm{m}^{3}$ ) | TPR | 6/8/90 | 1 | 12:50 |  |  |  |  | 5162.6700 |  |  | 1493.33 | 21.3333 |  | 6677.33 |  |
| Density ( $\# / / \mathrm{m}^{3}$ ) | TPR | 6/8/90 | 2 |  |  | 0.32 |  |  | 933.3300 |  | 0.32 | 506.67 | 30 |  | 1520 |  |
| Mean size (mm) | TPR | 6/8/90 | 1 | 12:50 |  |  |  |  | 0.6935 |  |  | 0.165 | 0.0788 |  | 0.5733 |  |
| Mean size (mm) | TPR | 6/8/90 | 2 |  |  | 0.393 |  |  | 0.6752 |  | 0.882 | 0.1755 | 0.0928 |  | 0.4781 |  |
| Biomass (mg/m3) | TPR | 7/17/90 | 1 | 12:50 |  | 0.0275 |  |  | 190.1800 |  |  | 3.2606 | 0.2099 | 67.6463 | 193.6837 |  |
| Biomass (mg/m3) | TPR | 7/17/90 | 2 | 12:20 |  | 0.2527 | 0.0583 |  | 41.8332 |  |  | 2.6562 | 0.3149 | 0.4308 | 45.1152 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 7/17/90 | 1 | 12:50 |  | 0.64 |  |  | 14015.8680 |  |  | 4416 | 256 | 0.32 | 18688.739 | $L G B=900$ |
| Density ( $\# / \mathrm{m}^{2}$ ) | TPR | 7/17/90 | 2 | 12:20 |  | 6.08 | 1.6 |  | 4821.3300 |  |  | 3605.33 | 384 | 0.32 | 8818.67 | LGB=ARANEIDA (SPIDER) |
| Mean size (mm) | TPR | 7/17/90 | 1 | 12:50 |  | 0.6145 |  |  | 0.5447 |  |  | 0.2034 | 0.0971 | 4.3322 | 0.458 |  |
| Mean size (mm) | TPR | 7/17/90 | 2 | 12:20 |  | 0.5959 | 0.8755 |  | 0.4675 |  |  | 0.2028 | 0.1072 | 0.798 | 0.3438 |  |
| Biomass (mg/m3) | TPR | 9/2/90 | 1 |  |  | 0.057 | 0.1083 |  | 29.2513 |  |  | 0.046 | 0.7347 |  | 30.1974 |  |
| Biomass (mg/m3) | TPR | 9/2/90 | 2 | 11:00 |  | 0.0975 | 0.2701 |  | 79.3262 |  | 0.0168 | 0.0473 | 0.5904 | 20.5266 | 80.3434 |  |
| Density (\#/m) | TPR | 9/2/90 | 1 |  |  | 0.96 | 2.4 |  | 1488.0000 |  |  | 104 | 896 |  | 2491.36 |  |
| Density (\#/m) | TPR | 9/2/90 | 2 | 11:00 |  | 2.24 | 5.92 |  | 4016.0000 |  | 0.32 | 160 | 720 | 0.16 | 4904.64 | LGB=900 |
| Mean size (mm) | TPR | 9/2/90 | 1 |  |  | 0.7079 | 0.9267 |  | 0.6720 |  |  | 0.1679 | 0.0933 |  | 0.4431 |  |
| Mean size (mm) | TPR | 9/2/90 | 2 | 11:00 |  | 0.6145 | 0.9402 |  | 0.6692 |  | 0.9831 | 0.1424 | 0.0919 | 3.6678 | 0.5677 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | TPR | 9/6/90 | 1 | 11:50 |  | 0.0388 | 0.0777 |  | 110.4000 |  |  | 0.1211 | 2.1254 |  | 112.76 |  |
| Density (\#/m) | TPR | 9/6/90 | 1 | 11:50 |  | 0.96 | 1.44 |  | 6016.0000 |  |  | 384 | 2592 |  | 8994.4 |  |
| Mean size (mm) | TPR | 9/6/90 | 1 | 11:50 |  | 0.5828 | 0.9853 |  | 0.6534 |  |  | 0.1438 | 0.087 |  | 0.4685 |  |
| Biomass (mg/m3) | TPR | 6/19/91 | 1 | 18:00 |  | 0.0158 |  |  | 121.3900 |  |  | 0.1281 | 0.1225 |  | 121.66 |  |
| Biomass (mg/m3) | TPR | 6/19/91 | 2 | 17:00 |  |  |  |  | 128.0900 |  |  | 0.7024 | 0.1924 |  | 128.98 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 6/19/91 | 1 | 18:00 |  | 0.64 |  |  | 5034.6700 |  |  | 277.33 | 149.33 |  | 5461.97 |  |
| Density (\#/m) | TPR | 6/19/91 | 2 | 17:00 |  |  |  |  | 5589.3300 |  |  | 1088 | 234.67 |  | 6912 |  |
| Mean size (mm) | TPR | 6/19/91 | 1 | 18:00 |  | 0.477 |  |  | 0.7279 |  |  | 0.165 | 0.1127 |  | 0.6825 |  |
| Mean size (mm) | TPR | 6/19/91 | 2 | 17:00 |  |  |  |  | 0.7137 |  |  | 0.1918 | 0.1041 |  | 0.6108 |  |
| Biomass (mg/m3) | TPR | 7/17/91 | 1 |  |  | 1.3598 |  |  | 70.0496 |  |  | 1.1478 | 0.3779 | 257.34 | 72.94 |  |
| Biomass (mg/m3) | TPR | 7/17/91 | 2 |  |  | 0.0142 |  |  | 34.8080 |  |  | 0.7423 | 0.4264 |  | 35.9909 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 7/17/91 | 1 |  |  | 33.7067 |  |  | 4556.8000 |  |  | 1996.8 | 460.8 | 0.4267 | 7048.53 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | TPR | 7/17/91 | 2 |  |  | 0.64 |  |  | 2216.0000 |  |  | 968 | 520 |  | 3704.64 |  |
| Mean size (mm) | TPR | 7/17/91 | 1 |  |  | 0.5875 |  |  | 0.5666 |  |  | 0.1839 | 0.1063 | 6.1445 | 0.4286 |  |
| Mean size (mm) | TPR | 7/17/91 | 2 |  |  | 0.4541 |  |  | 0.5627 |  |  | 0.204 | 0.111 |  | 0.4055 |  |
| Biomass (mg/m3) | TPR | 8/15/91 | 1 | 13:50 |  |  |  |  | 93.9171 |  |  | 0.0234 | 0.2099 |  | 94.1504 |  |
| Biomass (mg/m3) | TPR | 8/15/91. | 2 |  |  |  |  |  | 176.1500 |  |  | 0.0564 | 4.0934 |  | 180.3 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 8/15/91 | 1 | 13:50 |  |  |  |  | 6835.2000 |  |  | 76.8 | 256 |  | 7168 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 8/15/91 | 2 |  |  |  |  |  | 12607.8200 |  |  | 128 | 4992 |  | 17727.6023 |  |
| Mean size (mm) | TPR | 8/15/91 | 1 | 13:50 |  |  |  |  | 0.5687 |  |  | 0.1419 | 0.1062 |  | 0.5476 |  |
| Mean size (mm) | TPR | 8/15/91 | 2 |  |  |  |  |  | 0.5830 |  |  | 0.1544 | 0.1033 |  | 0.4448 |  |


| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass (mg/m3) | TPR | 9/18/91 | I |  |  | 2.9754 | 0.0662 |  | 150.5200 |  |  | 0.0215 | 9.8662 |  | 163.45 |  |
| Biomass (mg/m3) | TPR | 9/18/91 | 2 |  |  |  |  |  | 93.0729 |  |  | 0.3569 | 3.4112 |  | 96.8411 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 9/18/91 | , |  |  | 69.76 | 1.92 |  | 8128.0000 |  |  | 64 | 12031.7833 |  | 20295.5114 |  |
| Density ( $\left(\mathrm{m}^{3}\right)$ | TPR | 9/18/91 | 2 |  |  |  |  |  | 4896.0000 |  |  | 1216 | 4160 |  | 10271.8815 |  |
| Mear size (mm) | TPR | 9/18/91 | 1 |  |  | 0.6029 | 0.8646 |  | 0.6603 |  |  | 0.1544 | 0.1142 |  | 0.3348 |  |
| Mean size (mm) | TPR | 9/18/91 | 2 |  |  |  |  |  | 0.6636 |  |  | 0.1454 | 0.1225 |  | 0.3831 |  |
| Biomass (mg/m3) | TPR | 6/20/92 | 1 |  |  |  |  |  | 84.7574 |  |  | 0.5687 | 0.1889 |  | 85.515 |  |
| Biomass (mg/m3) | TPR | 6/20/92 | 2 |  |  |  |  |  | 129.7300 |  |  | 1.7513 | 0.1837 | 43.1504 | 131.6596 |  |
| Density (\#/m) | TPR | 6/20/92 | 1 |  |  |  |  |  | 3148.8000 |  |  | 2048 | 230.4 |  | 5427.2 |  |
| Density ( ${\mathrm{H} / \mathrm{m}^{3} \text { ) }}^{\text {d }}$ | TPR | 6/20/92 | 2 |  |  |  |  |  | 5984.0000 |  |  | 3200 | 224 | 1.6 | 9408 | LGB $=900$ |
| Mean size (mm) | TPR | 6/20/92 | 1 |  |  |  |  |  | 0.7412 |  |  | 0.1312 | 0.0673 |  | 0.4824 |  |
| Mean size (mm) | TPR | 6/20/92 | 2 |  |  |  |  |  | 0.6893 |  |  | 0.175 | 0.0959 | 1.9577 | 0.5005 |  |
| Biomass (mg/m 3 ) | TPR | 8/3/92 | 1 |  |  | 0.3439 | 0.0505 |  | 109.8900 |  |  | 0.5041 | 0.2624 |  | 111.05 |  |
| Biomass (mg/m3) | TPR | 8/3/92 | 2 |  |  | 0.0906 | 0.0545 |  | 146.0200 |  |  | 1.5778 | 0.2099 | 43.3182 | 147.9518 |  |
| Density (\#/m) | TPR | 8/3/92 | 1 |  |  | 7.68 | 1.6 |  | 12799.7062 |  |  | 1536 | 320 |  | 14665.2679 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 8/3/92 | 2 |  |  | 1.92 | 0.96 |  | 14335.7362 |  |  | 3136 | 256 | 0.32 | 17730.8681 | LGB $=900$ |
| Mean size (mm) | TPR | 8/3/92 | 1 |  |  | 0.6184 | 0.8182 |  | 0.4601 |  |  | 0.1453 | 0.0589 |  | 0.4185 |  |
| Mean size (mm) | TPR | 8/3/92 | 2 |  |  | 0.6381 | 0.9966 |  | 0.4834 |  |  | 0.1711 | 0.1271 | 3.7341 | 0.4231 |  |
| Biomass (mg/m3) | TPR | 8/23/92 | 1 |  |  | 0.0767 | 0.0657 |  | 152.0900 |  |  | 0.2389 | 0.8922 |  | 153.37 |  |
| Biomass (mg/m3) | TPR | 8/23/92 | 2 |  |  |  | 0.1905 |  | 144.6100 |  |  | 0.2963 | 1.9732 |  | 147.07 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 8/23/92 | 1 |  |  | 2.24 | 1.28 |  | 13119.8841 |  |  | 1088 | 1088 |  | 15299.245 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 8/23/92 | 2 |  |  |  | 5.12 |  | 11468.5893 |  |  | 921.6 | 2406.4 |  | 14801.6455 |  |
| Mean size (mm) | TPR | 8/23/92 | 1 |  |  | 0.5455 | 0.9477 |  | 0.5308 |  |  | 0.12 | 0.0671 |  | 0.4686 |  |
| Mean size (mm) | TPR | 8/23/92 | 2 |  |  |  | 0.8835 |  | 0.5541 |  |  | 0.1469 | 0.1148 |  | 0.4575 |  |
| Biomass (mg/m3) | TPR | 9/22/92 | 1 |  |  |  | 0.0057 |  | 101.2200 |  |  | 0.067 | 5.7728 |  | 107.07 |  |
| Biomass (mg/m3) | TPR | 9/22/92 | 2 |  |  | 0.0197 | 0.0745 |  | 133.0000 |  |  | 0.0139 | 0.4921 |  | 134.96 |  |
| Density ( $\mathrm{\#} / \mathrm{m}^{3}$ ) | TPR | 9/22/92 | 1 |  |  |  | 0.32 |  | 9792.0000 |  |  | 384 | 7040 |  | 17216.3071 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 9/22/92 | 2 |  |  | 0.64 | 0.64 |  | 9344.0000 |  |  | 0.32 | 2048 |  | 13057.499 |  |
| Mean size (mm) | TPR | 9/22/92 | 1 |  |  |  | 0.7017 |  | 0.5314 |  |  | 0.1187 | 0.0893 |  | 0.3414 |  |
| Mean size (mm) | TPR | 9/22/92 | 2 |  |  | 0.5264 | 1.1482 |  | 0.5944 |  |  | 0.9164 | 0.1193 |  | 0.4569 |  |
| Biomass (mg/m3) | TPR | 6/19/93 | 1 | 19:30 |  | 0.0207 | 0.038 |  | 242.3600 |  |  | 1.5612 | 0.105 | 168.88 | 244.09 |  |
| Biomass (mg/m3) | TPR | 6/19/93 | 2 | 20:00 |  |  | 0.0235 |  | 82.0935 |  |  | 4.0832 | 0.1574 | 87.5894 | 86.3606 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 6/19/93 | 1 | 19:30 |  | 0.64 | 1.28 |  | 7040.0000 |  |  | 1920 | 128 | 1.92 | 9091.84 | $L G B=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 6/19/93 | 2 | 20:00 |  |  | 0.64 |  | 3840.0000 |  |  | 6848 | 192 | 0.64 | 10881.0298 | LGB=900 |
| Mean size (mm) | TPR | 6/19/93 | 1 | 19:30 |  | 0.5397 | 0.8293 |  | 0.7915 |  |  | 0.2127 | 0.0983 | 3.1616 | 0.66 |  |
| Mean size (mm) | TPR | 6/19/93 | 2 | 20:00 |  |  | 0.891 |  | 0.6633 |  |  | 0.1868 | 0.1191 | 3.7477 | 0.354 |  |
| Biomass (mg/m3) | TPR | 7/31/93 | 1 | 12:00 |  |  | 0.2234 |  | 121.2000 |  |  | 0.4921 | 1.0496 | 55.1322 | 122.9678 |  |
| Biomass (mg/m3) | TPR | 7/31/93 | 2 | 11:00 |  |  | 0.0568 |  | 154.3400 |  |  | 4.662 | 0.6298 |  | 155.49 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 7/31/93 | 1 | 12:00 |  |  | 6.08 |  | 10751.7709 |  |  | 1280 | 1280 | 0.64 | 13318.6437 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | TPR | 7/31/93 | 2 | 11:00 |  |  | 0.64 |  | 9088.0000 |  |  | 896 | 768 |  | 10752.5136 |  |
| Mean size (mm) | TPR | 7/31/93 | 1 | 12:00 |  |  | 0.8602 |  | 0.5156 |  |  | 0.1575 | 0.117 | 2.8009 | 0.4431 |  |
| Mean size (mm) | TPR | 7/31/93 | 2 | 11:00 |  |  | 1.1944 |  | 0.5775 |  |  | 0.1673 | 0.1075 |  | 0.5098 |  |
| Biomass (mg/m3) | TPR | 9/11/93 | 1 |  |  |  | 0.3803 |  | 14.81 .8100 |  |  | 0.2482 | 4.8232 | 44.7892 | 177.2708 |  |
| Biomass (mg/m3) | TPR | 9/11/93 | 2 |  | 0.0025 |  | 0.7597 |  | 68.4600 |  |  | 0.4859 | 5.7728 |  | 75.4808 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 9/11/93 | , | 11:00 |  |  | 10.6667 |  | 6400.0000 |  |  | 1152 | 5888 | 0.4267 | 13450.8584 | $L G B=900$ |
| Density (\#/m²) | TPR | 9/11/93 | 2 |  | 0.64 |  | 38.4 |  | 2560.0000 |  |  | 2432 | 7040 |  | 12070.9097 | DIA $=141$ |
| Mean size (mm) | TPR | 9/11/93 | 1 |  |  |  | 0.8614 |  | 0.7533 |  |  | 0.1294 | 0.0992 | 3.4306 | 0.4137 |  |
| Mean size (mm) | TPR | 9/11/93 | 2 |  | 0.4215 |  | 0.6968 |  | 0.7404 |  |  | 0.1263 | 0.1772 |  | 0.288 |  |
| Biomass (mg/m3) | TPR | 7/24/94 | 1 | 14:00 |  |  | 1.6344 |  | 288.3500 |  |  | 0.2649 | 2.0992 |  | 292.34 |  |
| Biomass (mg/m3) | TPR | 7/24/94 | 2 | 16:00 |  |  | 1.0838 |  | 122.4800 |  |  | 1.1066 | 0.2099 |  | 124.88 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 7/24/94 | 1 | 14:00 |  |  | 66.56 |  | 17151.8117 |  |  | 256 | 2560 |  | 20034.1075 |  |
| Density (H/m ${ }^{3}$ ) | TPR | 7/24/94 | 2 | 16:00 |  |  | 34.56 |  | 12799.7062 |  |  | 1536 | 256 |  | 14626.486 |  |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean size (mm) | TPR | 7/24/94 | 1 | 14:00 | - |  | 0.7237 |  | 0.6050 |  |  | 0.2333 | 0.2276 |  | 0.5524 |  |
| Mean size (mm) | TPR | 7/24/94 | 2 | 16:00 |  |  | 0.8004 |  | 0.4918 |  |  | 0.203 | 0.1342 |  | 0.4559 |  |
| Biomass (my/m3) | TPR | 9/11/94 | 1 |  |  | 0.0389 | 0.4494 |  | 245.7800 |  |  | 2.5943 | 5.4579 | 115.86 | 254.32 |  |
| Biomass (mg/m3) | TPR | 9/11/94 | 2 |  |  | 1.6673 | 3.2825 |  | 62.9013 |  |  | 1.3783 | 2.519 |  | 71.7484 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 9/11/94 | 1 |  |  | 1.92 | 16.64 |  | 11775.7886 |  |  | 3754.67 | 6656 | 0.32 | 22205.4715 | LGB $=900$ |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 9/11/94 | 2 |  |  | 128 | 85.76 |  | 1834.6700 |  |  | 2048 | 3072 |  | 7168.43 |  |
| Mean size (mm) | TPR | 9/11/94 | 1 |  |  | 0.4322 | 0.7798 |  | 0.6674 |  |  | 0.1968 | 0.1306 | 5.1833 | 0.427 |  |
| Mean size (mm) | TPR | 9/11/94 | 2 |  |  | 0.3528 | 0.8238 |  | 0.8026 |  |  | 0.1969 | 0.1421 |  | 0.3387 |  |
| Biomass (mg/m3) | TPR | 7/26/95 | 1 |  |  |  | 0.0479 |  | 112.5600 |  |  | 3.8156 |  |  | 116.43 |  |
| Biomass (m/m3) | TPR | 7/26/95 | 2 |  |  |  | 0.0487 |  | 80.0163 |  |  | 1.4993 | 0.0109 |  | 81,5753 |  |
| Biomass (m/m3) | TPR | 7/26/95 | 3 |  |  |  | 0.0647 |  | 281.2700 |  |  | 3.3683 |  |  | 284.71 |  |
| Density (\#/m) | TPR | 7/26/95 | 4 |  |  |  | 0.0494 |  | 197.9100 |  |  | 8.204 | 0.0437 |  | 206.2 |  |
| Density (\#/m) | TPR | 7/26/95 | 1 |  |  |  | 1.4667 |  | 5573.3300 |  |  | 2826.67 |  |  | 8401.47 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TPR | 7/26/95 | 2 |  |  |  | 1.3333 |  | 3906.6700 |  |  | 1160 | 13.3333 |  | 5081.33 |  |
| Density ( $\mathrm{H} / \mathrm{m}^{3}$ ) | TPR | 7/26/95 | 3 |  |  |  | 1.6 |  | 8693.3300 |  |  | 1866.67 |  |  | 10561.3637 |  |
| Density (H/m ${ }^{3}$ ) | TPR | 7/26/95 | 4 |  |  |  | 1.0667 |  | 8053.3300 |  |  | 4586.67 | 53.3333 |  | 12694.3363 |  |
| Mean size (mm) | TPR | 7/26/95 | 1 |  |  |  | 0.8369 |  | 0.6275 |  |  | 0.2469 |  |  | 0.4995 |  |
| Mean size (mm) | TPR | 7/26/95 | 2 |  |  |  | 0.8674 |  | 0.6438 |  |  | 0.2479 | 0.3141 |  | 0.5526 |  |
| Mean size (mm) | TPR | 7/26/95 | 3 |  |  |  | 0.9192 |  | 0.7698 |  |  | 0.2825 |  |  | 0.6837 |  |
| Mean size (mm) | TPR | 7/26/95 | 4 |  |  |  | 0.9475 |  | 0.6884 |  |  | 0.2808 | 0.1714 |  | 0.539 |  |
| Biomass (mg/m3) | TPR | 9/16/95 | 1 | 13:20 |  |  | 1.2726 |  | 35.2551 |  |  | 0.1082 | 1.2923 |  | 37.9283 |  |
| Biomass (mg/m3) | TPR | 9/16/95 | 2 |  |  | 0.3943 | 5.4854 |  | 148.1200 |  |  | 0.1675 | 0.656 |  | 154.82 |  |
| Density (\#/m ${ }^{3}$ ) | TPR | 9/16/95 | 1 | 13:20 |  |  | 61.6 |  | 1160.0000 |  |  | 184 | 1576 |  | 2981.6 |  |
| Density (\#/m) | TPR | 9/16/95 | 2 |  |  | 32 | 131.84 |  | 3904.0000 |  |  | 224 | 800 |  | 5091.84 |  |
| Mean size ( mm ) | TPR | 9/16/95 | 1 | 13:20 |  |  | 0.6965 |  | 0.7840 |  |  | 0.1828 | 0.114 |  | 0.391 |  |
| Mean size (mm) | TPR | 9/16/95 | 2 |  |  | 0.3469 | 0.8976 |  | 0.8570 |  |  | 0.2064 | 0.1287 |  | 0.7118 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | TUYA | 9/9/87 | 1 |  |  |  | 17.9762 |  | 139.8600 | 110.54 | 176.21 | 6.4324 | 4.9856 |  | 456 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | TUYA | 9/9/87 | 2 |  |  |  | 18.3596 |  | 266.0700 | 903.12 | 522.86 | 3.1401 | 6.2976 |  | 1719.85 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 9/9/87 | 1 |  |  |  | 88.5333 |  | 3254.4000 | 76.8 | 4470.4 | 14879.3996 | 6080 |  | 28849.6135 | CAL=HETEROCOPE |
| Density (\#/m ${ }^{3}$ ) | TUYA | 9/9/87 | 2 |  |  |  | 40.96 |  | 5075.2000 | 691.2 | 11701.728 | 6432 | 7680 |  | 31600.9399 | CAL $=$ HETEROCOPE |
| Mean size (mm) | TUYA | 9/9/87 | , |  |  |  | 1.5615 |  | 0.8979 | 3.2957 | 0.8543 | 0.1685 | 0.1321 |  | 0.362 |  |
| Mean size (mm) | TUYA | 9/9/87 | 2 |  |  |  | 1.7366 |  | 0.9645 | 3.1497 | 0.8877 | 0.1743 | 0.1215 |  | 0.6193 |  |
| Biomass (mg/m3) | TUYA | 7/21/88 | 1 | 8:30 | 18.5126 |  |  |  | 125.5300 |  | 122.03 | 0.6314 | 0.2493 | 2071.1 | 266.96 | note:dia stuff crossed out on sheet??? |
| Biomass (mg/m3) | TUYA | 7/21/88 | 2 | 9:30 |  |  | 0.5673 | 0.0189 | 91.3080 | 31.6525 | 249.74 | 0.3514 | 1.6619 |  | 375.3 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 7/21/88 | 1 | 8:30 | 16 |  |  |  | 4016.0000 |  | 4000 | 976 | 304 | 48 | 9360 | LGB 100 |
| Density (\#/m) | TUYA | 7/21/88 | 2 | 9:30 |  |  | 2.4889 | 0.7111 | 3093.3300 | 38.7556 | 8373.33 | 480 | 2026.67 |  | 14015.2226 | CAL $=020$ |
| Mean size (mm) | TUYA | 7/21/88 | 1 | 8:30 | 2.8011 |  |  |  | 0.7831 |  | 0.7718 | 0.1956 | 0.1593 | 2.4921 | 0.709 |  |
| Mean size (mm) | TUYA | 7/21/88 | 2 | 9:30 |  |  | 1.5669 | 0.7999 | 0.7832 | 2.6535 | 0.7692 | 0.2051 | 0.1708 |  | 0.6718 |  |
| Biomass (mg/m3) | TUYA | 8/24/88 | 1 | 12:30 |  | 0.085 | 1.6945 |  | 131.3400 | 45.6372 | 115.06 | 2.4642 | 0.656 | 2552.02 | 296.94 |  |
| Biomass (mg/m3) | TUYA | 8/24/88 | 2 | 13:00 |  | 0.5778 | 1.5358 |  | 53.8672 | 107.95 | 212.86 | 3.7025 | 1.1079 |  | 381.61 |  |
| Density (\#/m) | TUYA | 8/24/88 | 1 | 12:30 |  | 1.28 | 14.08 |  | 4160.0000 | 49.92 | 3456 | 3136 | 800 | 1.28 | 4.06515 | LGB-900, CAL $=020$ |
| Density (\#/m ${ }^{3}$ ) | TUYA | 8/24/88 | 2 | 13:00 |  | 9.6 | 7.4667 |  | 1493.3300 | 114.13 | 5226.67 | 4017.78 | 1351.11 |  | 12219.9661 | $\mathrm{CAL}=020$ |
| Mean size (mm) | TUYA | 8/24/88 | 1 | 12:30 |  | 0.751 | 1.1921 |  | 0.8063 | 2.792 | 0.8204 | 0.2101 | 0.1771 | 9.1531 | 0.6162 |  |
| Mean size (mm) | TUYA | 8/24/88 | 2 | 13:00 |  | 0.7155 | 1.3705 |  | 0.8460 | 2.8291 | 0.8659 | 0.223 | 0.1748 |  | 0.5942 |  |
| Biomass (mg/m3) | TUYA | 9/23/88 | 1 | 12:45 |  | 0.011 | 1.1294 |  | 78.6503 | 16.3885 | 13.008 | 3.6026 | 0.4854 |  | 113.57 | NOTE:DAPH WRITTEN IN ABOVE |
| Biomass (mg/m3) | TUYA | 9/23/88 | 2 | 12:00 |  | 0.177 |  |  | 92.8498 | 75.2602 | 17.2042 | 2.9231 | 1.7493 | 164.54 | 190.16 | NOTE:BOS CROSSED OUT AND D |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 9/23/88 | 1 | 12:45 |  | 0.32 | 12.48 |  | 1920.0000 | 17.6 | 272 | 3648 | 592 |  | 6462.4 | CAL $=020$ |
| Density (\#/m) | TUYA | 9/23/88 | 2 | 12:00 |  | 2.1333 |  |  | 2506.6700 | 83.2 | 480 | 3013 | 2133.33 | 27.7333 | 8246.4 | CAL $=020$ |
| Mean size (mm) | TUYA | 9/23/88 | 1 | 12:45 | , | 0.5562 | 1.1517 |  | 0.8885 | 2.812 | 0.9446 | + 0.2281 | 0.1843 |  | 0.4593 |  |
| Mean size (mm) | TUYA | 9/23/88 | 2 | 12:00 |  | 0.8318 |  |  | 0.8566 | 2.7814 | 0.8517 | 0.2273 | 0.1825 | 1.2355 | 0.4726 |  |
| Biomass (mg/m3) | TUYA | 6/10/89 | 2 | 9:30 |  |  | 1.5995 |  | 239.9800 |  | 5.5256 | 4.7865 | 2.649 | 259.57 | 254.54 |  |
| Density ( $/ \mathrm{m}^{3}$ ) | TUYA | 6/10/89 | 2 | 9:30 |  |  | 14.9333 |  | 12373.1421 |  | 110.93 | 3596.19 | 3230.48 | 2.1333 | 19327.6922 | LGB=FISH LARVA |

Appendix 4 cont'd.

| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean size (mm) | TUYA | 6/10/89 | 2 | 9:30 |  |  | 1.1715 |  | 0.6372 |  | 0.96576 | 0.2422 | 0.1362 | 3.6025 | 0.4826 |  |
| Biomass (mg/m3) | TUYA | 6/20/89 | 1 | 12:00 |  |  | 1.2796 |  | 190.8600 |  | 9.9719 | 8.6807 | 1.4694 | 0.009 | 212.27 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 6/20/89 | 1 | 12:00 |  |  | 7.2 |  | 10303.8612 |  | 1600 | 4288 | 1792 | 0.32 | 17991.1946 | LGB=UNK INSECT HEAD WIDTH |
| Mean size (mm) | TUYA | 6/20/89 | 1 | 12:00 |  |  | 1.4708 |  | 0.6436 |  | 0.4345 | 0.2649 | 0.0918 | 0.2194 | 0.4801 |  |
| Biomass (mg/m3) | TUYA | 7/15/89 | 1 | 10:00 |  |  | 9.6238 |  | 158.7400 | 42.7638 | 215.13 | 1.0643 | 6.0177 |  | 433.34 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 7/15/89 | 1 | 10:00 |  |  | 38.4 |  | 7850.6700 | 81.92 | 4352 | 4352 | 7338.67 |  | 24013.1979 | CAL $=020$ |
| Mean size (mm) | TUYA | 7/15/89 | 1 | 10:00 |  |  | 1.6314 |  | 0.6844 | 2.2105 | 0.8986 | 0.1343 | 0.1047 |  | 0.4531 |  |
| Biomass (mg/m3) | TUYA | 8/28/89 | 1 | 14:00 |  |  | 45.2537 |  | 274.6700 | 25.4263 | 13.2681 | 2.0596 | 0.6822 |  | 361.36 |  |
| Biomass (mg/m3) | TUYA | 8/28/89 | 2 | 13:00 |  |  | 15.4819 |  | 229.0000 | 41.4282 | 11.69 | 2.079 | 2.3179 |  | 301.99 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 8/28/89 | 1 | 14:00 |  |  | 204.8 |  | 9792.0000 | 33.28 | 448 | 8256 | 832 |  | 19565.9077 | CAL=020 |
| Density (\#/m ${ }^{3}$ ) | TUYA | 8/28/89 | 2 | 13:00 |  |  | 142.93 |  | 7040.0000 | 64 | 320 | 6080 | 2826.67 |  | 16473.277 | CAL=020 |
| Mean size (mm) | TUYA | 8/28/89 | 1 | 14:00 |  |  | 1.404 |  | 0.7713 | 2.6109 | 0.7893 | 0.1355 | 0.0907 |  | 0.4842 |  |
| Mean size (mm) | TUYA | 8/28/89 | 2 | 13:00 |  |  | 1.1057 |  | 0.8090 | 2.3555 | 0.8555 | 0.1535 | 0.1171 |  | 0.4578 |  |
| Biomass (mg/m3) | TUYA | 10/1/89 | 1 | 12:00 |  |  | 35.1209 |  | 120.0600 | 16.5877 | 21.7467 | 0.7122 | 2.4891 |  | 196.71 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 10/1/89 | 2 | 12:30 |  |  | 148.86 |  | 122.5000 | 54.1979 | 14.7679 | 1.4081 | 1.4245 |  | 343.2 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 10/1/89 | 1 | 12:00 |  |  | 219.43 |  | 4352.0000 | 21.3333 | 621.71 | 2560 | 3035.43 |  | 10809.8583 | CAL $=020$ |
| Density (\#/m ${ }^{3}$ ) | TUYA | 10/1/89 | 2 | 12:30 |  |  | 518.1 |  | 3596.1900 | 58.1333 | 335.24 | 3382.86 | 1737.14 |  | 9627.66 | CAL $=020$ |
| Mean size (mm) | TUYA | 10/1/89 | 1 | 12:00 |  |  | 1.3754 |  | 0.7656 | 2.6228 | 0.8338 | 0.1425 | 0.0953 |  | 0.4498 |  |
| Mean size (mm) | TUYA | 10/1/89 | 2 | 12:30 |  |  | 1.6366 |  | 0.8306 | 2.8088 | 0.9153 | 0.1654 | 0.1113 |  | 0.5254 |  |
| Biomass (mg/m3) | TUYA | 6/6/90 | 1 |  |  | 0.3035 |  |  | 146.3000 |  | 11.3904 | 7.6347 | 1.6094 | 569.88 | 167.24 |  |
| Biomass (mg/m3) | TUYA | 6/6/90 | 2 |  |  |  |  |  | 146.7900 |  | 9.0824 | 6.3648 | 2.5803 |  | 164.82 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 6/6/90 | 1 |  |  | 10.24 |  |  | 6144.0000 |  | 240.64 | 6144 | 1962.67 | 5.12 | 14506.4117 | LGB=FISHLARV. |
| Density (\#/m ${ }^{3}$ ) | TUYA | 6/6/90 | 2 |  |  |  |  |  | 6346.6700 |  | 185.6 | 6613.33 | 3146.67 |  | 16292.21 |  |
| Mean size (mm) | TUYA | 6/6/90 | 1 |  |  | 0.5098 |  |  | 0.6388 |  | 0.9397 | 0.233 | 0.1011 | 3.4865 | 0.4001 |  |
| Mean size (mm) | TUYA | 6/6/90 | 2 |  |  |  |  |  | 0.6289 |  | 0.953 | 0.2151 | 0.1031 |  | 0.3631 |  |
| Biomass (mg/m3) | TUYA | 7/16/90 | 1 | 10:30 |  |  |  |  | 69.5137 | 97.0425 | 116.3 | 0.5772 | 4.8631 |  | 288.3 |  |
| Biomass (mg/m3) | TUYA | 7/16/90 | 2 | 9:50 |  |  |  |  | 39.1936 | 98.8975 | 88.8553 | 0.3949 | 3.4237 |  | 230.76 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 7/16/90 | 1 | 10:30 |  |  |  |  | 2304.0000 | 121.6 | 2645.33 | 4053.33 | 5930.67 |  | 15054.627 | CAL=020 |
| Density ( / m $^{3}$ ) | TUYA | 7/16/90 | 2 | 9:50 |  |  |  |  | 1371.4300 | 157.87 | 1950.48 | 1645.71 | 4175.24 |  | 9300.72 | CAL=020 |
| Mean size (mm) | TUYA | 7/16/90 | 1 | 10:30 |  |  |  |  | 0.7831 | 2.6235 | 0.8656 | 0.1102 | 0.0923 |  | 0.3592 |  |
| Mean size (mm) | TUYA | 7/16/90 | 2 | 9:50 |  |  |  |  | 0.7446 | 2.3959 | 0.8616 | 0.1334 | 0.0971 |  | 0.3983 |  |
| Biomass (mg/m3) | TUYA | 10/7/90 | , | 15:00 |  | 0.0169 | 0.0524 |  | 131.2300 | 44.2416 | 21.3962 | 1.5293 | 0.4023 | 0.0069 | 198.88 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 10/7/90 | 2 | 16:00 |  |  | 0.1475 | 0.048 | 56.4030 | 23.5169 | 21.1138 | 1.9691 | 1.6619 |  | 104.86 |  |
| Density ( $\#^{\prime} / \mathrm{m}^{3}$ ) | TUYA | 10/7/90 | 1 | 15:00 |  | 0.4267 | 1.28 |  | 2752.0000 | 43.093 | 384 | 3328 | 490.67 | 0.4267 | 6999.89 | CAL $=020$, LGB $=$ CNK INSECT (HEA |
| Density (\#/m ${ }^{3}$ ) | TUYA | 10/7/90 | 2 | 16:00 |  |  | 2.1333 | 0.8 | 1344.0000 | 25.8667 | 384 | 4373.33 | 2026.67 |  | 8156.8 | CAL $=020$ |
| Mean size (mm) | TUYA | 10/7/90 | 1 | 15:00 |  | 0.5921 | 0.9207 |  | 0.9221 | 2.9063 | 1.0003 | 0.1688 | 0.0905 | 0.1836 | 0.5221 |  |
| Mean size (mm) | TUYA | 10/7/90 | 2 | 16:00 |  |  | 1.0739 | 1.0216 | 0.8617 | 2.7754 | 0.9933 | 0.1657 | 0.0976 |  | 0.311 |  |
| Biomass (mg/m3) | TUYA | 6/18/91 | 1 | 8:00 |  |  | 0.4345 |  | 274.2400 |  | 4.7187 | 9.7383 | 0.3499 |  | 289.48 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 6/18/91 | 1 | 8:00 |  |  | 2.56 |  | 4480.0000 |  | 197.12 | 6826.67 | 426.67 |  | 11933.0099 |  |
| Mean size (mm) | TUYA | 6/18/91 | 1 | 8:00 |  |  | 1.481 |  | 0.9468 |  | 0.6541 | 0.2344 | 0.0786 |  | 0.5035 |  |
| Biomass (mg/m3) | TUYA | 7/23/91 | 1 |  |  |  | 5.6607 | 0.0538 | 79.7070 | 61.753 | 241.79 | 3.3904 | 8.0469 |  | 400.4 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 7/23/91 | 1 |  |  |  | 22.4 | 1.0667 | 4266.6700 | 65.0667 | 8533,33 | 24746.2844 | 9813.33 |  | 47448.2282 | CAL $=020$ |
| Mean size (mm) | TUYA | 7/23/91 | 1 |  | : |  | 1.6066 | 0.9897 | 0.6613 | 2.8138 | 0.7323 | 0.1095 | 0.0919 |  | 0.2719 |  |
| Biomass (mg/m3) | TUYA | 9/4/91 | 1 |  |  | 0.743 | 5.6187 | 0.2728 | 103.8900 | 25.4612 | 8.0603 | 2.4873 | 8.3268 |  | 154.86 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 9/4/91 | 1 |  |  | 35.84 | 74.24 | 5.12 | 2304.0000 | 23.04 | 250.88 | 8021.33 | '10154.534 |  | 20868.9212 | CAL $=020$ |
| Mean size (mm) | TUYA | 9/4/91 | 1 |  |  | 0.4326 | 1.0256 | 0.9864 | 0.9171 | 2.9907 | 0.7963 | 0.1486 | 0.0898 |  | 0.2195 |  |
| Biomass (mg/m3) | TUYA | 10/10/91 | 1 | 10:30 |  | 0.075 | 3.4558 | 0.0035 | 52.0147 | 13.4707 | 5.4525 | 1.6399 | 1.0496 |  | 77.1617 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 10/10/91 | 1 | 10:30 |  | 7.3143 | 10.9714 | 1.8286 | 1408.0000 | 14.6286 | 138.97 | 6869.33 | 1280 |  | 9731.05 | CAL=020. $\mathrm{HOL}=141$ |
| Mean size (mm) | TUYA | 10/10/91 | 1 | 10:30 |  | 0.3131 | 1.7471 | 0.3338 | 0.8465 | 2.7965 | 0.8749 | 0.135 | 0.0754 |  | 0.2467 |  |
| Biomass (mg/m3) | TUYA | 6/25/92 | 1 |  |  | 0.1538 |  |  | 281.1200 |  | 1.6441 | 2.7225 | 0.2099 |  | 285.85 |  |
| Biomass ( $\mathrm{mg} / \mathrm{m} 3$ ) | TUYA | 6/25/92 | 2 |  |  | 7.3078 | 2.9341 |  | 455.9100 |  | 1.7735 | 9.6339 | 1.0496 |  | 478.61 |  |
| Density (\#/m ${ }^{3}$ | TUYA | 6/25/92 | 1 |  |  | 5.12 |  |  | 10495.9076 |  | 102.4 | 1792 | 256 |  | 12651.4412 |  |


| Property | Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | Tor | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Density ( $/ 1 / \mathrm{m}^{3}$ ) | TUYA | 6/25/92 | 2 |  |  | 426.67 | 51.2 |  | 19199.9457 |  | 89.6 | 4693.33 | 1280 |  | 25740.6791 |  |
| Mean size (mm) | TUYA | 6/25/92 | 1. |  |  | 0.5093 |  |  | 0.7496 |  | 0.594 | 0.266 | 0.1822 |  | 0.6683 |  |
| Mean size (mm) | TUYA | 6/25/92 | 2 |  |  | 0.4035 | 0.9743 |  | 0.7154 |  | 0.6159 | 0.2784 | 0.211 |  | 0.6056 |  |
| Biomass (mg/m3) | TUYA | 7/26/92 | 1 |  |  |  | 0.314 | 0.1449 | 256.1000 | 6.8455 | 128.89 | 2.5836 | 0.6298 |  | 395.5 |  |
| Biomass (mg/m3) | TUYA | 7/26/92 | 2 |  |  |  |  |  | 240.7200 | 0.6522 | 93.6973 | 4.214 | 1.7712 | 330.26 | 341.05 |  |
| Density (\#/m) | TUYA | 7/26/92 | 1 |  |  |  | 2.24 | 7.36 | 14335.7362 | 19.84 | 3712 | 2560 | 768 |  | 21405.235 | CAL $=020$ |
| Density (\#/m²) | TUYA | 7/26/92 | 2 |  |  |  |  |  | 13919.7040 | 3.2 | 2560 | 10079.9876 | 2160 | 3.2 | 28726.3204 | LGB $=900 . \mathrm{CAL}=020$ |
| Mean size (mm) | TUYA | 7/26/92 | 1 |  |  |  | 1.367 | 0.6775 | 0.6475 | 1.855 | 0.8152 | 0.2 | 0.1116 |  | 0.6051 |  |
| Mean size (mm) | TUYA | 7/26/92 | 2 |  |  |  |  |  | 0.6357 | 1.6158 | 0.8286 | 0.1571 | 0.1045 | 3.4112 | 0.4454 |  |
| Biomass (mg/m3) | TUYA | 8/24/92 | 1 |  |  |  | 0.5418 | 0.843 | 573.2000 | 14.3563 | 0.9481 | 7.1122 | 0.2099 |  | 597:21 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 8/24/92 | 1 |  |  |  | 7.68 | 21.12 | 13823.8814 | 11.52 | 36.88 | 11263.9315 | 256 |  | 25410.8972 | CAL=020 |
| Mean size (mm) | TUYA | 8/24/92 | 1 |  |  |  | 1.0046 | 0.8298 | 0.8923 | 3.1202 | 0.8144 | 0.1808 | 0.1281 |  | 0.5701 |  |
| Biomass (mg/m3) | TUYA | 8/25/92 | 2 |  |  |  | 0.2312 | 1.3777 | 534.5200 | 6.6927 | 1.9574 | 11.5745 | 0.2332 |  | 556.58 |  |
| Density (\#/ $\mathrm{m}^{3}$ ) | TUYA | 8/25/92 | 2 |  |  |  | 5.3333 | 35.2 | 15075.4401 | 5.3333 | 52.2667 | 23039.8687 | 284.44 |  | 38497.2759 | CAL $=020$ |
| Mean size (mm) | TUYA | 8/25/92 | 2 |  |  |  | 0.9294 | 0.8562 | 0.8414 | 3.1388 | 0.8543 | 0.1576 | 0.1163 |  | 0.4271 |  |
| Biomass (mg/m3) | TUYA | 9/18/92 | 1 |  |  |  | 1.3694 | 1.4442 | 853.8200 | 5.8206 | 2.4303 | 3.8191 | 0.4198 |  | 869.13 |  |
| Biomass (mg/m3) | TUYA | 9/18/92 | 2 |  |  |  | 0.1142 | 1.7141 | 303.2500 | 1.7555 | 2.0271 | 19.0269 |  |  | 327.88 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 9/18/92 | 1 |  |  |  | 6.4 | 24.32 | 23039.8687 | 5.12 | 51.2 | 12799.7062 | 512 |  | 36438.4588 | CAL $=020$ |
| Density (\#/m ${ }^{3}$ ) | TUYA | 9/18/92 | 2 |  |  |  | 2.1333 | 37.3333 | 8533.3300 | 3.2 | 42.6667 | 17066.3245 |  |  | 25685.0254 | CAL $=020$ |
| Mean size ( mm ) | TUYA | 9/18/92 | 1 |  |  |  | 1.3907 | 1 | 0.8506 | 3.0307 | 0.93 | 0.1409 | 0.073 |  | 0.591 |  |
| Mean size (mm) | TUYA | 9/18/92 | 2 |  |  |  | 0.9952 | 0.9068 | 0.8454 | 2.2636 | 0.9384 | 0.2295 |  |  | 0.4366 |  |
| Biomass (mg/m3) | TUYA | 6/16/93 | 1 | 23:00 |  |  |  | 0.0484 | 466.4900 |  | 0.4119 | 6.286 | 2.7989 | 472.99 | 476.03 |  |
| Biomass (my/m3) | TUYA | 6/16/93 | 2 |  |  |  | 0.0264 | 0.146 | 121.2700 |  | 0.866 | 4.7838 | 4.9331 | 167.18 | 132.03 |  |
| Density (\#/mi ${ }^{3}$ ) | TUYA | 6/16/93 | 1 | 23:00 |  |  |  | 30.9333 | 12799.7062 |  | 68.2667 | 613.33 | 3413.33 | 9.6 | 22935.0663 | LGB $=900$ |
| Density (\#/m ${ }^{3}$ ) | TUYA | 6/16/93 | 2 |  |  |  | 2.56 | 84.48 | 4992.0000 |  | 184.32 | 4352 | 6016 | 2.56 | 15633.6361 | LGB $=900$ |
| Mean size (mm) | TUYA | 6/16/93 | 1 | 23:00 |  |  |  | 0.3088 | 0.7911 |  | 0.4242 | 0.204 | 0.0977 | 2.6363 | 0.5177 |  |
| Mcan size (mm) | TUYA | 6/16/93 | 2 |  |  |  | 0.5843 | 0.3213 | 0.6916 |  | 0.4036 | 0.2208 | 0.1101 | 2.9286 | 0.3317 |  |
| Biomass (mg/m3) | TUYA | 8/4/93 | 1 |  |  | 0.0489 | 0.3768 | 1.5845 | 164.3000 | 25.0296 | 97.9433 | 1.8187 | 7.4172 |  | 298.52 |  |
| Biomass (mg/m3) | TUYA | 8/4/93 | 2 |  |  |  | 0.1646 | 42.6239 | 61.8923 | 15.4306 | 59.8842 | 0.615 | 3.382 |  | 183.99 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 8/4/93 | 1 |  |  | 1.0237 | 2.0473 | 51.1836 | 5973.3300 | 20.4734 | 4096 | 8874.67 | 9045.33 |  | 28064.0277 | CAL $=020$ |
| Density (\#/ ${ }^{3}$ ) | TUYA | 8/4/93 | 2 |  |  |  | 0.5333 | 2346.67 | 2204.4400 | 12.2667 | 2204.44 | 2631.11 | 4124.44 |  | 13523.8393 | CAL=020 |
| Mean size (mm) | TUYA | 8/4/93 | 1 |  |  | 0.646 | 1.4325 | 0.788 | 0.7588 | 3.1067 | 0.6429 | 0.1265 | 0.0924 |  | 0.329 |  |
| Mem size (mm) | TUYA | 8/4/93 | 2 |  |  |  | 1.8061 | 0.6564 | 0.7493 | 3.1397 | 0.642 | 0.1302 | 0.0965 |  | 0.3984 |  |
| Biomass (mg/m3) | TUYA | 9/1/93 | 1 |  | 0.0241 |  |  | 4.0137 | 298.9100 | 5.6764 | 2.2239 | 1,4853 | 2.729 |  | 315.06 |  |
| Biomass (mg/m3) | TUYA | 9/1/93 | 2 |  |  |  |  | 5.4953 | 65.7325 | 3.0071 | 2.8154 | 0.9077 | 1.7318 |  | 79.6899 |  |
| Density ( $\#^{\left(1 m^{3}\right.}$ ) | TUYA | 9/1/93 | 1 |  | 2.56 |  |  | 138.24 | 8704.0000 | 5.12 | 107.52 | 6784 | 3328 |  | 19069.0914 | CAL $=020$, DIA $=141$ |
| Density (\#/ $/ \mathrm{m}^{3}$ ) | TUYA | 9/1/93 | 2 |  |  |  |  | 117.76 | 1984.0000 | 2.56 | 94.72 | 4864 | 2112 |  | 9175.04 | CAL $=020$ |
| Mean size (mm) | TUYA | 9/1/93 | 1 |  | 0.5672 |  |  | 0.7874 | 0.8321 | 3.0014 | 0.6532 | 0.1312 | 0.0933 |  | 0.4531 |  |
| Mean size (mm) | TUYA | 9/1/93 | 2 |  |  |  |  | 0.923 | 0.8195 | 3.0691 | 0.7395 | 0.1232 | 0.0862 |  | 0.2827 |  |
| Biomass (mg/m3) | TUYA | 6/18/94 | 2 | 19:00 | 0.7091 |  | 0.0479 | 0.0298 | 824.3800 |  | 0.6934 | 3.1993 | 9.0965 |  | 838.15 |  |
| Density (\#/mi ${ }^{3}$ ) | TUYA | 6/18/94 | 2 | 19:00 | 16 |  | 0.5333 | 5.8667 | 52905.3997 |  | 10.6667 | 2133.33 | 11093.2806 |  | 66165.2564 | DIA juv epi or heterocope |
| Mean size (mm) | TUYA | 6/18/94 | 2 | 19:00 | 0.9415 |  | 1.199 | 0.4558 | 0.6065 |  | 1.0485 | 0.2607 | 0.1674 |  | 0.5219 |  |
| Biomass (mg/m3) | TUYA | 7/28/94 | 1 | 15:30 |  | 14.3473 | 21.2283 |  | 881.7200 |  | 22.5413 | 2.2504 | 0.4198 |  | 942.51 |  |
| Biomass (mg/m3) | TUYA | 7/28/94 | 2 | 18:00 |  |  |  |  | 2492.0700 |  | 52.9712 | 6.1127 | 1.3995 |  | 2552.55 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 7/28/94 | 1 | 15:30 |  | 328.21 | 287.18 |  | 24063.5668 |  | 164.1 | 5632 | 512 |  | 30987.0334 |  |
| Density (\#/m ${ }^{3}$ ) | TUYA | 7/28/94 | 2 | 18:00 |  |  |  |  | 58878.9423 |  | 221.87 | 15359.957 | 1706.67 |  | 76167.5524 |  |
| Mean size (mm) | TUYA | 7/28/94 | 1 | 15:30 |  | 0.6109 | 10843 |  | 0.8392 |  | 1.3749 | 0.1628 | 0.1222 |  | 0.7071 |  |
| Mean size (mm) | TUYA | 7/28/94. | 2 | 18:00 |  |  |  |  | 0.8903 |  | 1.6657 | 0.1625 | 0.1299 |  | 0.7288 |  |
| Biomass (mg/m3) | TUYA | 9/4/94 | 1 | 13:30 |  | 0.1787 |  | 29.6238 | 1091.5300 |  | 20.5623 | 10.6413 | 0.5598 |  | 1153.1 |  |
| Biomass (mg/m3) | TUYA | 9/4/94 | 2 |  |  |  |  | 77.4775 | 2622.3400 |  | 22.2022 | 16.1234 | 6.2976 |  | 2744.45 |  |
| Density ( $\# / \mathrm{m}^{3}$ ) | TUYA | 9/4/94 | 1 | 13:30 |  | 5.12 |  | 373.76 | 28330.2235 |  | 153.6 | 15359.957 | 682.67 |  | 44905.5479 |  |
| Density (\#/m) | TUYA | 9/4/94 | 2 |  |  |  |  | 819.2 | 62293.1080 |  | 187.73 | 17066.3245 | 7680 |  | 88046.0749 |  |

## Appendix 4 cont'd.

| Lake | Date | Site | Time | DIA | BOS | DPH | HOL | CYC | CAL | DPT | NP | ROT | LGB | TOT | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TUYA | 9/4/94 | 1 | 13:30 |  | 0.5592 |  | 1.0224 | 0.8684 |  | 1.3309 | 0.1993 | 0.1387 |  | 0.6313 |  |
| TUYA | 9/4/94 | 2 |  |  |  |  | 1.0881 | 0.8950 |  | 1.2609 | 0.221 | 0.1841 |  | 0.7049 |  |
| TUYA | 8/1/95 | 1 | 14:00 |  | 2.7074 | 0.7502 | 181.43 | 4.7320 |  | 31.636 | 5.785 | 4.7232 |  | 231.77 |  |
| TUYA | 8/1/95 | 2 | 18:45 |  | 18.9096 | 0.5474 | 278.65 | 4.5673 |  | 52.5841 | 8.0748 | 6.7058 |  | 369.95 |  |
| TUYA | 8/1/95 | 1 | 14:00 |  | 213.33 | 27.7333 | 4010.67 | 83.2000 |  | 1194.67 | 4053.33 | 5760 |  | 15342.6366 |  |
| TUYA | 8/1/95 | 2 | 18:45 |  | 711.11 | 32 | 6186.67 | 80.0000 |  | 1777.78 | 4622.22 | 8177.78 |  | 21587.3832 |  |
| TUYA | 8/1/95 | 1 | 14:00 |  | 0.3491 | 0.7401 | 0.8701 | 0.9868 |  | 0.6882 | 0.2539 | 0.1236 |  | 0.406 |  |
| TUYA | 8/1/95 | 2 | 18:45 |  | 0.4727 | 0.6543 | 0.8385 | 0.9696 |  | 0.716 | 0.2694 | 0.1193 |  | 0.4223 |  |
| TUYA | 9/11/95 | 1 |  |  | 22.4274 | 218.48 |  | 1.6228 |  | 573.05 | 7.6924 | 0.4198 |  | 823.69 |  |
| TUYA | 9/11/95 | 2 |  | 0.0936 | 40.1685 | 94.1665 |  | 114.6400 |  | 407.14 | 5.0428 | 0.8528 |  | 662.1 |  |
| TUYA | 9/11/95 | 1 |  |  | 1024 | 12799.7062 |  | 43.7333 |  | 22783.4769 | 5888 | 512 |  | 43051.6697 |  |
| TUYA | 9/11/95 | 2 |  | 0.8 | 1680 | 5680 |  | 2720.0000 |  | 9120 | 3760 | 1040 |  | 24000.484 |  |
| TUYA | 9/11/95 | 1 |  |  | 0.4508 | 0.6511 |  | 0.8301 |  | 0.709 | 0.2466 | 0.1064 |  | 0.6153 |  |
| TUYA | 9/11/95 | 2 |  | 1.3091 | 0.462 | 0.6582 |  | 0.8840 |  | 0.8229 | 0.2471 | 0.1131 |  | 0.6446 |  |


[^0]:    ${ }^{1}$ Within attachment 2 of the letters to govenment prepared by the Canadian and U.S. sections of the Pacific Salmon Commission it specifies that for a given brood year Tahltan Lake origin fry will be outplanted to Tuya Lake only when the escapement into Tahltan Lake for that year exceeds 15,000 adult sockeye (PSC 1989, Appendix 2).

[^1]:    ${ }^{2}$ Eggs for outplants to both Tahltan and Tuya lakes are taken at Tahltan Lake and share a common history to the fry stage.

[^2]:    ${ }^{3}$ It is currently estimated to be as high as $70 \%$ to $80 \%$ depending on annual run strength. Foot surveys of the interconnecting stream indicate amual spawning populations of 500-1500.

[^3]:    ${ }^{4}$ 1) an outplant of 934,000 fry in 1991 and ; 2) and estimated female escapement to Little Trapper Lake of 3,889 in 1990
    ${ }^{5}$ The wild sockeye present in Trapper Lake are almost certainly kokanee and apparently do not emigrate out of the lake.

[^4]:    ${ }^{6}$ The analyses presented here are for plankton of all sizes, including rotifers and nauplii larvae.

[^5]:    ${ }^{7}$ Based on scale pattern analysis, the proportion of 0-check fish is higher for samples collected from the creek spawning population; this implies that these fish leave the system early and do nod over-winter (as fry) in freshwater.

[^6]:    ${ }^{2}$ Egg-take targets in millions (M) are shown in parentheses.

[^7]:    ${ }^{\text {a }}$ Egg-take targets in millions (M) are shown in parentheses.

[^8]:    a Measurements are from specimens preserved in denatured (94\%) ethanol and are not directly comparable to fresh measurements. Total beachseine catches are for index sites only and do not include supplemental catches (see Methods, section 3.4). Total sampled refers to no. of juveniles sampled for thermal marks. In some cases, subsamples were selected from the total sampled, proportional to numbers captured in individual sels or traws. Percentages may not agree exactly with sample sizes for lenglh/weight ( n ) since some fish observed for marks may not have been measurable.

[^9]:    ${ }^{2}$ Measurements are from specimens preserved in $90 \%$ denatured) ethanol and are not directly comparable to fresh measurements. Total beachseine catches are for index sites only and do not include supplemental catches (see Methods, section 3.4). Total sampled refers to no. of juveniles sampled for thernal marks. In some cases, sub-samples were selected from the total sampled, proportional to numbers captured in individual sets or frawis. Percentages may not agree exactly with sample sizes for length/weight ( n ) since some fish observed for marks may not have been measurable.

[^10]:    ${ }^{\text {a }}$ Measurements are from specimens preserved in denatured ( $94 \%$ ) ethanol and are not directly comparable to fresh measurements. Total beachseine catches are for index sites only and do not include supplemental catches (see Methods, section 3.4). In some cases, sub-samples were selected from the total sampled, proportional to numbers captured in individual sets or traws. Percentages may not agree exactly with sample sizes for lenglh/weight (n) since some fish observed for marks may not have been meaurable.

[^11]:    ${ }^{\text {a }}$ Measurements are from specimens preserved in denatured ( $94 \%$ ) ellhanol and are not directly comparable to fresh measurements. Total beachseine catches are for index sites only and do not include supplemental catches (see Methods, section 3.4). Tolal sampled refers to no. of juveniles sampled for thenmal marks. In some cases, subsamples were selected from the total sampled, proportional to numbers captured in individual sets or trawis. Percentages may not agree exactly with sample sizes for length/weight (n) since some fish observed for marks may not have been measurable.

[^12]:    a Measurements are from specimens preserved in denatured ( 94 or $90 \%, 1993$ and 1994, respectively) ethanol and are not directly comparable to fresh measurements. Total beachseine catches are for index sites only and do not include supplemental catches (see Methods, section 3.4). Total sampled refers to no. of juveniles sampled for thermal marks. In some cases, subsamples were selected from the total sampled, proportional to numbers captured in individual sets or trawls. Percentages may not agree exactly with sample sizes for length/weight (n) since some fish observed for marks may not have been measurable.

[^13]:    not a true survival as it does not account for smolts which migrated at age $1+$; also, this age class is likely underrepresented because of the increased net avoidance associated with larger size.
    ${ }^{\text {b }}$ estimate extremely poor because of exceptionally small trawl sample (3 fish).
    ${ }^{c} 16.6 \%$ omitting BY 1993 (probably more realistic).

[^14]:    ${ }^{2}$ measurements are from fresh (unpreserved) fish.

[^15]:    ${ }^{\text {a }}$ measurements are from fresh (unpreserved) fish.

[^16]:    ${ }^{\text {a }}$ measurements are from fresh (unpreserved) fish.
    bonly 1 fish captured

[^17]:    ${ }^{2}$ measurements are from fresh (unpreserved) fish.

[^18]:    ${ }^{2}$ reliability of estimate low due to small sample size $(\mathrm{n}=32)$

[^19]:    ${ }^{2} \mathrm{C}=$ clear, $\mathrm{C} / \mathrm{S}=$ clear/stained, $\mathrm{G}=$ glacial, $\mathrm{C} / \mathrm{G}=$ clear/glacial
    ${ }^{\mathrm{b}}$ derived from Secchi depth; euphotic depth estimates were calculated from revised average Secchi depths compiled from additional sample years (see section 3.62 ).
    ${ }^{c} \mathrm{EV}=$ euphotic depth X surface area. 1 EV unit $=1$ million cubic meters.
    ${ }^{\mathrm{d}}$ as reported in PSC 1988.
    ${ }^{0}$ estimated from euphotic volume by formula (PSC 1988): number adults ( 1000 's) $=29.7+(1.81 \mathrm{X} \mathrm{EV}$ ) Note: the Enhancement Sub Committee recommends assuming a more conservative estimate of one-half this number until proven.
    ${ }^{f}$ wild plus enhanced, using a biostandard fry to adult survival of $2 \%$. Note: in accordance with the more conservative adult production estimates, the Enhancement Sub-Committee recommends a safe maximum of one-half this number, until corroboration by further research.

